

# Outer Dowsing Offshore Wind

## Habitats Regulations Assessment

### Predator Control Evidence Base and Roadmap

Date: November 2024

Document Reference: 7.7.5

Revision: 2.0

Company:		<b>Outer Dowsing Offshore Wind</b>		Asset:		<b>Whole Asset</b>	
Project:		<b>Whole Wind Farm</b>		Sub Project/Package:		Whole Asset	
Document Title or Description:		Predator Control Evidence Base and Roadmap					
Internal Document Number:		PP1-ODOW-DEV-CS-PLA-0040_02		3 <sup>rd</sup> Party Doc No (If applicable):		N/A	
Rev No.	Date	Status / Reason for Issue	Author	Checked by	Reviewed by	Approved by	
V1.0	March 2024	DCO Application	GoBe	GoBe	Shepperd & Wedderburn	Outer Dowsing	
V2.0	November 2024	Deadline 2 Examination update	GoBe	GoBe	Shepperd & Wedderburn	Outer Dowsing	

## Table of Contents

Terminology .....	6
1 Executive Summary .....	8
2 Introduction .....	9
3 Site Selection .....	11
4 Plémont Seabird Reserve.....	12
4.1 Background .....	12
4.2 Overview of the Selected Area .....	12
4.3 Predators within the Plémont Seabird Reserve.....	13
4.3.1 Summary of Non-Native Predators Present .....	13
4.4 Control measures.....	14
4.4.1 Predator-proof fence .....	14
4.4.2 Predator control.....	14
4.5 Potential for Population Growth .....	15
4.5.1 Scale of Compensation .....	17
4.6 Connectivity with the National Site Network .....	18
5 Monitoring and Adaptive Management.....	19
5.1 Guillemot and Razorbill Implementation Plans .....	19
5.2 Post-Implementation Monitoring.....	19
6 Implementation .....	21
6.1 Exclusivity Agreement.....	21
6.2 Land Availability .....	21
6.3 Reserve Establishment and Management .....	21
7 Funding .....	24
8 References .....	25
Annex 1 – Plémont Seabird Reserve: Feasibility Study Report for a Predator-Exclusion Fence .....	27
Annex 2 – Plémont Seabird Sanctuary: Design Statement.....	28
Annex 3 – Plémont Seabird Sanctuary: Management Plan .....	29

## Table of Tables

Table 6-1 Indicative timescales for establishment of Plemont Seabird Reserve .....	23
Table 7-1 Anticipated costs for establishment and management of Plémont Seabird Reserve .....	24

## Acronyms & Definitions

### Abbreviations / Acronyms

Abbreviation / Acronym	Description
<b>AEoI</b>	Adverse Effect on Integrity
<b>ANS</b>	Artificial Nesting Structure
<b>BOTE</b>	Birds of the Edge
<b>COWSC</b>	Collaboration on Offshore Wind Strategic Compensation
<b>CSG</b>	Compensation Steering Group
<b>dDCO</b>	draft Development Consent Order
<b>DCO</b>	Development Consent Order
<b>DESNZ</b>	Department for Energy Security and Net Zero, formerly Department of Business, Energy and Industrial Strategy (BEIS), which was previously Department of Energy & Climate Change (DECC)
<b>FFC</b>	Flamborough and Filey Coast
<b>GT R4 Ltd</b>	The Applicant. The special project vehicle created in partnership between Corio Generation (and its affiliates), Gulf Energy Development and TotalEnergies
<b>GCP</b>	Guillemot Compensation Plan
<b>GCIMP</b>	Guillemot Compensation Implementation and Monitoring Plan
<b>MRF</b>	Marine Recovery Fund
<b>NTJ</b>	National Trust Jersey
<b>ORBA</b>	Offshore Restricted Build Area
<b>RCIMP</b>	Razorbill Compensation Implementation and Monitoring Plan
<b>RCP</b>	Razorbill Compensation Plan
<b>RIAA</b>	Report to Inform Appropriate Assessment
<b>SNCB</b>	Statutory Nature Conservation Body
<b>SoS</b>	Secretary of State
<b>SPA</b>	Special Protection Area
<b>UK</b>	United Kingdom

## Terminology

Term	Definition
<b>The Applicant</b>	GT R4 Ltd. The Applicant making the application for a DCO. The Applicant is GT R4 Limited (a joint venture between Corio Generation, Total Energies and Gulf Energy Development (GULF)), trading as Outer Dowsing Offshore Wind. The Project is being developed by Corio Generation (a wholly owned Green Investment Group portfolio company), TotalEnergies and GULF.
<b>Baseline</b>	The status of the environment at the time of assessment without the development in place.
<b>Compensatory Measures</b>	Stage 3 of the Habitats Regulations Assessments (see Derogation) involves the development of compensation measures for any features which the report to inform appropriate assessment was unable to conclude no adverse effect on integrity on.
<b>deemed Marine Licence (dML)</b>	A marine licence set out in a Schedule to the Development Consent Order and deemed to have been granted under Part 4 (marine licensing) of the Marine and Coastal Access Act 2009.
<b>Derogation</b>	Stage 3 of the Habitats Regulations Assessments which is triggered once it is determined that you cannot avoid adversely affecting the integrity of a designated site. Involves assessing if alternative solutions are available to achieve the same goals as the project, if there are imperative reasons of overriding public interest, and required compensatory measures.
<b>Development Consent Order (DCO)</b>	An order made under the Planning Act 2008 granting development consent for a Nationally Significant Infrastructure Project (NSIP) from the Secretary of State (SoS) for Department for Energy Security and Net Zero (DESNZ).
<b>Effect</b>	Term used to express the consequence of an impact. The significance of an effect is determined by correlating the magnitude of the impact with the sensitivity of the receptor, in accordance with defined significance criteria.
<b>Evidence Plan</b>	A voluntary process of stakeholder consultation with appropriate Expert Topic Groups (ETGs) that discusses and, where possible, agrees the detailed approach to the Environmental Impact Assessment (EIA) and information to support Habitats Regulations Assessment (HRA) for those relevant topics included in the process, undertaken during the pre-application period.
<b>Habitats Regulations Assessment (HRA)</b>	A process which helps determine likely significant effects and (where appropriate) assesses adverse impacts on the integrity of European conservation sites and Ramsar sites. The process consists of up to four stages of assessment: screening, appropriate assessment, assessment of alternative solutions and assessment of imperative reasons of over-riding public interest (IROPI) and compensatory measures.
<b>Impact</b>	An impact to the receiving environment is defined as any change to its baseline condition, either adverse or beneficial.
<b>Intertidal</b>	The area between Mean High Water Springs (MHWS) and Mean Low Water Springs (MLWS)

<b>Mitigation</b>	Mitigation measures, or commitments, are commitments made by the Project to reduce and/or eliminate the potential for significant effects to arise as a result of the Project. Mitigation measures can be embedded (part of the project design) or secondarily added to reduce impacts in the case of potentially significant effects.
<b>Offshore Restricted Build Area</b>	The area within the array area, where no wind turbine generator, offshore transformer substation or offshore accommodation platform shall be erected.
<b>Outer Offshore (ODOW) Dowsing Wind</b>	The Project.
<b>Order Limits</b>	The area subject to the application for development consent, the limits shown on the works plans within which the Project may be carried out.
<b>Strategic Compensation</b>	Collaborative approach by developers and/or government departments to secure compensation for adverse effects on the conservation objectives of a Marine Protected Area.
<b>The Project</b>	Outer Dowsing Offshore Wind including proposed onshore and offshore infrastructure.
<b>The Inspectorate Planning</b>	The agency responsible for operating the planning process for Nationally Significant Infrastructure Projects (NSIPs).
<b>Study Area</b>	Area(s) within which environmental impact may occur – to be defined on a receptor-by-receptor basis by the relevant technical specialist.

## 1 Executive Summary

1. The Report to Inform Appropriate Assessment (RIAA; Document 7.1) has concluded that there would be no Adverse Effect on Integrity (AEI) to the common guillemot, *Uria aalge* (hereafter 'guillemot'), and razorbill, *Alca torda* (hereafter 'razorbill') features of the Flamborough and Filey Coast (FFC) Special Protection Area (SPA) due to displacement, both when considering the project alone and in combination with other plans or projects.
2. Following consultation with Natural England and other relevant consultees through the Evidence Plan Process, the Project has however provided a 'without prejudice' derogation case for both guillemot and razorbill in relation to the FFC SPA; alongside this, a number of options for relevant compensation measures were developed as far as possible at the point of application. In the event that the Secretary of State determines potential for Adverse AEI and considers that compensation is required, the Project has provided sufficient confidence that compensation measures are available, securable and deliverable.
3. Information is provided on the proposed Plémont Seabird Reserve, including the potential for predator control and the resultant seabird population growth / scale of compensation that this measure could provide. Connectivity of the Plémont site with the FFC SPA and wider National Site Network is also discussed
4. This document provides the evidence base and roadmap for the delivery of predator control, focusing on the proposed Plémont Seabird Reserve in Jersey. The document has been updated following the submission of the planning application for Plémont Seabird Sanctuary by Birds on the Edge Partnership in November 2024 (P/2024/1198). This document provides updates on fence design, location and updates to planned monitoring of the measure to ensure that the reserve's effectiveness is maintained throughout the duration of the project.



## 2 Introduction

5. The Report to Inform Appropriate Assessment (RIAA; Document 7.1) has concluded that there would be no Adverse Effect on Integrity (AEI) to the common guillemot, *Uria aalge* (hereafter 'guillemot'), and razorbill, *Alca torda* (hereafter 'razorbill') features of the Flamborough and Filey Coast (FFC) Special Protection Area (SPA) due to displacement, both when considering the project alone and in combination with other plans or projects.
6. Following consultation with Natural England and other relevant consultees through the Evidence Plan Process, the Project has however provided a 'without prejudice' derogation case for both guillemot and razorbill in relation to the FFC SPA; alongside this, a number of options for relevant compensation measures have been developed as far as possible at the point of application. In the event that the Secretary of State determines potential for AEI and considers that compensation is required, the Project has provided sufficient confidence that compensation measures are available, securable and deliverable.
7. This document provides the evidence base and roadmap for the delivery of predator control, focussing on the proposed Plémont Seabird Reserve in Jersey.
8. Section 3 sets out the process followed in the selection of suitable sites for this measure, specifically where there is clear evidence that predator control is feasible and would lead to an increase in the annual productivity and recruitment of guillemot and razorbill into the regional population of the southern North Sea.
9. Section 4 provides further information on the proposed Plémont Seabird Reserve, including the potential for predator control and the resultant seabird population growth / scale of compensation that this measure could provide. Connectivity of the Plémont site with the FFC SPA and wider National Site Network is also discussed in Section 4.6.
10. Implementation of the measure, including monitoring and adaptive management, as well as the funding required, are provided in Sections 5 to 7. Discussions regarding the development of this measure have been framed around an earlier version of the Defra compensation guidance (published in 2021). However, although still under consultation, updated draft guidance has been published recently (Defra, 2024). The new proposals prioritise Ecological Effectiveness when considering compensation, i.e. the ecological outcome and the confidence that the measures will be effective.
11. This document is supported by The Plémont Seabird Reserve Feasibility Study Report undertaken in 2021 by Birds on the Edge for the National Trust Jersey (NTJ), provided as Annex 1.
12. A planning application for Plémont Seabird Sanctuary was made by Birds on the Edge Partnership in November 2024 (P/2024/1198). This included further detail on the design, management and monitoring of the fence and a design statement which presents the technical specifications of the predator proof fence is provided in Annex 2. Detail on management and monitoring can be found in the management plan which is presented in Annex 3. The management plan details the following elements:

- The installation and decommissioning of the predator-proof fence;
  - The eradication and biosecurity programmes; and
  - The monitoring and incursion response programme.
13. Final monitoring and adaptive management measures for this measure would be set out in the Guillemot Compensation Implementation and Monitoring Plan and the Razorbill Compensation Implementation and Monitoring Plan, as required, following consultation with the respective steering groups. This is secured in paragraph 4(a)(vi), Parts 2 and 3, Schedule 22 of the draft Development Consent Order (dCO) (3.1).
14. It is envisaged that compensation for guillemot and razorbill, if required, could be delivered through a suite of measures appropriate for both species. The extent to which this could be necessary being dependent on the final quantum of compensation being deemed as required by the Secretary of State (SoS). The supporting measures are discussed in more detail in the Additional Measures for Compensation of Guillemot and Razorbill (document reference 7.7.6) and the Artificial Nesting Structure (ANS) Evidence and Roadmap (document reference 7.7.4).
15. This report should be read alongside the Project's Guillemot Compensation Plan (7.7.2) and the Razorbill Compensation Plan (7.7.3).
16. The report should also be read alongside the following documents submitted at Deadline 2:
- 19.8 Levels of precaution in the assessment and confidence calculations for offshore ornithology;
  - 19.9 Statutory Nature Conservation Bodies (SNCB) guidance and bioseasons for guillemot; and
  - 19.10 Rates of displacement in guillemot and razorbill.
17. These documents discuss the Applicant's approach to assessment which ultimately defines the levels of compensation potentially required for each species.
18. Predator control at Plémont Seabird forms the primary compensation measure for guillemot and razorbill when using the Applicant's approach of 50% displacement and 1% mortality, the adult proportion presented in Furness (2015), with 50% apportioning to FFC SPA for guillemot (ASI-099 and 19.10). This measure can be supported, if necessary, by the suite of 'Additional Measures' in south-west England (see Additional Measures for Compensation of Guillemot and Razorbill document reference 7.7.6). Additional supporting compensation could be provided by ANS should that be deemed necessary (see ANS Evidence and Roadmap, document reference 7.7.4).
19. Compensation requirements for guillemot and razorbill, calculated using the Applicant's approach and Natural England's anticipated approach, have been presented in each of the species-specific Compensation Plans: the Guillemot Compensation Plan (document reference 7.7.2) and the Razorbill Compensation Plan (document reference 7.7.3).

### 3 Site Selection

20. A desk-based site review, focussed on locations which could support populations of breeding guillemot and/or razorbill, and where predator control measures would be feasible, was undertaken. Islands were preferred to mainland sites due to the greater potential for full predator eradication or control/exclusion. The site selection process was informed by Stanbury *et al.* (2017), taking into account factors such as the connectivity to other sites with predators, human population size and island area size.
21. Several sites from the following areas were identified:
- Channel Islands;
  - Isles of Scilly; and
  - Scottish Islands.
22. A shortlist of suitable sites was created, with input from consultation with relevant site managers as well as predator control experts. Information considered included, but was not limited to:
- The feasibility of undertaking predator control on the relevant site, including whether a full eradication, or a control/exclusion is possible;
  - The extent of the predation issue, and the predator species present;
  - Any site-specific requirements (including over the lifetime of the measure);
  - The expected quantifiable benefits to guillemot and razorbill as a result of eradication on the site;
  - The ability of the site to host measures in addition to existing site management; and
  - The connectivity of birds at the proposed site with relevant SPAs (e.g. FFC SPA) and the UK national site network.
23. These considerations guided the site selection process towards a location where predator control could be carried out, to the benefit of seabird populations and at a location where any population increases would benefit the FFC SPA colonies and the UK national site network.
24. Very few sites in England have evidence of ongoing predation related suppression of seabird populations, with several having already undergone successful predator control programmes. This reduced options considerably, but remaining areas identified as potentially feasible for an eradication/control programme are the Isles of Scilly and the Channel Islands.
25. The area between the Plémont and La Rocquerelle Headlands, on the north coast of Jersey, was identified as a suitable location; studies into local predator populations have already been carried out and the potential to restore populations to historic levels at the site had been assessed.

## 4 Plémont Seabird Reserve

### 4.1 Background

26. A full feasibility assessment of the potential for predator control at the proposed Plémont Seabird Reserve was carried out by Birds on the Edge for the NTJ in 2021. The feasibility study (Annex 1 – Plémont Seabird Reserve: Feasibility Study Report for a Predator-Exclusion Fence) provides the detailed evidence base for this measure, outlining the current state of the bird populations at the site, the historical context, information on the presence of mammalian predators and the measures required to remove predators from the reserve, as well as ongoing management measures that will be required over the lifetime of the reserve. The planning application (P/2024/1198) has now been submitted by Birds on the Edge Partnership (Birds on the Edge Partnership, 2024).
27. The Project has extended the term of the exclusivity agreement with respect to the funding of the establishment of the reserve from November 2024- November 2025. As part of the extension to the exclusivity agreement, the Project has also funded the Project Officer role for the reserve for a further year between November 2024- November 2025.
28. It should be noted that, at the time of the Project’s initial involvement, there was insufficient commitment to funding to develop the Reserve. However, in the event that compensation is required, the Project has committed to provide the additional funding required to implement and maintain the scheme during the lifetime of the project.
29. The following sections provide further information on the evidence base and roadmap for delivery of this measure, should it be required, including:
- Overview of the selected area, suggested predator control measures and the potential for population growth / scale of compensation delivered (Section 4);
  - monitoring and adaptive management for this measure (Section 5);
  - implementation of the measure (Section 6); and
  - funding of the measure (Section 7).

### 4.2 Overview of the Selected Area

30. The two mile stretch of coast between the Plémont and La Rocquerelle Headlands in northern Jersey, comprises cliffs, promontories, bays and rocky shores. The coastline is backed by steep bracken and gorse covered slopes, bordering onto agricultural land with some small settlements. A public footpath runs through the area and parts of the described area are accessed for swimming, angling, clay-pigeon shooting and other forms of outdoor recreation. The site also hosts a variety of historic sites, including three Neolithic sites, as well as fortifications dating from the Iron age to the Second World War.

31. The entire site falls within the boundaries of the Jersey Coastal Park, which is mostly owned by the Government of Jersey, and managed by the Natural Environment Department. Management for biodiversity mainly consists of the removal of bracken and gorse to encourage growth of native coastal flowers and grasses.
32. Small scale commercial fisheries operate offshore, however resulting impacts on seabirds are low as operations are minimised during the breeding season through a voluntary code of conduct, discouraging access to those areas closest inshore, known as the 'Seabird Protection Zone'.

### 4.3 Predators within the Plémont Seabird Reserve

#### 4.3.1 Summary of Non-Native Predators Present

33. Studies into non-native invasive predators in the Plémont area were carried out by the 'Birds on the Edge' partnership (<http://www.birdsontheedge.org/>).
34. Four species of non-native invasive predators have been identified within the proposed reserve (Brown rat, *Rattus norvegicus*, European hedgehog *Erinaceus europaeus*, feral ferret *Mustela furo* and feral cat *Felis catus*), with studies using traps, direct observations from project officers, thermal imaging, motion-triggered cameras, flavoured wax blocks, and footprint tunnels. Further monitoring of some species has been carried out in order to inform population estimates for the site, or behavioural aspects such as home ranges.
35. Trapping, images from monitoring and trail cameras, and direct observations were the most effective means of identifying predators within the study area, contributing to a total of 408 detections where species could be ascertained.
36. In total, 53 occurrences of brown rat have been recorded (comprising a minimum of 13 individuals), with 219 hedgehogs (32 individuals), 57 ferret (17 individuals) and 7 cats (4 individuals) recorded. Numbers of individuals are likely to be underestimates, especially where individuals of the animals look alike, as trapped animals were not marked during the beginning of the monitoring program. Greatest numbers of each species were encountered around human habitation to the south of the Plémont headland, and towards the eastern end of the proposed reserve.
37. Data informing which of these predators is limiting seabird breeding at the Plémont Seabird Reserve are not currently available. However, it is known that the rise in numbers of ferrets (with none recorded on Jersey 100 years ago but 17 individuals trapped locally in three years recently, (*C. Selares* (Birds on the Edge (BOTE)) pers comms) broadly coincides with the decreases in breeding seabirds at the site.
38. For detailed information on the site baseline please see the supplementary Plémont Seabird Reserve Feasibility Study Report (Annex 1 – Plémont Seabird Reserve: Feasibility Study Report for a Predator-Exclusion Fence).

## 4.4 Control measures

### 4.4.1 Predator-proof fence

39. In order to control non-native predators, a predator-proof fence encircling the area to be protected would be constructed (see Annex 2 – Plémont Seabird Sanctuary: Design Statement for technical specification), and lethal and non-lethal traps designed to capture the species defined in 4.1 would be deployed (see Annex 3 – Plémont Seabird Sanctuary: Management Plan for management plan). Concurrent monitoring of predator numbers using the methods described in 4.3 will allow the success of the scheme to be assessed, and will inform the need for, and nature of, any adaptive management measures to be implemented.

#### 4.4.1.1 Fence specifications

40. The predator proof fence will be built to the following specifications:

- A 2 m high fence including additional hood or cap to deter predators from climbing over.
- A mesh size no larger than 7 mm, with the mesh extending 50 cm from the base to prevent burrowing under
- Sited to follow the contours of the land with a buffer zone containing no vegetation or trees within 4 m (2 m either side of the fence).

41. Fences built to these specifications have been used successfully in predator control measures in New Zealand e.g. the Wharariki Sanctuary (Nature Trust, 2024), Hawaii (Young *et al.*, 2012), Australia (Smith *et al.*, 2020) and the Azores (Benedicto *et al.*, 2019). Full detail on the fence design is included in Annex 2 – Plémont Seabird Sanctuary: Design Statement for technical specification.

#### 4.4.1.2 Fence route

42. The predator-proof fence route is 907 m long along the east coast of Plémont forming an enclosed ecological sanctuary of 3.34 hectares. The application site comprises a narrow strip of land located on the cliffs between Le Petit Plémont and Le Creux Gaborel and it is purposefully well below the coastal footpath. The site forms part of the coastal cliffs with slopes covered mainly in bracken. Some small patches of ivy, bramble and gorse are found amongst the bracken. Full detail on the fence design is included in Annex 2 – Plémont Seabird Sanctuary: Design Statement for technical specification.

### 4.4.2 Predator control

43. For the control of non-native predators, the following suite of measures will be deployed:

#### 4.4.2.1 Rats

- Rat specific kill traps on a grid of 25 – 50 m squares,
- Additional rat specific kill traps at locations where rats have been detected,
- Bait stations at the same spatial scale,

- Feeding stations with live traps where rats have been detected, and
- Baited camera traps, rodent detector cards, dusk surveys and surveys for tracks and signs to monitor for presence (and as such, inform the location of the control measures).

#### 4.4.2.2 Hedgehogs

- Deployment of feeding stations with humane traps in areas occupied by hedgehogs, and other areas of suitable habitat.
- Monitoring through the use of baited traps, dusk surveys and surveys for tracks and signs

#### 4.4.2.3 Ferrets

- Deployment of feeding stations with humane traps in areas occupied by ferrets, and other areas of suitable habitat.
- Monitoring through the use of baited traps, dusk surveys and surveys for tracks and signs.

#### 4.4.2.4 Cats

- Deployment of feeding stations with humane traps in areas occupied by cats, and other areas of suitable habitat.
- Monitoring through the use of baited traps, dusk surveys and surveys for tracks and signs.

44. Further detail on the predator control programme can be found in Annex 3 – Plémont Seabird Sanctuary: Management Plan.

## 4.5 Potential for Population Growth

45. Predator eradication programs have had positive impacts in other colonies (for example Canna, Lundy and the Shiant), but predicting the impact of a successful predator eradication program inevitably carries some uncertainty. This said, the following factors provide confidence that there is potential for population growth as a result of the proposed control measures:

- There is suitable habitat within the proposed Reserve to hold many more breeding auks and other cliff-nesting seabirds than the current numbers.
- Other parts of the Channel Islands hold larger colonies of guillemot and razorbill: there are guillemot colonies of 90 birds on Alderney, 135 on Herm and 235 on Sark; there is a razorbill colony of 58 birds on Alderney (SMP database). This suggests that there is suitable foraging for both species within the species' foraging ranges from the proposed Plémont Seabird Reserve site.
- Numbers of native predators (e.g., gulls) are low in the area, which may also be a factor supporting population growth, although any increase in auk numbers may result in a subsequent increase in native predators.
- Historically, the site has held maximum populations of 300 guillemot and 300 razorbill. With options for habitat management capable of increasing this capacity by creating more accessible ledges for breeding birds, the reserve has the potential to restore guillemot and razorbill populations beyond these historic levels.

46. There is currently no guillemot breeding population at the Reserve, although individual birds are noted in the area, potentially searching for suitable breeding habitat and birds are seen annually below the cliffs at Grosnez point, just west of the site, in the breeding season, with birds noted flying up to the cliffs on occasion. With this behaviour noted, and the regularity of occurrence here and off the reserve in the breeding season, it is possible that breeding is occurring undetected (see Plémont Seabird Reserve Feasibility Study Report (Annex 1)).
47. The population of birds breeding in the vicinity of the Reserve appears relatively stable (42 in 2007, 44 in 2011, 30 in 2014) and these colonies could provide breeding birds for Plémont once the predator eradication is complete. The presence of individuals around the colony in the breeding season suggests that potential breeders do investigate the area.
48. The current razorbill population is 27 individuals on the cliffs at the reserve (data provided by Birds on the Edge post publication of the feasibility report). With slightly different breeding habitat requirements to guillemot, once the eradication of rats and other non-native predators is complete, this small population should be able to expand, and without creating competition for guillemots.
49. Colonisation of new (or recolonisation of historic) natural breeding sites in auks has been well documented.
50. Colonies can grow and expand rapidly where conditions are right (Swann, 1983) and after successful eradication programs, expanding populations of both species will readily occupy new and historically occupied sites (Booker *et al.*, 2018). Large colonies of Brunnich's guillemot and razorbill that were wiped out by hunting pressures have re-established even when source colonies are very distant. Boertmann (2023) describes recolonisation by Brunnich's guillemot where source colonies are at least 270km away, with razorbill recolonising due to an expanding local population. This colony was initially recolonised by kittiwakes, suggesting heterospecific habitat copying may have been a factor in the recolonisation of auks. If this is the case, predator control and subsequent increases in razorbill may also encourage guillemot to recolonise the Plémont site.



#### 4.5.1 Scale of Compensation

51. The scale of compensation that will be delivered by the predator control measure is defined by the current population of guillemot and razorbill in the context of historical peaks. If this measure returns guillemot and razorbill populations to previous maxima, the populations could grow to historical highs of 300 individual guillemot and 300 individual razorbill on the cliffs. Applying a standard correction factor (multiply number of individuals by 0.67) to express this as a population of breeding pairs, compensation quanta could be at least 200 pairs of guillemot and 200 pairs of razorbill from the Plémont Seabird Reserve. The Project's impacts on guillemot and razorbill (using the Applicant's approach) are 25.9 birds (18.2 birds following the introduction of the Offshore Restricted Build Area (ORBA)) and 11.8 birds (10.5 birds following the introduction of the ORBA) respectively. Using region-specific productivity rates published in Horswill and Robinson (2015), and based on historical maxima, populations would be restored to 200 breeding pairs of each species; it is therefore possible to calculate the level of compensation provided per year for each species. Evidence from Lundy shows that colonies grow rapidly following predator eradication, including colonisation in locations where breeding birds had been absent<sup>1</sup>.
52. For guillemot, using the Hornsea 4 approach to calculating compensation, the requirement of 25.9 birds necessitates 110 breeding pairs. Where the colony is restored to 200 breeding guillemot pairs, this would provide compensation at a ratio of 1.8:1.
53. Similarly, for razorbill, using the Hornsea 4 approach to calculating compensation, the requirement of 11.8 birds necessitates 103 breeding pairs. Where the colony is restored to 200 breeding razorbill pairs, this would provide compensation at a ratio of 1.9:1
54. For guillemot, following the introduction of the ORBA, and using the Hornsea 4 approach to calculating compensation, the requirement of 18.2 birds necessitates 77.4 breeding pairs. Where the colony is restored to 200 breeding guillemot pairs, this would provide compensation at a ratio of 2.6:1.
55. For razorbill, following the introduction of the ORBA, and using the Hornsea 4 approach to calculating compensation, the requirement of 10.5 birds necessitates 92 breeding pairs. Where the colony is restored to 200 breeding razorbill pairs, this would provide compensation at a ratio of 2.2:1.

---

<sup>1</sup>[https://lfs-resources.s3.amazonaws.com/j6/LFS\\_Journal\\_Vol\\_6\\_Part\\_5.pdf](https://lfs-resources.s3.amazonaws.com/j6/LFS_Journal_Vol_6_Part_5.pdf)

## 4.6 Connectivity with the National Site Network

56. Both guillemot and razorbill show a high degree of breeding philopatry (>90%), meaning that they return to the same colonies year on year. However, they do show a lower degree of colony philopatry (guillemot = ~50%; razorbill = ~80%). This means that roughly 50% of guillemots disperse away from the colony where they hatched and recruit to non-natal colonies (Swann and Ramsay, 1983; Lyngs, 1993; Harris *et al.*, 1996; Lavers *et al.*, 2007). This is evidenced by some colonies showing very high rates of growth, indicating that immigration into the colony is occurring (Hudson, 1982).
57. Coulson (2016) states that ‘the expression of philopatry is probably variable within a species and is influenced by environmental conditions and population pressures and so should not be considered a constant for individual species’. Detection bias could lead to over-estimation of philopatry (i.e. individuals returning to a studied nesting site are more likely to be detected than those that do not). The prospect of both interannual variation and the overestimation of rates suggests that any published philopatry rates should be treated with caution. This uncertainty, along with the fact that the Plémont Seabird Reserve will be contributing birds back into the biogeographic population, means that philopatry does not need to be considered further (RR-045).
58. Hornsea Four provided considerable evidence of connectivity of guillemot and razorbill in the Channel Islands with North Sea populations and beyond from ringing studies (Ørsted, 2022) which was accepted by the SoS in their decision, such that predator eradication in the Channel Islands has been relied upon as a compensatory measure for guillemot for that project. Ringing recoveries show how some birds raised at FFC SPA (and other North Sea colonies) spend at least some of the non-breeding season in the area around the Channel Islands.
59. Guillemots can recruit to colonies 780km from the natal colony. Dispersal during the non-breeding season between UK North Sea breeding sites and the English Channel is shown to be regular through both kernel density analyses of tagged birds, and ring recoveries. Birds ringed at east coast UK colonies have been recovered in the Channel Islands, suggesting that a proportion of east coast UK breeders end up in Channel Islands waters in the non-breeding season.
60. Razorbill can disperse over bigger ranges than guillemot. Of 314 ringed birds in eastern Canada, 41 dispersed to an island cluster averaging 541 km away, and four dispersed to islands approximately 783 kms away. The largest distance recorded during this study was 1,737 kms, and one razorbill recruited to a colony in eastern Canada from Handa in western Scotland, a distance of 3,200 kms.
61. The provision of suitable breeding habitat in an area used by North Sea breeding birds during the non-breeding season is likely to encourage colonisation of these birds at Plémont. The re-established colonies at Plémont should have high enough productivity to contribute breeding birds to sites within the wider National Site Network. Birds raised at the Plémont Seabird Reserve will have the potential to recruit to the FFC SPA.

## 5 Monitoring and Adaptive Management

### 5.1 Guillemot and Razorbill Implementation Plans

62. An outline monitoring and adaptive management plan has been provided for both guillemot (the Guillemot Compensation Implementation and Monitoring Plan (GCIMP)) and razorbill (the Razorbill Compensation Implementation and Monitoring Plan (RCIMP)) and these plans will be developed further post consent, in consultation with the species-specific Compensation Steering Group (CSG) members, as required. It is anticipated that the relevant elements of Annex 3 – Plémont Seabird Sanctuary: Management Plan for management (submitted as part of the planning application in November 2024) would be incorporated into the GCIMP and RCIMP that are required under Schedule 22, Part 2 and Part 3 of the dDCO(AS1-024).
63. Once the measure is in place, if monitoring suggests that the control programme is less successful than planned, an assessment will be undertaken to establish the reasons for the lack of success and to identify methods of improving the control programme.
64. Support to the Plémont Reserve as a compensation measure may be implemented by the Project alone, however the concept of predator control more generally is currently within the library of compensatory measures delivered through the Collaboration on Offshore Wind Strategic Compensation (COWSC) group, that will be available through the Marine Recovery Fund (MRF) and as such may be delivered at a strategic, cross-project level. The dDCO provides a mechanism for the Project to deliver compensation through the MRF.

### 5.2 Post-Implementation Monitoring

65. Monitoring of both targeted predators and relevant seabirds will be undertaken following implementation of the predator control programme to establish the success of this measure. This will include monitoring for signs of re-invasion and the identification of any increases in predator numbers above target levels.
66. To assess the response of guillemot and razorbill and other seabird species to the predator removal programme, breeding populations and productivity will also be monitored. Data will be compared to pre-eradication levels to assess any changes as a result of the measure. Productivity and any population changes will also be evaluated in a local, regional, and national context, comparing any changes to other guillemot and razorbill colonies to assess the success of the project. This process may involve undertaking seabird censuses and productivity monitoring at other local or regional guillemot and razorbill colonies.
67. Monitoring of the efficacy of the existing voluntary agreement with sea users to restrict close approach to breeding cliffs will allow the assessment of levels of disturbance through seaward disturbance and the potential for strengthening the nature of the agreement, or the visibility of the measure as adaptive management.
68. Annual monitoring of seabird numbers will continue for the operational phase of the Project.

69. The success of the measure relies on the successful eradication of target predators within the fenced area but also the ongoing maintenance of the reserve through maintenance of the fence and sustained biosecurity measures to prevent and deal with re invasion of predators, particularly from rats along the shoreline. The technical elements of the eradication measure have been developed with renowned experts in non-native predator eradication. A fence operational plan, an eradication plan and biosecurity plans will be produced as part of the development of the guillemot GCIMP and RCIMP pursuant to paragraph 4(a) of Part 2 and Part 3 of Schedule 22 of the dDCO (document reference 3.1). It is anticipated that the relevant elements of Annex 3 – Plémont Seabird Sanctuary: Management Plan for management (submitted as part of the planning application in November 2024) would be incorporated into these plans. The risk of reinvasion through the intertidal zone will be adequately addressed through biosecurity measures detailed in Annex 3 which highlights that a biosecurity protocol will be followed that will provide clear context and guidance to follow to prevent, detect and respond to any physical incursion.
70. The biosecurity protocol will feature:
- A detailed list of all known incursion pathways described and categorised with a risk score.
  - A table with all the key actions, resources and responsibilities to ensure that the excluded predators do not re-invade the ecosanctuary via any of the pathways.
  - A monitoring programme will cover the entirety of the ecosanctuary and will feature a combination of passive surveillance and active detection.
  - An incursion response plan to be initiated on a confirmed sign of an excluded species.
71. Further details regarding the monitoring programme can be found in Annex 3 – Plémont Seabird Sanctuary: Management Plan.

## **6 Implementation**

### **6.1 Exclusivity Agreement**

72. As outlined above, the Applicant has entered into an exclusivity agreement with the NTJ with regard to the funding of the proposed Plémont Seabird Reserve project and with the intention to enable full establishment of the Reserve, should compensation for guillemot and/or razorbill be required. The Applicant is currently funding a full time Project Officer role at NTJ who is progressing the planning of the project.

### **6.2 Land Availability**

73. The site falls within the boundaries of the Jersey Coastal Park, which is mostly owned by the Government of Jersey, and managed by the Natural Environment Department. On the 3<sup>rd</sup> March 2024 the Applicant received confirmation from the landowner that land rights are granted in principle to install the fence pending planning approval, to allow the project to continue. This can be found in the Letter from Jersey Government Anti- Predator Fence East of Plémont, Jersey, Channel Islands (document reference PD1-099).

### **6.3 Reserve Establishment and Management**

74. NTJ has previously installed a trial fence in support of the future Planning Application within the area where the reserve would be established. The full planning application for the establishment of the full fence and the Reserve was submitted in November 2024.

75. Public opinion on the establishment of the Reserve has been assessed through a public survey, informed by an awareness campaign including a web page, information boards and leaflets, social media and public events. The results of this survey was presented to the Planning Department as part of the Planning Application.

76. The Planning Application is expected to be determined by Q1 or early Q2 2025, at which time, all planning consents for the establishment of the fence will be in place. As such, following a successful planning application, it is expected that all necessary consents for this measure will be secured prior to the determination of the Project's DCO.

77. Once a lease with the landowner has been secured, construction would be carried out in line with a construction plan, drawn up in consultation with relevant stakeholders.

78. On the completion of the predator-proof fence, the predator eradication plan, drawn up in consultation with stakeholders, would be implemented. This work would run concurrently with a biosecurity strategy (defining schedules for activities such as routine fence maintenance and vegetation management) and a monitoring strategy, aimed at reducing the likelihood of re-invasion, and should this occur, enabling detection and further eradication. Monitoring would also enable the progress of the predator eradication to be charted and may inform when predator eradication can be declared complete. Annex 3 – Plémont Seabird Sanctuary: Management Plan, provides further detail, the relevant parts of which will be incorporated into the guillemot GCIMP and RCIMP pursuant to paragraph 4(a) of Part 2 and Part 3 of Schedule 22 of the dDCO (AS1-024).
79. Following determination of the Project DCO, and if the SoS considers that compensation is required for guillemot and/or razorbill, the Project is confident that the establishment of the proposed Plémont Seabird Reserve can be secured and deliverable prior to the operation of the Project. An indicative programme for the establishment of the Reserve is set out in Table 6-1.

Table 6-1 Indicative timescales for establishment of Plemont Seabird Reserve

Phase	Indicative time based on current project timeline	Task	2024	2025	2026	2027	2028	2029	2030
<b>Project milestones</b>									
Consent	Q3/Q4 2025	Anticipated consent award							
Construction	2027	Start of offshore construction							
<b>Reserve Establishment</b>									
Consent	Q1 2025 (or early Q2 2025)	Anticipated grant of planning permission for full fence							
Establishment	Q4 2025	Construction of fence							
	Q4 2025 onwards	Eradication programme							
Management	Q1 2026	Reserve established and ongoing management implemented							

## 7 Funding

80. The anticipated costs of the development, implementation and ongoing maintenance and monitoring of the proposed Plémont Seabird Reserve are outlined in Table 7-1 below. These costs are also included within the Compensation Funding Statement (document reference 7.9) which outlines how the Applicant and its ultimate parent companies would fund compensation measures should they be required.

Table 7-1 Anticipated costs for establishment and management of Plémont Seabird Reserve

Phase	Cost
DEVEX	£458,384
CAPEX	£1,567,303
OPEX	£3,241,997
Total	£5,267,684



## 8 References

Benedicto, R.J., Hervías, S., de la Cruz, A., Geraldés, P., Costa, L.T., and Gil, A. (2019) 'The socio-economic impact of conservation: the Safe Islands for Seabirds LIFE project' *Oryx*. ;53(1):109-116. doi:10.1017/S0030605317000205

Birds on the Edge Partnership (2024) Planning application to Erect a 907m Predator Exclusion Fence along to coast East of Plémont to create a seabird sanctuary. P/2024/1198. Available at [REDACTED]  
(Accessed 20 November 2024)

Boertmann, D. (2023). 'Re-establishment of an extinct breeding colony of Brünnich's Guillemot *Uria lomvia* in West Greenland'. *Seabird* 35

Booker, H., Slader, P., Price, D., Bellamy, A.J. and Frayling, T., (2018). 'Cliff Nesting Seabirds On Lundy: Population Trends From 1981 To 2017'. *Journal of the Lundy Field Society*, 6, 2018.

Coulson, J. (2016) 'A Review of Philopatry in Seabirds and Comparisons with Other Waterbird Species'. *Waterbirds* 39/3: 229-240.

Defra (Department for Environment, Food & Rural Affairs) (2024). Consultation on policies to inform updated guidance for Marine Protected Area (MPA) assessments. Available [https://consult.defra.gov.uk/offshore-wind-environmental-improvement-package/consultation-on-updated-guidance-for-environmental/supporting\\_documents/090224%20OWEIP%20Consultation%20on%20updated%20policies%20to%20inform%20guidance%20for%20MPA%20assessments .pdf](https://consult.defra.gov.uk/offshore-wind-environmental-improvement-package/consultation-on-updated-guidance-for-environmental/supporting_documents/090224%20OWEIP%20Consultation%20on%20updated%20policies%20to%20inform%20guidance%20for%20MPA%20assessments.pdf). Accessed on 28<sup>th</sup> February 2024.

Harris, M. P., Wanless, S., and Barton, T.R. (1996) 'Site use and fidelity in the Common Guillemot *Uria aalge*.' *Ibis* 138.3 : 399-404.

Lavers, J.L., Ian, L. J., and Antony, W. . (2007) 'Natal and breeding dispersal of Razorbills (*Alca torda*) in eastern North America.' *Waterbirds* 30.4 : 588-594.

Lyngs, P.E. (1993) 'Colony interchange in Baltic guillemots *Uria aalge*'. *Dansk Orn. Foren. Tidsskr.* ;87:247-50.

Nature Trust. (2024). [REDACTED] Accessed March 2024.

Orsted. (2022). Hornsea Project Four Compensation measures for FFC SPA: Ecological Connectivity of Compensation Measures Annex 1 Deadline: 3, Date: 21 April 2022 Document Reference: G3.4.1

Swann, R.L., and Ramsay, A.D.K. (1983) 'Movements from and age of return to an expanding Scottish Guillemot colony', *Bird Study*, 30:3, 207-214, DOI: 10.1080/00063658309476798

Seabird Monitoring Program database, [REDACTED] accessed February 2024

Smith, D., Waddell, K., Allen, B.L. (2020) 'Expansion of Vertebrate Pest Exclusion Fencing and Its Potential Benefits for Threatened Fauna Recovery in Australia'. *Animals*, 10, 1550.

Young L.C., Vanderwerf E.A., Mitchell C., Yuen E., Miller C.J., Smith D.G., Swenson C. (2012). 'The use of predator proof fencing as a management tool in the Hawaiian Islands: a case study of Ka`ena Point Natural Area Reserve'. Honolulu (HI): Pacific Cooperative Studies Unit, University of Hawaii at Manoa. Technical report, 180. 87 pages.

## **Annex 1 – Plémont Seabird Reserve: Feasibility Study Report for a Predator-Exclusion Fence**

# PLÉMONT SEABIRD RESERVE

## FEASIBILITY STUDY REPORT FOR A PREDATOR-EXCLUSION FENCE



Cristina Sellarés de Pedro

2021



ROY OVERLAND CHARITABLE TRUST



DRAFT



## PLÉMONT SEABIRD RESERVE

### FEASIBILITY STUDY REPORT FOR A PREDATOR-EXCLUSION FENCE

Cristina Sellarés de Pedro

Birds On The Edge

This report was prepared for the Birds On The Edge partnership as part of the duties of the project officer.

Partners:

**The National Trust for Jersey**

**Durrell Wildlife Conservation Trust**

**Government of Jersey Natural Environment Department**

Jersey

February 2021

## EXECUTIVE SUMMARY

This report assesses the justification for and feasibility of creating the Plémont Seabird Reserve with the erection of a predator-proof fence around the puffin nesting site, and the eradication of all invasive predators from within. The reserve will be located between the Plémont headland and Grève de Lecq bay, where most of the land is managed by Jersey's Natural Environment Department.

The results of two years of ecological research in the area, as well as an examination of historical bird records, have determined the critical situation of the puffin population in Jersey, standing at four confirmed pairs in 2020. The razorbills fare only slightly better at an estimate of six to eight pairs, and there are no guillemots breeding in Jersey. These three related species of auk used to be found in large colonies in this area in the early 1910s, with 200-300 breeding pairs each. It is believed that their dramatic decline and local extinction over the last century was caused by a combination of predation from introduced mammals and the encroachment of bracken on their grassland habitats.

The Jersey puffins are not nesting in their typical setting of burrows in coastal grassy slopes, but instead are nesting below the vegetation line, on almost-vertical cliffs, in rock crevices, between boulders, and under rocky outcrops. This is probably a result of historical predation and the present effect of invasive predators, which have been found on the slopes above the cliffs. The surveys revealed the presence of four species of invasive mammalian predator in the coastal slopes, namely brown rat, feral ferret, cat and European hedgehog. All are known to predate on adults, chicks or eggs of ground-nesting seabirds in other parts of the world, including various offshore British islands.

The surveys also revealed small populations of other native wildlife classified as threatened or featured in local biodiversity action plans, such as peregrine falcon, European shag, common stonechat, green lizard, and slow worm, amongst others, making this an area of high conservation potential.

The Birds on the Edge partnership believes that to ensure a future for the puffins in Jersey their habitat must be made safe and suitable, and after analyzing the existing management options, this assessment concludes that the most cost-effective and environmentally acceptable solution is to create a Seabird Reserve where the present colony is, to protect it with a predator-proof fence, and to remove any invasive predators found within it. A number of predator-exclusion fences have been established around the world, creating wildlife reserves as 'mainland islands'. Once the fence is completed, the invasive species inside the enclosed area can then be eradicated or relocated using similar methods as the ones used on offshore islands.

Installation of the predator-exclusion fence at the site appears technically feasible, and despite some challenges such as steep rocky cliffs, there is local expertise from geoenvironmental contractors which have done works in similar conditions.

Based on current available techniques and the small size of the proposed site, removal of the invasive predators from the reserve is feasible, with increased chances of success if all four species are targeted at once using a combination of methods. Eradications of a wide range of invasive mammals have been successful in fenced reserves and offshore islands in several countries. Their positive environmental impact is well documented as well methods available to minimise any potential negative effects on non-target species, which are usually short-termed. This work has the support from local wildlife and animal welfare agencies, which will assist with technical advice and take responsibility of feral animals removed from the reserve. The potential for re-invasion will be dependent on the final design of the fence as well as the robustness of the monitoring and bio-security measures put in place.

The fence will have other positive effects for the area's biodiversity by facilitating habitat management works, such as bracken removal and the use of a grazing flock. The removal of invasive predators will potentially result in increased populations of native land birds, reptiles and small mammals, which would otherwise be kept at low densities or eventually extirpated by the invasive predators.

A consultation process has begun and key stakeholders have been identified, however further work will be needed to engage the local community and encourage participation in the decision-making process. This will ensure the continuity and long-term success of this project. Some potential concerns raised by the site users are likely to be related to visual impact and accessibility; these issues should be addressed with the right positioning and landscaping of the fence, as well as the provision of access gates at strategic points.

The benefits of this project to the Island community relate to the enhancement of the natural value of this popular site, and thus promoting physical and mental wellbeing. The project will also generate economic benefits in the form of local employment, training apprenticeships, research opportunities for local students, and increased business in the tourism and hospitality sectors.

The proposed fence will have a maximum approximate length of 2,938 metres and protect 32.3 hectares of land. Conservation fences are expensive projects and the total cost of this fence is estimated at 879,921GBP. The costs of eradication works will be in comparison much lower, but the fence will require long-term maintenance and bio-security measures to prevent re-invasions. It is important to have a clear plan for financing the initial works and the long-term management before embarking on this project.

In summary:

1. The most effective way to ensure the future of the puffins in Jersey is to create a predator-proof fence around their nesting site and to remove all invasive predators from within. The creation of the Plémont Seabird Reserve will also benefit other native species present at the site as well as the plant community and habitats.
2. The Plémont Seabird Reserve will benefit the local community and users of this area with an enhancement of the natural habitats and wildlife at the site, promoting well-being and awareness for conservation of nature. The project will also generate economic benefits for various sectors as well as training and educational opportunities.
3. Based on available techniques developed in similar projects, the installation of the fence and the eradication of invasive predators are technically feasible. This project will require a fence design specialist, high quality materials, detailed operational plans, environmental risk assessments, a robust biosecurity strategy, and a far-reaching public engagement and awareness plan, in order to secure the support of the local community and the key stakeholders.
4. The fence is estimated to be approximately 2,938m long and will protect 32.3ha, costing an estimate 879,921GBP. There will be ongoing costs, albeit significantly lower, as the fence will require maintenance and long-term biosecurity measures.



## TABLE OF CONTENTS

EXECUTIVE SUMMARY	4
TABLE OF CONTENTS	6
LIST OF TABLES AND FIGURES	8
ACKNOWLEDGEMENTS	10
1. INTRODUCTION	11
1.1 Purpose and structure of this report	11
1.2 The Birds On The Edge partnership	11
1.3 The Plémont Seabird Reserve project	12
1.4 Predator-exclusion fence and eradication	13
1.5 The Feasibility criteria	14
1.6 Terminology	14
2. GOAL, OBJECTIVES AND OUTCOMES	16
2.1 Goal	16
2.2 Objectives and outcomes	16
3. STUDY AREA AND RESERVE SITE	18
3.1 General Description	18
Location	18
Habitat	19
Land use and human activities	20
Public access	23
Buildings	23
Weather	26
3.2 Land ownership and management	26
4. SITE ECOLOGY AND CONSERVATION VALUE	29
4.1 Methodology	29
Desk Review	29
Bait and footprint tunnels	29
Puffin watches	30
Seabird population and productivity monitoring	30
Coordinated census	31
Playback survey	31
Bird ringing	32
Automatic acoustic recording	32
Breeding bird survey	33
Dusk surveys with thermal imager	34
Motion-activated cameras	35
Live trapping and mark/recapture	36
Radio-tracking	36
Refugia monitoring	36
4.2 Results	37
4.2.1 Puffins and other seabirds: Past and present populations	37
Puffins	37
Razorbills	41

Guillemots	44
European storm petrels	46
Manx shearwaters	47
Other seabirds	49
4.2.2 Land birds	50
4.2.3 Native reptiles and amphibians	51
4.2.4 Non-target mammals	52
4.3 Conservation value	56
<b>5. THE TARGET SPECIES, IMPACTS AND BENEFITS OF EXCLUSION</b>	<b>57</b>
5.1 The Target Species	57
Brown rats	57
European hedgehogs	59
Feral ferrets	62
Cats	67
5.2 Impacts of the Invasive Mammalian Predators	68
Brown rat	68
European hedgehog	69
Feral ferret	70
Cat	71
5.3 Benefits of Removal and Exclusion of Mammalian Invasive Predators	72
<b>6. FEASIBILITY ANALYSIS</b>	<b>75</b>
6.1 Technical feasibility	75
6.1.1 Construction of a predator-exclusion fence	75
Requirements for an effective fence	75
Location and route options	77
6.2.2 Removal of invasive mammalian predators	80
Brown rats	81
European hedgehogs	82
Feral ferrets	83
Cats	84
Overall likelihood of eradication success	84
6.2 Sustainability	85
6.3 Social acceptability	87
6.4 Political and legally acceptability	93
6.5 Environmental acceptability	95
6.6 Capacity	97
6.7 Affordability	98
6.8 Project contingencies	98
<b>7. ANALYSIS OF AVAILABLE OPTIONS: COSTS AND BENEFITS</b>	<b>99</b>
7.1 'No Action' Option	99
7.2 'Sustained Control' Option	100
7.3 'Fence and Eradication' Option	101
<b>8. CONCLUSIONS</b>	<b>104</b>
<b>REFERENCES</b>	<b>107</b>
<b>APPENDIX I. BIRDS OBSERVED OR ASSOCIATED TO THE STUDY AREA</b>	<b>113</b>
<b>APPENDIX I. MAMMALS AND REPTILES OBSERVED OR ASSOCIATED TO THE STUDY AREA</b>	<b>115</b>
<b>APPENDIX III. OTHER SPECIES MENTIONED IN THE REPORT</b>	<b>116</b>

## LIST OF TABLES AND FIGURES

- Table 1. Objectives and outcomes.
- Table 2. Leisure craft observed in the SPZ between March and July.
- Table 3. Historical buildings in or in the vicinity of the reserve.
- Table 4. Puffin nest summary 2017-2020 (green: nest visits; blue: food delivered).
- Table 5. Results of the seabird census and productivity survey in the study area in 2020.
- Table 6. Land birds recorded during the breeding bird survey in 2020.
- Table 7. Terrestrial mammals in Jersey.
- Table 8. Appearances on cameras 1-6 between 2018 and 2020.
- Table 9. Distribution of species at the camera sites.
- Table 10. Number of rabbits at dusk surveys.
- Table 11. Number and source of records of invasive predators in study area 2018-2020.
- Table 12. Potential negative effects, assessment and mitigation.
- Table 13. Reserve sections (from west to east).
- Table 14. Potential invasive pathways for the Seabird Reserve and summary of prevention strategies.
- Table 15. List of stakeholders identified.
- Table 16. Preliminary list of non-target species that might require mitigation actions to avoid negative impacts.
- Table 17. Indicative costs for the fence installation and eradication works (in GBP)
- Table 18. Costs and benefits of the three main management options.
- 
- Image 1. Population trends of auks in Jersey (1911 to 2020).
- Image 2. Channel Islands and offshore islets.
- Image 3. Location of study area in Jersey.
- Image 4. Close-up of study area (blue square) and reserve site (white outline).
- Image 5. Coastal cliffs between Petit Plémont and the Grand Becquet headland
- Image 6. Plémont headland.
- Image 7. Public footpath between Plémont and Grève de Lecq.
- Image 8. Seabird Protection Zone.
- Image 9. A group on closed-top kayaks following the edge of the Plémont headland.
- Image 10. Two different groups on open-top kayaks exploring the caves below puffin and razorbill nests.
- Image 11. Public footpath (green), inhabited buildings (red), and historical buildings numbered according to Table 3.
- Image 12. Ownership of the land within the proposed reserve boundaries and fence (white line).
- Image 13. Coastal cliffs east of Plémont headland managed by the Natural Environment Department.
- Image 14. National Trust land at Plémont after ground restoration works.
- Image 15. Restored wildlife pond within the National Trust land at Plémont.
- Image 16. Flavoured wax block before bait tests.
- Image 17. Favoured wax blocks after bait tests.
- Image 18. Footprint inked card with bait before test in tunnel TT4.
- Image 19. Footprint inked card with bait after test in tunnel TT4.
- Image 20. Area covered on the playback survey.
- Image 21. Set up of the acoustic recorder.
- Image 22. Location of the acoustic recorder facing east from the Plémont headland.
- Image 23. Audiogram of an oystercatcher's repeated single note call obtained from the recorder.
- Image 24. Breeding bird survey transect route (yellow) and proposed fence boundary (white).
- Image 25. Vantage points (yellow dots) and areas covered by the dusk survey.
- Image 26. Three puffins near the nesting sites in the study area.
- Image 27. Population trend of the puffin in Jersey (1911-2020).
- Image 28. Location of puffin nests in 2020 (red dots), former breeding range (red outline) and boundary of the proposed fence (white line).
- Image 29. A puffin on its way to take several fish to a nest in the study area.
- Image 30. A puffin stands outside its nest (bottom left quarter) below fulmar nests in a rocky vertical cliff in the study area.
- Image 31. A small raft of razorbills below their breeding site in the study area.
- Image 32. Population trend of the razorbill in Jersey (1911-2020).

Image 33. Location of razorbill nests in 2020 (red dots) and boundary of the proposed fence (white).

Image 34. Location of razorbill nests in 2020 (red dots), previous known breeding site in the 1930s (yellow dot) and other sites with sporadic presence during the breeding season.

Image 35. Razorbills standing outside their nests.

Image 36. A razorbill carrying fish to one of the nests in the study area.

Image 37. A guillemot nearby the razorbill and puffin nesting area.

Image 38. Population trend of the guillemot in Jersey (1911-2020).

Image 39. Guillemot former breeding range in Jersey and locations with sporadic presence in 2020.

Image 40. A storm petrel released after being ringed.

Image 41. Two storm petrels in suitable breeding habitat.

Image 42. A Manx shearwater released after being caught and ringed.

Image 43. Manx shearwater breeding habitat in Lundy Island.

Image 44. A lesser black-backed gull with its recently fledged chick.

Image 45. A wren on a motion-triggered camera.

Image 46. A pheasant on a motion-triggered camera.

Image 47. Records and recent observations of reptiles and amphibians in the study area.

Image 48. Rabbit on a motion-triggered camera.

Image 49. Rabbit on a motion-triggered camera.

Image 50. Jersey bank vole on a motion-triggered camera.

Image 51. Evidence of a mole tunnel near the NTJ pond at Plémont.

Image 52. Location and breakdown of the motion-triggered cameras for bait trials and monitoring.

Image 53. Two rabbits on the thermal imager.

Image 54. A brown rat on a motion-triggered camera.

Image 55. Distribution of brown rats in the study area (yellow and red), location of traps (pink).

Image 56. A brown rat in a live trap.

Image 57. Distribution of European hedgehogs in the study area (yellow and red), location of traps (pink).

Image 58. Hedgehog H6 being weighed before release.

Image 59. Hedgehog H6 showing its colour mark (deep purple) as well as a strimmer injury on the centre spikes.

Image 60. Minimum distances travelled by hedgehogs H19 and H13 based on start and end location.

Image 61. Locations of hedgehog H17 within a period of six months.

Image 62. Distribution of feral ferrets in the study area (yellow and red), location of traps (pink).

Image 63. Two ferrets trapped on the same day before release.

Image 64. Ferret wearing a radio-collar before release.

Image 65. Locations of the dens identified (red dots) and proposed fence boundary (white).

Image 66. A selection of locations of ferrets F9 and F13 in relation to the proposed fence boundary.

Image 67. Locations of ferrets F15 and F16 including visits to the proposed reserve area.

Image 68. A large rabbit warren with multiple entrances where various ferrets were located at different times.

Image 69. A den located in a bank on the edge of a horse field in a built-up area.

Image 70. Location and identification of cats recorded in the study area (red) and location of motion-triggered cameras (yellow).

Image 71. A cat on a motion-triggered camera.

Image 72. A cat encountered by the project officer.

Image 73. General pest and predator-proof fence design.

Image 74. Proposed primary route of the predator-exclusion fence at the Plémont Seabird Reserve.

Image 75. Proposed location of the predator-proof fence and partitions for four optional sections.

Image 76. Plémont headland as seen from land.

Image 77. Example of a predator-exclusion fence in a coastal area.

Image 78. Example of a predator-exclusion fence below the line of sight.

Image 79. Example of an access gate on a predator-proof fence.

Image 80. Members of the Rotokare Scenic Reserve Trust maintaining the predator-proof fence that they manage.

All images are copyright of the author unless otherwise credited. The cover and back cover images, as well as images 5,6,14 and 76 have been reproduced with the kind permission of the National Trust for Jersey. All maps produced with a license from Digimap (Jersey) Ltd.

## ACKNOWLEDGEMENTS

This work was carried out for the Birds On The Edge partnership as part of the “Plémont Seabird Reserve project”. The partnership is a joint initiative between The National Trust for Jersey, Durrell Wildlife Conservation Trust and the Government of Jersey’s Natural Environment Department. The Roy Overland Charitable Trust provided funding for the role of the Birds On The Edge Project Officer, who carried out the ecological research, feasibility study, and production of this report.

Biz Bell of Wildlife Management International Limited has been the main technical advisor on all aspects of this study, which has adapted material from her publication “*Feasibility Study for a ‘mainland island’ on Saint Lucia: a pest-proof fence at a mainland site*” (2014).

Additional funding for research equipment and training was kindly provided by:

Mr Gavin Breeze  
The Jersey Ecology Trust Fund  
Jersey Action For Wildlife  
The De Lancey & de la Hanty foundation

Thanks to the following people and organisations:

Vicky Heaney, Darren Mason and the staff of the Isles Of Scilly Wildlife Trust for their training in seabird survey and rodent monitoring techniques.

Dean Jones and Joshua Harris, The Landmark Trust Lundy, for their training in seabird survey, biosecurity and habitat management techniques.

Roland Gauvain, Alderney Wildlife Trust, for facilitating the participation in the annual puffin burrow survey in Burhou.

Margaret and Richmond Austin, Channel Island Ringing Scheme, for preparing and providing ringing records upon request.

Antony Paintin, Ornithology Section, Société Jersiaise, and Lynda Higgins, Société Sercquaise, for facilitating access to historical bird records.

Tim Liddiard, Nina Cornish, Lindsey Napton, Liz Walsh, Keiran Luce and the rangers of Natural Environment Department, for sharing their expertise on habitats and survey methods, and for their assistance in the installation of artificial nest boxes.

Amy Hall, Jersey International Centre of Advanced Studies, for providing training and advice on trapping mammals.

Geomarine Jersey rope team, for their assistance in the installation of puffin nest boxes in cliffs.

The National Trust for Jersey staff and associates, for sharing their expertise, and the Lands Team for their assistance in clearing habitats and installing artificial nest boxes.

Fieldwork was carried out with the kind permission of private landowners and the Natural Environment Department, Government of Jersey.

The animal trapping and bird ringing work were carried out with licences from the Natural Environment Department and the Channel Island Ringing Scheme.

# 1. INTRODUCTION

## 1.1 Purpose and structure of this report

This feasibility study and report have been produced on behalf of the Birds On The Edge partnership by its project officer Cristina Sellarés de Pedro, who is employed by the National Trust for Jersey with a grant from the Roy Overland Charitable Trust.

The purpose of this feasibility study is to assess the feasibility of erecting a predator fence near the puffin colony in the north coast of the Island and eradicating all terrestrial invasive predators from inside the fenced area, as well as to evaluate its potential as a measure of long-term protection for the colonies of puffins and other endangered seabirds. The present report addresses three questions: 1) why this needs to be done, 2) whether it can be done, and 3) how it can be done.

The remainder of the Introduction section offers context to the Birds On The Edge partnership and the role of the fence and eradication within the BOTE Plémont Seabird Reserve project, as a summary of key findings of the first two years of this project, the rationale behind the proposal at the centre of the feasibility study, an insight into the feasibility criteria and a guide for the terminology used. Chapters 2, 3 and 4 cover the aims of this work and describe the study area and proposed reserve site, from a natural, historical and cultural perspective. Here are presented the results from various ecological surveys highlighting the status of the puffin colony as well as many other native species in the area. Chapter 5 focuses on the presence and distribution of the four target species, the invasive predators found in the area, with an overview of their known impact on island ecosystems in British Islands and around the world. Chapters 6, 7 and 8 present the feasibility analysis itself, the costs and benefits of alternative management options, and the final conclusions of this report.

This report will be made available on first instance to all relevant Government of Jersey authorities as well as St Ouen's Parish authorities, key stakeholders and affected parties.

## 1.2 The Birds On The Edge partnership

The Birds On The Edge partnership (BOTE hereafter) is a joint initiative between the National Trust for Jersey, the Government of Jersey Natural Environment Department, and Durrell Wildlife Conservation Trust, which aims to restore Jersey's coastal lands and habitats, in order to enhance local populations of rare or endangered birds.

In other to achieve its goals, BOTE works with many local and international organisations such as the Société Jersiaise, the Jersey National Park, the Channel Islands Bird Ringing Scheme (CIBS), the Jersey International Centre for Advanced Studies (JICAS), the Jersey Biodiversity Centre, the British Trust for Ornithology (BTO) and various authorities and counterparts in other Channel Islands.

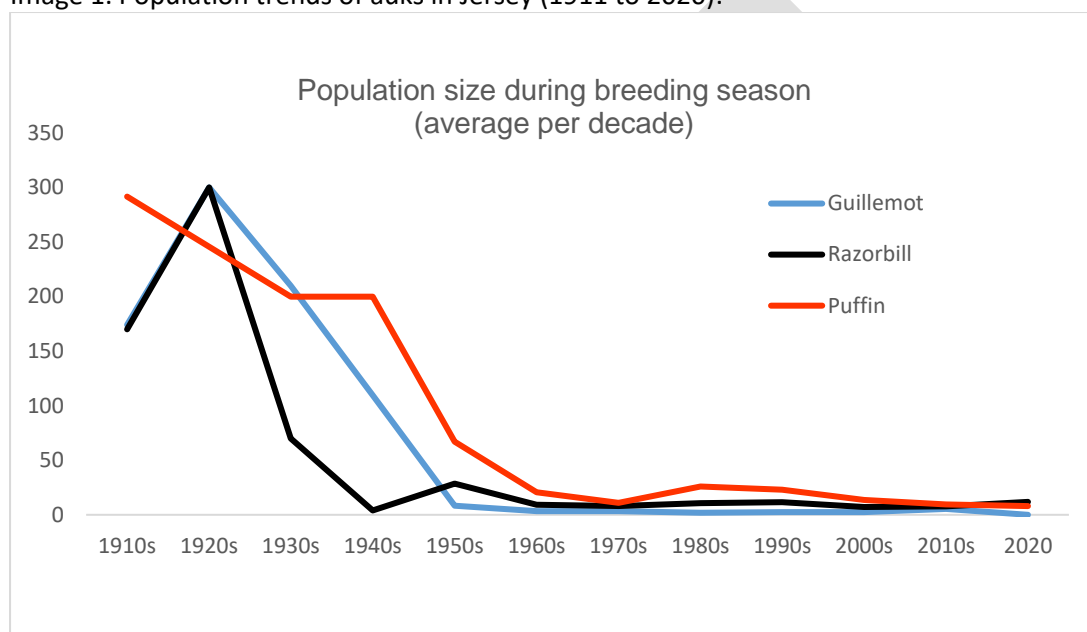
Since its inception in 2011 BOTE has launched, promoted and undertaken local projects in a variety of fields of nature conservation and management, such as habitat surveys, grassland restoration, bracken management, conservation grazing, long-term bird monitoring schemes, farmland bird monitoring and conservation, hedge restoration, winter bird crop management, species re-introduction, and several awareness campaigns and community-focused events<sup>1,2,3</sup>. Over the years the work of BOTE has been financially supported by a wide array of local and international funding bodies, government grants, charities, trusts, and private donations. In 2018 the Roy Overland Charitable Trust granted BOTE with funding which allowed the Project Officer to undertake a three-year study and management trial based in the north coast of Jersey. The National Trust petitioned

for this funding in view of the need to ascertain the ecological status and conservation needs of the Island's puffins, in order to safeguard their future and that of many other endangered seabirds. The work focused on gathering information on Jersey's puffin colony, as well as trialling ecological monitoring techniques and producing a long-term management plan for the species and its habitat.

### 1.3 The Plémont Seabird Reserve project

The work carried out by the BOTE project officer has revealed the unequivocal declines suffered by puffins and other seabirds in Jersey over the last century, in particular the two other species of auks which used to breed alongside puffins (Image 1). The project has also evaluated the habitat of many of the coastal areas where the present colonies survive, and quantified the presence, abundance and distribution of all known invasive terrestrial predators which can potentially threaten seabirds and other native wildlife (chapters 4 and 5).

Image 1. Population trends of auks in Jersey (1911 to 2020).



It is believed that the present location of the remaining puffin nests, which is mostly on the cliff-faces below the slopes, and amongst rocky boulders and crevices instead of burrows, reflects historical predation which would have happened at nests on the slopes. This is the natural puffin breeding habitat, however in Jersey these slopes are fully accessible to the terrestrial invasive predators found during the survey, with some of them encountered directly on the slopes. Therefore, it is likely that past predation pushed the puffins to nest in the cliff faces and that presence of predators on the slopes puts a swift end to any attempt from the puffins to re-colonise the slopes.

In view of this evidence, and of the importance, ecologically and culturally, of preserving the only remaining colony of puffins in Jersey, BOTE proposes the establishment of a reserve which will safeguard the species, its habitat and its other native inhabitants for the long term.

The reserve aims to encompass the puffin colony and its former breeding grounds, covering portions of land and sea. The land portion will be made safe and suitable with the installation of a predator-exclusion fence, the removal of the invasive predators and the management of bracken manually and by a grazing flock. The seabird breeding opportunities in the reserve can be enhanced with the installation of nest boxes or artificial burrows for puffins and other burrowing seabirds, and made

more attractive to colonial breeders with the display of decoys or by playing audio recordings from other colonies. By sea, the protection of the colonies will involve awareness and education campaigns to reduce disturbance caused by human activities, and a biosecurity protocol to prevent re-invasion of predators. The present feasibility study report relates to the management of the invasive predators in the reserve; therefore, it covers the installation of the fence and the eradication of predators from within.

Works related to the reserve have already started on some fronts, such as raising awareness with education materials, boards, during public various events, and via de media. Habitat management is ongoing at the Plémont headland as well as the National Trust for Jersey land, and puffin nest boxes have been installed and provided with monitoring and biosecurity measures in two areas near existing nesting sites<sup>4</sup>.

#### **1.4 Predator-exclusion fence and eradication**

Introduced species have had, and still are having, devastating impacts around the world. These impacts are magnified on islands, where species have evolved and adapted to live without land-based predators such as mammals. These introductions have caused the decline and even extinction of many species through competition of resources, direct predation and changes in their habitats<sup>5,6,7,8,9</sup>.

There exists a vast and detailed body of work that covers eradications of invasive mammals such as rats from islands, in particular from New Zealand where such techniques were trialed and perfected. These techniques have become increasingly complex and ambitious, and therefore expensive. The price tag and the possible impact on local communities means that conservation managers need to justify the benefits of such projects and provide scientific evidence that warrants such extreme measures. However, when applied correctly, the results can be extremely beneficial, as proven by for most of the eradications that have been successful to date.

When a landscape or area is too large, or too complex to attempt a full eradication, such as the case of Jersey itself, the best alternative is the use of a predator-exclusion fence. Predator-proof fencing is a proven technology developed in New Zealand, with more than 50 fences been constructed to the date. These fences are capable of excluding animals as small as a baby mouse and are designed to prevent animals from digging under or climbing over the fence.

By protecting an area of land, the fence is effectively creating a 'mainland island' or even a 'coastland island' when it is by the sea. There are nowadays a number of 'mainland islands' around the world, from Australia and New Zealand where the technique was first developed, to Hawaii or the Azores, their design varying in accordance to the target invasive species, the characteristics of the site, the native species to be protected and social factors such as the involvement of the community and the public access strategy.

Once the fence has been installed, techniques developed for offshore islands to remove invasive predators and to prevent re-invasions can be applied and adapted to the fenced reserve with expected positive results.



## 1.5 The Feasibility criteria

The Pacific Invasives Initiative<sup>10</sup> is an international organisation compiling and providing technical expertise for conservation specialised in islands, including seabird conservation, invasive eradications and habitat restoration. Its guidelines propose that for an eradication project to be successful it must fulfil the following seven criteria:

- **Technically feasible:** Can the technique(s) be used at the project site to remove all individuals of the target populations?
- **Sustainable:** Can we prevent re-invasion of the target species?
- **Socially acceptable:** Does the project have full support from the community and other key stakeholders?
- **Politically and legally acceptable:** Will we be able to secure all required permits and consents?
- **Environmentally acceptable:** Can we ensure a manageable impact to the environment?
- **Capacity:** Do we have, or can we acquire all the required skilled people, resources and equipment?
- **Affordability:** Will we be to secure the required funding?

## 1.6 Terminology

**Native:** The International Union for the Conservation of Nature (IUCN) defines 'native' or 'indigenous' as species or race that occurs naturally in an area<sup>11</sup> (for this review, Jersey), i.e. whose dispersal has occurred independently of deliberate human translocation.

**Non-native:** The Joint Nature Conservation Council (JNCC) defines 'non-native', 'non-indigenous', 'alien' or 'exotic' as a species or race that does not occur naturally in an area, i.e. it has not previously occurred there, or its dispersal into the area has been mediated by humans<sup>12</sup>.

**Introduction:** The deliberate or accidental release by human agency of an organism(s) into the wild by humans in areas where the species or race is not native<sup>12</sup>.

**Feral:** An animal (or its descendants) that has been kept in domestication but which, following escape or release, now lives in the wild state.

**Naturalised:** A non-native species or race that, following escape or release, has become established in the wild in self-maintaining populations.

**Invasive non-native species:** A species which has been introduced into areas outside its natural range through human actions and is posing a threat to native wildlife.

**Target species of this project:** All the invasive non-native terrestrial predators found within the reserve area, which the project aims to remove and exclude from the reserve. Namely the brown rat, feral ferret, European hedgehog and cat.

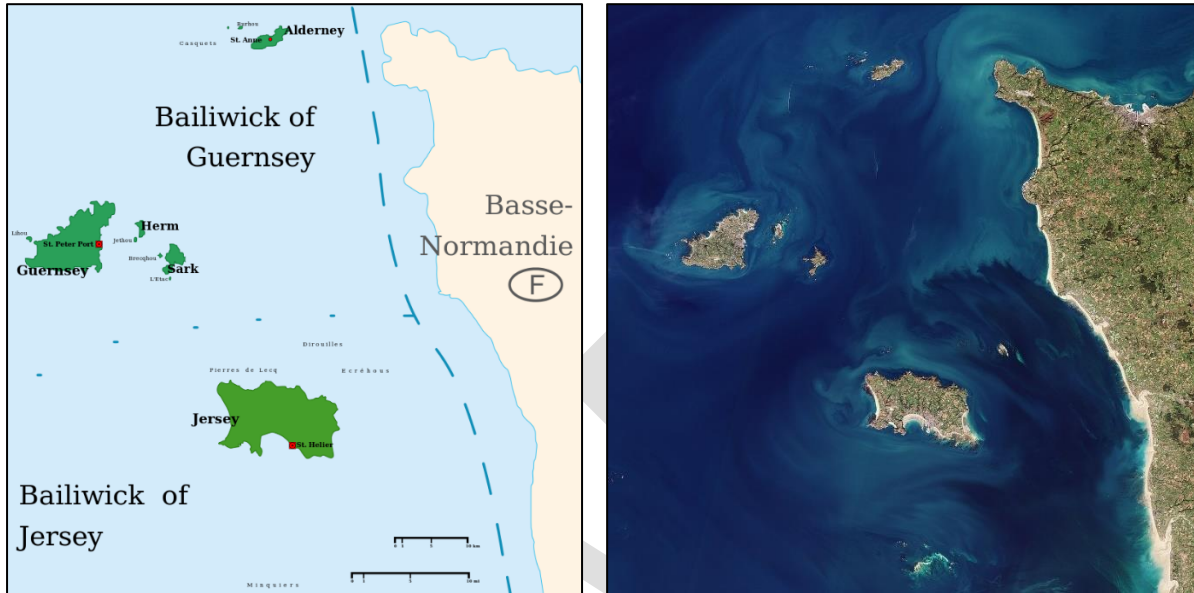
**Eradication:** The complete removal of a species from a location into which there is little chance of reinvasion by natural dispersal.

**Control:** A reduction of the population size of a species, by sustained and constant effort.

**Reserve area / site:** In this report, the portion of land and sea designated as the Plémont Seabird Reserve. On land, it refers to the area protected by the predator-exclusion fence (Image 3).

**Study area:** The area comprising the reserve area and surrounding grounds where the different ecological surveys were carried out. The study area did not have exact boundaries as some of the research involved observing and recording animals far from where the observer was standing, or radio-tracking animals to wherever they had travelled. Most of the research was carried out within 50-200m from the boundary of the proposed reserve; however, on certain instances it extended several kilometers from it (Image 3).

Image 2. Channel Islands and offshore islets.



© "Hannes2" in Wikipedia <https://commons.wikimedia.org/w/index.php?curid=1506987>

Image 3. Location of study area in Jersey.



## 2. GOAL, OBJECTIVES AND OUTCOMES

### 2.1 Goal

The goal of the proposed project is to install a fence around the Plémont Seabird Reserve and to remove all known invasive mammalian predators from within, in order to create a safe and suitable breeding habitat for Jersey’s puffins and any other native species in the reserve.

### 2.2 Objectives and outcomes

The objectives that the project will achieve and the outcomes that will be produced as a result of achieving the objectives are:

Table 1. Objectives and outcomes.

OBJECTIVES	OUTCOMES
1. Build fence	1.1 Planning documents and permission secured.
	1.2 Invasive mammalian predators are now isolated.
	1.3 Puffins and other seabirds are protected from re-invasions.
	1.4 Wildlife is protected from unmanaged human disturbance.
2. Remove invasive predators	2.1 New conditions support increase of puffin colony and expansion of breeding grounds.
	2.2 Re-establishment or colonization of new members of the seabird community such as storm petrels and Manx shearwaters.
	2.3 Increase in population of a range of terrestrial passerine species and other predation-vulnerable terrestrial and waterfowl species, breeding and migrant.
	2.4 A re-balancing of natural predator-prey relationships and a bottom-up trophic effect within the reserve occurs.
	2.5 Increase in population of terrestrial reptiles, amphibians, mammals and invertebrates and overall increase in biodiversity and species richness.
	2.6 Production of mitigation plans for non-target native species partially isolated within fenced area.
	2.7 Biosecurity measures are functional and effective.
	2.8 Re-invasion detection measures continue in perpetuity and response measures are in place to direct the management of re-invasions.
2. Restore seabirds’ breeding habitat	3.1 Invasive bracken and other dominant plant species are removed or reduced.
	3.2 Conservation grazing flock is introduced.
	3.3 Plant community diversity increases.

4. Develop a participatory decision-making process	4.1 Partners have a regular forum in which to share information, collaborate on planning, and coordinate future project components.
	4.2 Fence installation and invasive predator eradication are developed and implemented with the community's awareness and support.
	4.3 The Island's community supports ongoing biosecurity and response measures.
5. Community engagement	5.1 Seasonal awareness campaigns are conducted and education materials are produced and made available to all Islanders.
	5.2 The community has an ongoing role in the decision-making process and management of the reserve.
	5.3 Participatory activities and events are organised to reach and engage a wide variety of demographic sectors of the community.
	5.4 Internship and research opportunities for local students emerge.
6. Improve quality of life and livelihoods on Jersey	6.1 Increased access to natural habitats and wildlife and encouragement of outdoor activities.
	6.2 Increased regulation and facilitation of commercial leisure activities and work in coordination with local operators.
	6.3 Hospitality businesses in the Seabird Reserve vicinity benefit from an increase of tourism and public visiting the area.
	6.4 Additional capacities are developed within the Channel Islands that relate to the planning and implementation of complex large-scale conservation projects.
7. Leverage each partners' existing capacities and build new ones in line with their missions	7.1 Each partner in the BOTE partnership has a role and contributes to the project's larger goals accordingly, leading to success of the overall project.
	7.2 Each partner develops additional capacities that facilitate implementation of their mission in the future.

### 3. STUDY AREA AND RESERVE SITE

#### 3.1 General Description

##### Location

Jersey is the largest of the Channel Islands, with a total area of 118km<sup>2</sup> (45.6 square miles) and 70km of coastline (43 miles). It lies 30.5km from the Normandy Coast, France, and approximately 137km south of Great Britain. Jersey is inhabited and, with 107,800 people in 2019<sup>13</sup>, densely populated. Jersey's coast is subject to the large tidal range of the Bay of St Malo, the 3<sup>rd</sup> largest in the world. Most seabird colonies in Jersey are found in coastal cliffs, which occupy the landscape of the north coast as well as the south-west portion of Jersey.

The study area is a two mile-stretch of coast in the north-west of Jersey, between the Plémont Headland and La Grève de Lecq Bay. The study area expands inland, into agricultural and urban areas depending on the research that was carried out. Some parts of the research took the project officer several kilometres from the edge of the proposed reserve, in particular when radio-tracking target invasive mammals.

The proposed site for the Plémont Seabird Reserve is the coastal strip of land between Plémont Bay and La Rocquerelle headland, coming inland to the cliffs below the public footpath. This is the only place in Jersey where puffins and razorbills breed. The proposed reserve would fall within the boundaries of the Jersey Coastal Park<sup>14</sup>.

Image 4. Close-up of study area (blue square) and reserve site (white outline).



## Habitat

The study area comprises steep cliffs with inlets, caves, rocky outcrops, headlands, promontories, and wooded valleys descending onto the sea from the interior land and farmland.

A 2020 ‘character assessment’ of Jersey’s landscape and seascape commissioned by the Government of Jersey placed the study area within the Grève de Lecq “coastal until”, which featured the following principal Character Types<sup>15</sup>:

- Cliffs and headlands
- Interior agricultural plateau
- Rocky shores and bays
- Offshore reefs and islands
- Shallow sea
- Deep sea

The cliffs from Petit Plémont eastwards (Image 5) are like most of Jersey’s north coast: slopes covered in dense and tall bracken - and as such, with little other species associated. Amongst the bracken there are occasional patches of gorse, bramble, and blackthorn, especially where the soil is shallower and around rocky outcrops. Along the coast, the bracken reaches almost all the way down to the cliffs, occasionally transitioning to an edge of coastal grassland where exposed rock and rocky outcrops start to appear. Closer to the top of the cliffs and near the footpath clumps of Holm oak and other trees can be found.

Image 5. Coastal cliffs between Petit Plémont and the Grand Becquet headland.



The terrain and habitats of the Plémont headland (Image 6) are distinctively different from the coastal cliffs. The bracken is much less vigorous and mixed with a larger number of species, and there are distinctive areas of different plant communities. These include short coastal grassland, gorse patches, open swards of short grasses and little tussocks, rocky ground with various lichens, as well as bracken-dominated areas, sometimes co-associated with bramble and other species<sup>16</sup>. The striking difference in landscape, character and communities of the Plémont headland, compared to

the rest of the study area, is possibly caused by its topography, exposed location, areas of shallow soil and the grazing effect of rabbits, which are abundant in the headland.

Image 6. Plémont headland.



### Land use and human activities

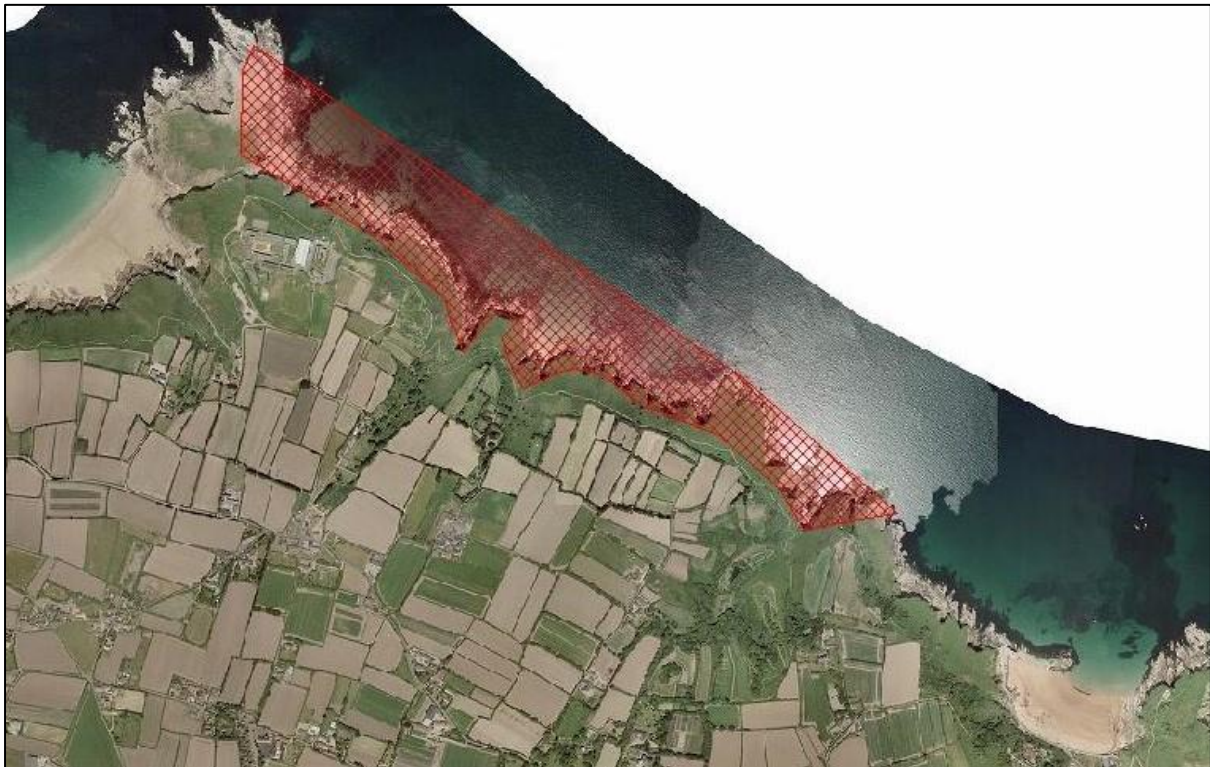
The agricultural fields above the cliffs are mainly used for growing seasonal commercial crops, silage crops or for grazing. The footpath (Image 7) is used by walkers and sight-seers, as well as by educational organisations and charities, professional dog-walkers and tour operators, with a focus on the landscape and the wildlife of the area. There is a small area below Petit Plémont where people sometimes swim, and four main sites where hobbyist anglers fish from. There is a clay target shooting range on a small headland, near the eastern end of the reserve.

Image 7. Public footpath between Plémont and Grève de Lecq.



The waters around the reserve are commercially fished with lobster and crab pots, as well as line-fishing from the boats when the conditions are suitable. This small fishing fleet is mainly based at Grève de Lecq bay. Recreational activities such as coastering and trips on paddle board, kayak and boat also happen in the area, both commercially and privately. However, there is a voluntary code of conduct known as the Seabird Protection Zone<sup>17</sup> (SPZ hereafter), which encourages people to avoid visiting the waters in the area during the seabird breeding season altogether, in order to avoid causing damage or disturbance to wildlife (Image 8).

Image 8. Seabird Protection Zone.



Although no surveys were undertaken with the sole aim to record human presence and activity in the study area, ad-hoc observations were recorded during the other surveys, especially of any activity in the Seabird Protection Zone. This research found that most tour operators avoided commercial activities within the SPZ during the breeding season, and that the commercial fishermen avoided unnecessary disturbance by limiting their time spent in the SPZ to the minimum necessary. The general public, however, visited the SPZ with boats, kayaks and paddle boards in several occasions during the breeding season (Table 2). A survey conducted amongst recreational users at the launching point of Grève de Lecq revealed a lack of awareness of the SPZ amongst the majority. A small minority of people revealed their knowledge of the SPZ, but expressed lack of concern based on the belief that their behaviour did not have a negative impact on the seabirds breeding in the SPZ.

Table 2. Leisure craft observed in the SPZ between March and July.

	2018	2019	2020
<b>N craft</b>	12	41	46
<i>during watches</i>	10	7	13
<b>N watch hours</b>	103.5	145.5	61.0



Despite not having a large sample size, with only three years of watches in the study area, an interesting difference was noted between the years 2019 and 2020, with the latter year seeing an increase of 360% in the number of leisure craft observed per hour in the SPZ during the breeding season (Images 9 and 10). It is not possible to ascertain the exact reasons for this increase without a survey amongst the craft users, however it is believed that it could be the result of the restrictions on other activities which were present at the time as part of the Government of Jersey Covid-19 pandemic strategy.

Image 9. A group on closed-top kayaks following the edge of the Plémont headland.



Image 10. Two different groups on open-top kayaks exploring the caves below puffin and razorbill nests.



## Public access

There is a public footpath between Plémont and the shooting range. The walking route eastward then veers inland via small lanes and between buildings, until it becomes a footpath again overlooking Grève de Lecq Bay. The path descends on the west side of the bay and ends behind the Prince of Wales establishment. The Plémont headland itself has no public footpath, but people access it by foot via a steep slope (Image 11).

Image 11. Public footpath (green), proposed fence location (white), inhabited buildings (red), and historical buildings numbered according to Table 3.



## Buildings

There exists a large number and variety of human-built structures in Jersey that reflect the activities and settlements across the ages. The first known human presence in Jersey dates back from 250,000 years ago, when nomadic groups used natural caves in the south-west for hunting activities. The first structures were built by settled communities during the Neolithic period, and comprised a variety of burial sites such as dolmens and passage graves across the Island.

There are three Neolithic sites in the vicinity of the study site<sup>18</sup> (Table 3). Within the proposed reserve there are components of the Stroingpoint Plémont, a multi-levelled defence complex built during the Occupation (Image 11).

Table 3. Historical buildings in or in the vicinity of the reserve.

<p>1. La Hougue Des Géonnais</p>	
<p>A Neolithic passage grave built around 6,000 years ago. Comprises a passage leading into a large open rectangular chamber. This monument had been damaged during excavations. Stones taken by quarrymen were replaced later with granite blocks to show their original position. Finds at the site also include flint tools, decorated pottery fragments and broken querns<sup>22</sup>.</p>	
<p>2. La Hougue Le Bêqui</p>	
<p>A large megalithic structure below a mound, found on private land and first mentioned in 1817. Type of site and date are presently unknown. Findings consisted of a capstone, twelve slabs, pottery and cremated bones<sup>20</sup>.</p>	
<p>3. La Rue de la Croute Tummulus Site</p>	
<p>Classified as an area of archaeological potential, it is believed to contain a mound of earth or stones raised over a grave, also known as a tumulus<sup>21</sup>.</p>	

4. Plémont Fort & Guardhouse



Believed to have been originally an Iron Age fort, the British built a guardhouse at this site in the 18th Century. The structure was substantially altered during the Occupation, turning it into a "reinforced field position", with an emplacement with twin machine guns, a mortar emplacement, an observation post, and a searchlight shelter with railroad tracks that could move the searchlight to twin operating sites.

5. Plémont Guardhouse Tower



This small tower, part of the original Guardhouse construction, was modified by the Germans during the Occupation and fitted with an anti-aircraft machine gun.

6. Machine Gun Emplacement and Shelter



This small bunker was built during the Occupation and is situated on the cliffs below the main M3 observation bunker.

7. M3 Army Observation Bunker



This large building dominates the complex from the high ground. It was designed to watch the coast and provide coordinates to the Artillery Batteries across the island.

The buildings in present use which are closest to the study area are private houses, farm buildings and sheds, and a café. The inhabited building closest to the reserve is at 329m distance of its edge.

### Weather

The prevailing westerly winds are often felt in the study area, especially in the headlands. The annual average wind speed is 15.2mph (data from Jersey Airport and Weatherspark statistics, 1980-2016). The Island can occasionally experience violent storms (64-72mph, Beaufort scale) which can be accompanied by gusts of hurricane-strength winds (73mph and over). The average rainfall in Jersey is 875mm/year (Jersey Met Climate Statistics, 1981-2010), with the wetter season being from September to March.

### 3.2 Land ownership and management

The portion of land comprising the Plémont headland and the cliffs below the public footpath, up to the eastern edge of Creux Gabourel headland, is owned by the Government of Jersey and is managed by the Natural Environment Department. The Department manages the land to promote biodiversity and to prevent encroaching of dominant vegetation such as bracken, gorse and bramble. These species are cut seasonally in certain areas, in order to open up ground and to help establish coastal grasses and native flowering plants (Image 13). A smaller portion of coast, found between the shooting range and Grève de Lecq, is also publicly owned. The land below the public footpath between Creux Gabourel and the shooting range is the Common du Fief de Vinchelez, and is privately owned (Image 12).

Image 12. Ownership of the land within the proposed reserve boundaries and fence (white line).

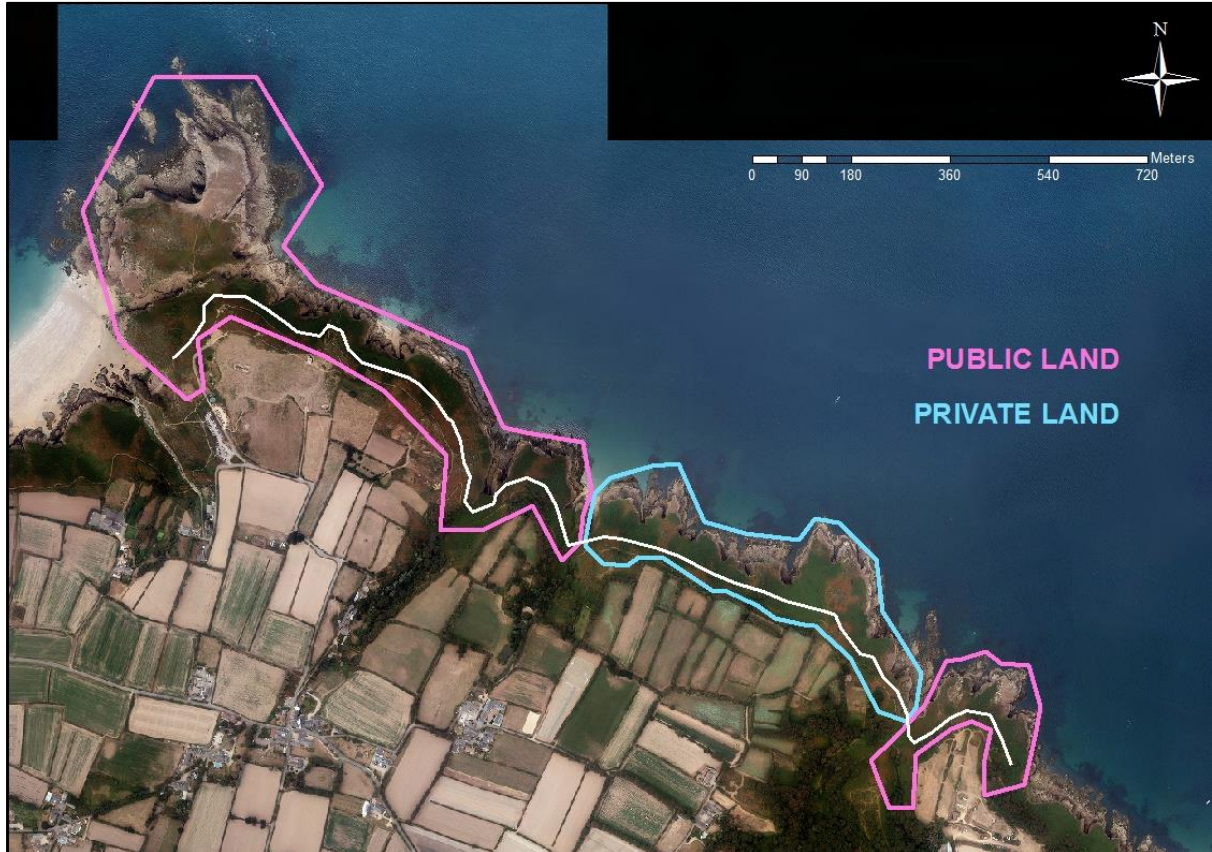


Image 13. Coastal cliffs east of Plémont headland managed by the Natural Environment Department.



The fields and land parcels above the footpath belong to various landowners. Amongst them is the National Trust for Jersey, which owns the land above the Plémont headland and adjacent fields, as well as a portion of coast between the shooting range and Grève de Lecq Bay. The Trust removed the derelict holiday camp buildings from this land and now manages it to increase biodiversity and restore coastal habitats. The agricultural fields owned by the Trust are leased to cattle or sheep farmers for grazing (Images 15 and 16).

Image 14. National Trust land at Plémont after ground restoration works.



Image 15. Restored wildlife pond within the National Trust land at Plémont.



## **4. SITE ECOLOGY AND CONSERVATION VALUE**

### **4.1 Methodology**

From 2018 until December 2020 a variety of ecological surveys were conducted in the study area as well as desktop reviews. The main aim of this research was to establish the historical decline and status of the puffin population and its present threats, but it also included other native species and threatened wildlife, as well as the presence and activities of people in the study area.

Some methods were species-specific, such as the puffin watches, whilst other methods collected data more broadly, such as motion-triggered cameras or baited live traps. Most of the research was carried out in the study area, and where possible the results distinguished between data originating from within the proposed boundaries of the reserve and data from outside of the reserve.

#### **Desk Review**

All available records of puffins, razorbills, guillemots, fulmars, storm petrels and Manx shearwaters in Jersey were compiled and analysed. Most of the records were obtained from the Société Jersiaise Ornithology Section, by examining more than 30 folders, scrapbooks and hand-written notes where the records are physically stored. Other sources were the reports from local seabird census, as well as individual reports of sightings from social media. Information from other Channel Islands was also collected when available.

#### **Bait and footprint tunnel trials**

During the months of October, November and December 2018 trials were undertaken to determine the validity of different monitoring techniques used in similar projects such as in the Isles of Scilly and Lundy Island. These techniques have been designed to detect the presence and distribution of rodents and other small mammals, and as such can be applied to study invasive predators. The techniques that were trialled were motion-triggered cameras, flavoured wax blocks, and footprint tunnels. Most of the techniques were monitored by motion-triggered cameras as a means to test their effectiveness and uptake by mammals in the area.

Flavoured wax blocks were used with the aim to study the bite markings on the block and to identify the species based on the tooth indentations. Other baits such as peanut butter and sardines were also used to test the reliability of the cameras. The footprint tunnels consisted of a long and narrow piece of white card which had inked edges and bait placed in the middle of it. The card was fitted inside a small black plastic tunnel. It was hoped that this technique would provide feet imprints of small mammals, as they would step onto the inked areas and leave footprints on the white cardboard when walking towards the bait. Typically, species can be identified from clear prints of feet, based on the shape and size of the foot as well as the stride distance.

During the trials these two techniques did not appear useful for their purpose, mainly due to the presence of non-target species which interfered with the results. As the cameras showed, the wax blocks were chewed by many species, such as field mouse, brown rat and bank vole, leaving the blocks so damaged that it was difficult to ascertain if the target species, mainly the brown rat, had been involved (Images 16 and 17). Similarly, the cardboards used on the footprint tunnels were heavily marked with footprints of many species and this made the identification of the target species very difficult. In the worst cases, the cardboards themselves were found chewed to small bits (Images 18 and 19).



Images 16 and 17. Favoured wax blocks before and after bait tests.



Images 18 and 19. Footprint inked card with bait before and after test in tunnel TT4.



### Puffin Watches

Watches were carried out at the puffin breeding sites each year between February and July of each year in order to determine presence, numbers, activity and breeding success. During the watches the same data was collected for razorbills, as many razorbill pairs could be monitored from the same watchpoints. Early on, watches required intensive working days with long hours, but each subsequent season the methodology was streamlined. By 2020 the information collected was reduced to census and productivity and the watches evolved into a locally-devised seabird monitoring strategy for population and productivity, which also covered all other seabirds breeding in the study area (see next).

### Seabird population and productivity monitoring

In 2020 all seabirds breeding in the study area, including puffins, were monitored following a methodology adapted from the Seabird Monitoring Handbook of Britain and Ireland (1995 JNCC)<sup>22</sup>. The purpose of this survey was to collect the required information regarding the present puffin population but also to establish a baseline of data on population size and breeding productivity for all seabirds at the study area. This data then could be compared against previous seabird census carried out island-wide, as well as used to monitor the effect of future habitat and wildlife management works.

This survey consists on mapping each potential seabird nest via photographs, and checking each at specific times of its life cycle in order to determine breeding status and success. The method can be applied to most diurnal seabirds but the timing varies depending on the species. This particular method was designed to monitor small colonies which cannot be accessed by land, therefore it was found to be the most suitable for the study area.

### Coordinated census

In order to ascertain the size of the puffin population during the breeding season, between one and three coordinated counts were carried out each year across the study area, by a team of volunteers watching from different vantage points. The aim of these counts was to record all puffins observed at sea or on land, and to cover as many blind spots as possible in order to ensure that all the breeding population was accounted for. Razorbill numbers and guillemots, if present, were also recorded during these counts.

### Playback survey

The method of recorded playback surveys is designed to detect burrowing nocturnal birds, such as storm petrels or Manx shearwaters, by enticing a vocal reply when their natural call is played on a device near, or at the entrance of, their burrow. This method is best conducted during the day, whilst the animals are inactive, and allows to detect burrow occupancy with minimal disturbance. However, not all animals reply and there is a known response rate for each species which needs to be applied to the results obtained<sup>23</sup>.

A playback survey was carried during the 2018 breeding season across the Plémont headland (Image 20), as it was all the land in the study area which could be found to contain visible burrows and which was accessible by foot. The headland was divided into working sections and recordings of both storm petrels and Manx shearwaters were played for a few minutes each, at every entrance of burrow or hole in the ground.

Image 20. Area covered on the playback survey.



## Bird Ringing

Scientific bird ringing has been carried out at the Plémont headland since 1999. Between one and three night-time ringing sessions, usually carried out in July, aim at trapping and ringing nocturnal seabirds such as storm petrels and Manx shearwaters. Health, breeding condition and various biometric data is collected as well, and all records are entered in the database of the Channel Islands Ringing Scheme. Any birds that already have a ring when caught have their ring code submitted via the official channels in order to receive the information from their original ringing (such as location, date, age etc).

## Automated Acoustic Recording

Acoustic recorders are increasingly being used as a way to monitoring the presence of species which use calls or other acoustic means to communicate and navigate, like bats or nocturnal seabirds<sup>27</sup>. The recorders are set up in chosen locations and programmed to record at a set frequency and time, or time intervals, for days, weeks or whole seasons.

An acoustic recorder was borrowed from the Agile Frog Project of Natural Environment Dep. and set up at the Plémont headland, in an area where Manx shearwaters are heard (Images 21 and 22). The recorder was set at the frequency most used to detect shearwaters and petrels, and programmed to record for a period of 20 minutes every hour between 10pm and 5am for the month of August 2020.

Image 21. Set up of the acoustic recorder.

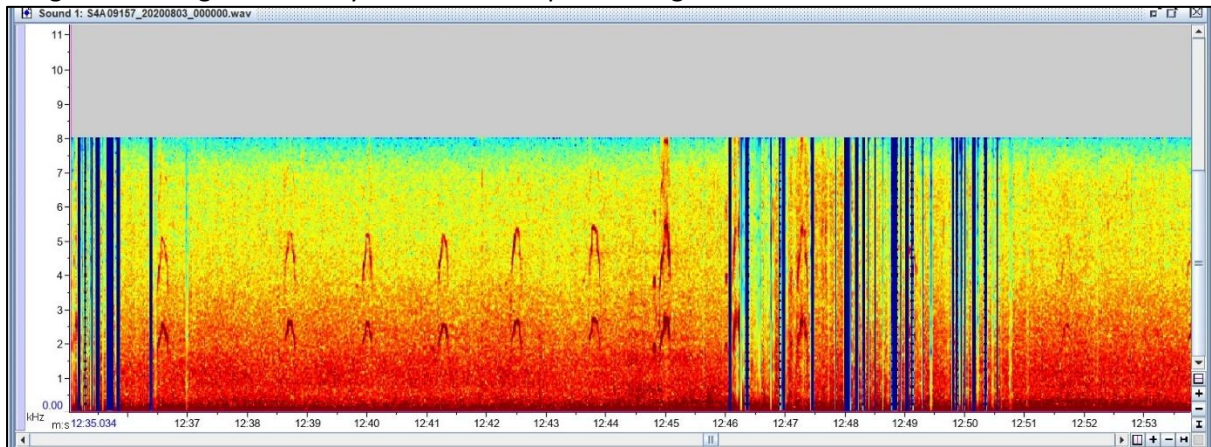


Image 22. Location of the acoustic recorder facing east from the Plémont headland.



The recordings analysed to date did not contain any sounds of Manx shearwater or storm petrel, but did capture the calls of various gulls and oystercatchers, which were easily identifiable (Image 23). The lack of recordings from the species of interest might be due to a wrong positioning of the equipment, or most likely might reflect a real absence of the species in the area, which is expected at this time of the year. Due to logistic constraints, the equipment could not be borrowed until August, which is considered late in the breeding season, and the presence of visiting or prospecting birds is probably significantly reduced.

Image 23. Audiogram of an oystercatcher's repeated single note call obtained from the recorder.



As the recorder succeeded in capturing and recoding the sounds of other seabird species, there is no reason to believe that it would not capture the nocturnal seabirds if they were in the area. This technique has proven to be a cost-effective method to monitor changes in populations of seabirds in the long-term<sup>28</sup>, and therefore it is suggested that a similar method is implemented in 2021 earlier in the season, and tested during nights in which shearwaters can be heard directly by researchers, in order to corroborate the location and set up that is most effective for the recorder.

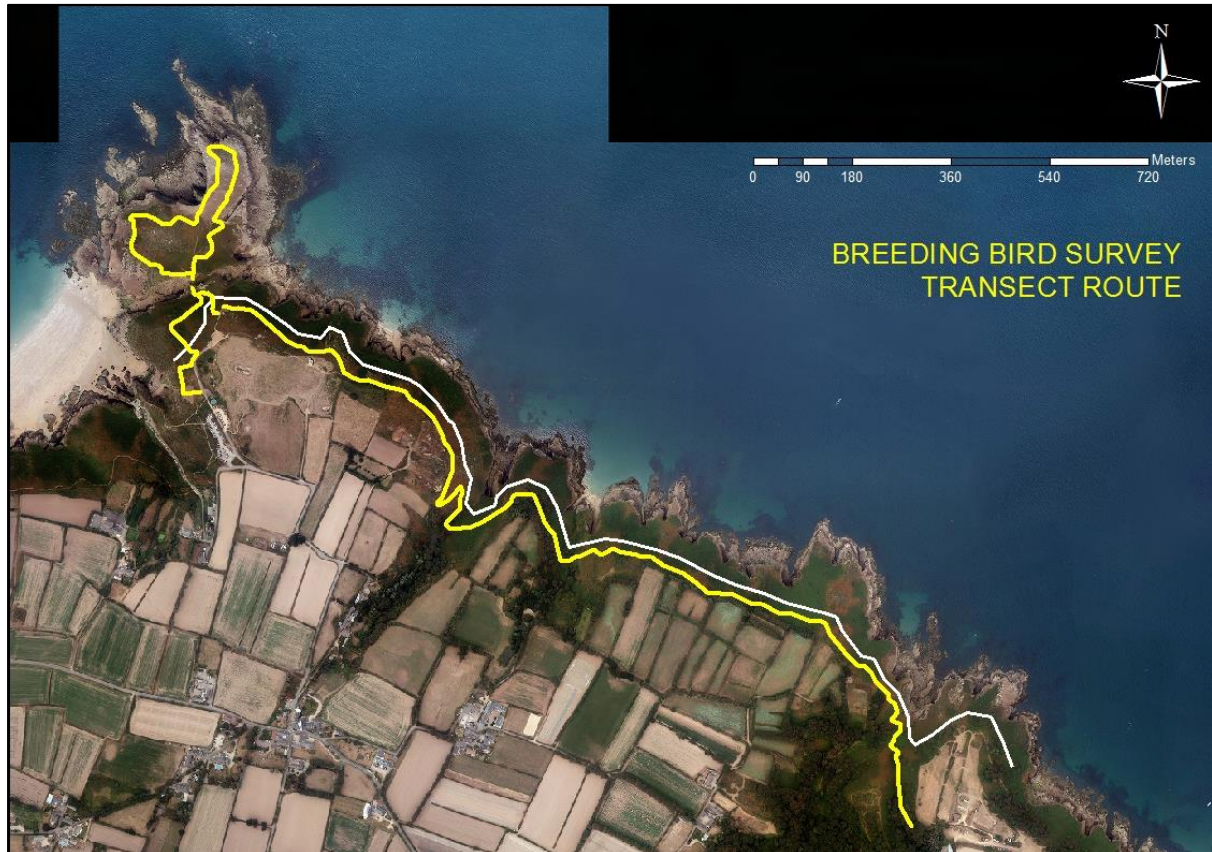
### Breeding bird survey

This method was adapted from the British Trust for Ornithology's Breeding Bird Survey, which is a long-term nation-wide survey undertaken by volunteers, each at a random quadrant of the territory which gets checked twice per year, each year<sup>26</sup>. The two checks correspond to the early and late stages of the breeding season. The survey provides data which allows for monitoring land bird species and their populations, showing trends over long periods of time.

The surveys consist broadly of walking a set route or transect and to record data on all birds observed during the walk: species, sex, activity, and location in reference to the transect. The survey is carried out twice during the breeding season.

A Breeding Bird Survey was devised and trialled in 2020 in order to detect land birds present at the study area and to monitor changes in their populations over time, especially in relation to potential management of habitats and invasive predators. A transect route with a length of 3292m was chosen to be surveyed, and was later broken down in three sections of 825, 382 and 2084m (Image 24). The 'early' survey was carried out in May and the 'late' survey was carried out in June, with all observation and environmental data entered on survey forms and maps.

Image 24. Breeding bird survey transect route (yellow) and proposed fence boundary (white).



### Dusk Surveys with Thermal Imager

The use of thermal imagers is becoming widespread in the field of wildlife research, as they become smaller in size and price. The most accessible format comes in the shape of a hand-held scope which is used to look for animals at night, although it can also detect an animal's heat during the day. Thermal scopes are very useful to find and record animals which cannot be easily detected with the naked eye, such as nocturnal animals or diurnal animals which are very well camouflaged.

A thermal imager was purchased for the project and was used to trial a monitoring programme for rabbits in 2020. Rabbits are seen on most of open ground in the survey area, and are the most likely main prey to feral ferrets, therefore monitoring their population in the long term would provide an insight into ecological changes in the community after the invasive predator eradication. As they are mostly nocturnal, various survey techniques were considered and it was decided that the use of a thermal imager would be the one less likely to cause disturbance in the study area.

Following guidelines from long-term rabbit monitoring programmes<sup>27,28</sup> trial dusk surveys started in June and were carried out each month, except in August, on three consecutive nights of similar weather; or three nights in a period of four consecutive nights, if a change in weather had disrupted the three-night strike. The first few surveys were carried out at sunset. In November they were also repeated one hour after sunset, and finding this time more suitable for logistical reasons, from December onwards they were carried out one hour after sunset. The dates for the surveys were chosen based on favourable weather conditions, avoiding nights with rain or high humidity. The survey recorded all rabbits and unknown mammals observed in the Plémont headland and the NTJ land above the headland (including the open grassland and adjacent fields). The observations were

made from three vantage points visited in the same sequence each time and consisted of three consecutive visual sweeps of each area, with the highest count of the three being recorded (Image 25). Of the three counts carried out each month, the maximum per area was used for analysis.

Image 25. Vantage points (yellow dots) and areas covered by the dusk survey.



### Motion-activated Cameras

Six automated motion-activated cameras were used initially in 2018 to monitor the testing of other techniques such as trapping and flavoured baits, but from November 2018 and up until early 2020 four of them were permanently set up to monitor invasive predators across the study area. The cameras were usually programmed to take three pictures and a 20-40 second video at each trigger, and the resting period between triggers was between 10 and 30 minutes, adjusted depending on the wind, which appeared to trigger some cameras when it was very strong.

The cameras collected images of any wildlife which appeared on their field of vision, from mammals of all sizes, birds, reptiles and amphibians, to even small invertebrates such as slugs, centipedes, and moths.

The four cameras which were mainly dedicated for monitoring all year round worked for an average of 100 days per year and produced 24,237 image files, of which 19,804 were photographs and 4,433 were videos. The videos amounted a total of 34 hours of footage. The cameras were triggered 6,805 times, but only 1,427 of them featured one or more vertebrate animals. Whilst this would not be considered a reliable tool to study the population size or movements of invasive predators, it is considered useful for detection and distribution purposes.

### **Live trapping and mark/recapture**

In 2017 a visiting specialist carried out a 7-day trapping survey in the study area, during which local staff including the BOTE project officer were trained on the basics of live-traps as a means of monitoring invasive predators<sup>29</sup>. In May 2018 a trapping programme was trialled with a total of 34 traps set in the study area, and as it produced positive results it was carried on until end of 2020. The traps were suitable for safely catching animals of a size between a small rat and a medium-sized cat, with a mesh size that would allow non-target species and small mammals to step through comfortably. The cages were labelled, and retro-fitted with dark plastic covers to prevent weather exposure. The staff was trained in wildlife handling and the required licence was secured from the Natural Environment Department.

Starting on May 2018 and throughout 2019, one trapping session of five consecutive nights was carried out each month. Bad weather, road access and seasonal changes in vegetation did not allow for all traps to be accessible or suitable for use at all times, so the number of traps being used at any time varied between 10 and 34. In 2020 the trapping effort was reduced by 50% and trapping sessions were carried out bi-monthly.

Every evening traps were baited and set, and each morning traps were checked and closed for the day. Any non-target species which was accidentally captured (only rabbits in this case) were released. Each target species (brown rat, feral ferret, European hedgehog and cat) had a different handling and processing protocol. Rats were photographed, weighed when possible, and released – it was not possible to mark them. Feral ferrets were measured and marked, and in some cases fitted with radio-tracking collars. Hedgehogs were measured, marked and released. Cats would have been taken to an appropriate charity, or marked and release depending on their status - but none was ever caught.

### **Radio-tracking**

Some of the feral ferrets caught during trapping rounds were fitted with radio-collars which could be tracked with a receiver. Ferrets are most active during the night, resting during the day in an underground den or burrow. Radio tracking was carried out during the day, in order to locate the dens or burrows. In some cases, the signal was not strong enough, or hidden by the terrain, and the den could not be located. However, when the conditions were right the ferret was located within meters of error, or even as close as the entrance to the den where it was resting.

The location of the den or area where each ferret was located was marked on a map, notes were taken on tracks or signs in the vicinity, and dens were photographed for reference.

### **Refugia Monitoring**

This type of monitoring scheme is targeted mainly at reptile species, although it can also detect amphibians and small mammals. It involves visual checks of special mats or 'refugias' laid out on the ground which different species are attracted to and use for hiding, resting or sunbathing. A refugia-based monitoring scheme has been successfully implemented across the Island as 'Reptilewatch JE' by the Jersey Amphibian and Reptile Group (JARG), which has been advising BOTE on this matter<sup>30</sup>.

In preparation for the surveys, 20 refugia mats were laid out across the study area in the spring of 2020 (Image 27). Unfortunately, the subsequent checks could not be carried out due to restrictions related to the Covid-19 pandemic. It is believed that this monitoring scheme can produce data adequate for long-term monitoring of reptiles, amphibians and even small mammals, and it is hoped that surveys will start in 2021.

## 4.2 Results

### 4.2.1 Puffins and other seabirds: Past and present populations

#### Puffins

Image 26. Three puffins near the nesting sites in the study area.



© Romano da Costa

The puffin, a small seabird member of the auk family, spends most of its life at sea and only comes to land to breed. Puffins breed on coastal grass slopes facing the sea, where the pair digs a burrow, or uses a rabbit warren, to line it with grass and make an underground nest. Puffins mature at about five years of age, returning to the area where they were born to look for a mate and breed. They arrive to their breeding grounds in March to return to their nest. The female lays one single egg, and the chick or 'puffling' is raised inside the burrow. When the puffling is almost ready to fledge, it will come out of the burrow at night to observe the landscape and orientate itself. This is done to prepare for fledging, puffling will fly off to sea at night and alone, and will not see its parents again after that. Immature puffins take 5 years to reach breeding age, however they visit the colonies where they were raised before they are fully mature, particularly at the end of the breeding season (July), to prospect for breeding partners and nesting opportunities.

Puffins are rare breeders in most of the Channel Islands and are seen very rarely in the winter. Their local stronghold is Burhou, Alderney, which is monitored by the Alderney Wildlife Trust. According to the Trust's latest Ramsar report, the population estimate in Burhou for 2020 was 167 pairs. Although this population has fluctuated around the 200 mark since the year 2000, it suffered a dramatic decline in the 1970s and '80s, when it shrank from 1,028 individuals in 1970 to only 330 in the '90s.

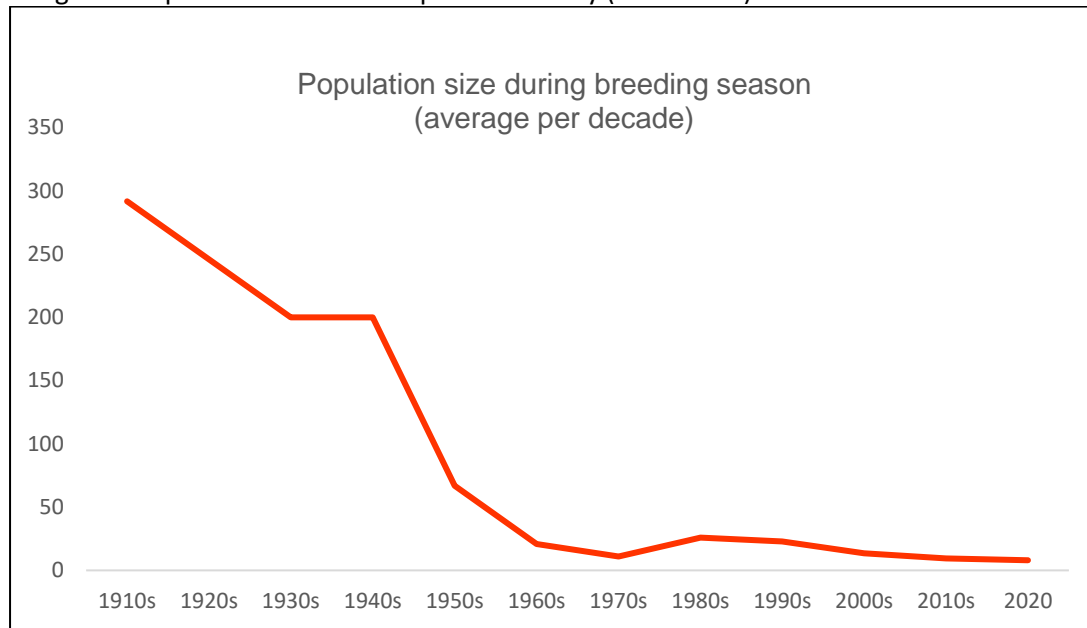
The smaller colonies in Sark and Herm (Guernsey) have both fluctuated between 50 and 10 individuals since the 1970s, and birds seen on the water near the breeding sites suggest a small breeding population in each island, although there is no data available which would confirm numbers of occupied burrows<sup>31</sup>.

The earliest records in Jersey date back from 1911, when 200 individuals were counted during breeding season. The population reached a peak in 1912-1914, with many dozens of pairs each year estimated to add up to around 500 individuals. The population decreased down to 80 pairs (160 birds) in 1950 and 1951, and from the late 1950s to the year 2000 the yearly average decreased



from 50 to 10 birds. Over the last 20 years (2000-2020) the breeding population has fluctuated between 10 and 2 individuals, and it presently stands at 4 confirmed pairs in 2020 (Image 27).

Image 27. Population trend of the puffin in Jersey (1911-2020).



The Jersey puffins appear to have always been between Plémont headland and Grève de Lecq Bay. The only local records not within this area, taken during the breeding season, feature puffins which were travelling past Grosnez point. It is likely that puffins from other islands visit Jersey during their foraging trips and that puffins breeding in Jersey visit waters around other nearby colonies too. The diet of the Jersey puffins appears to be mainly of sand eel, as seen on high-resolution pictures volunteered by local photographers, and puffins and razorbills are often observed diving and fishing on the waters directly below their nesting cliffs.

The recent efforts to monitor the puffins breeding in Jersey, led by the National Trust for Jersey and the BOTE partnership, have produced a reliable account of the size, distribution and productivity of this very small population over the past four years (Table 4).

Table 4. Puffin nest summary 2017-2020 (green: nest visits; blue: food delivered).

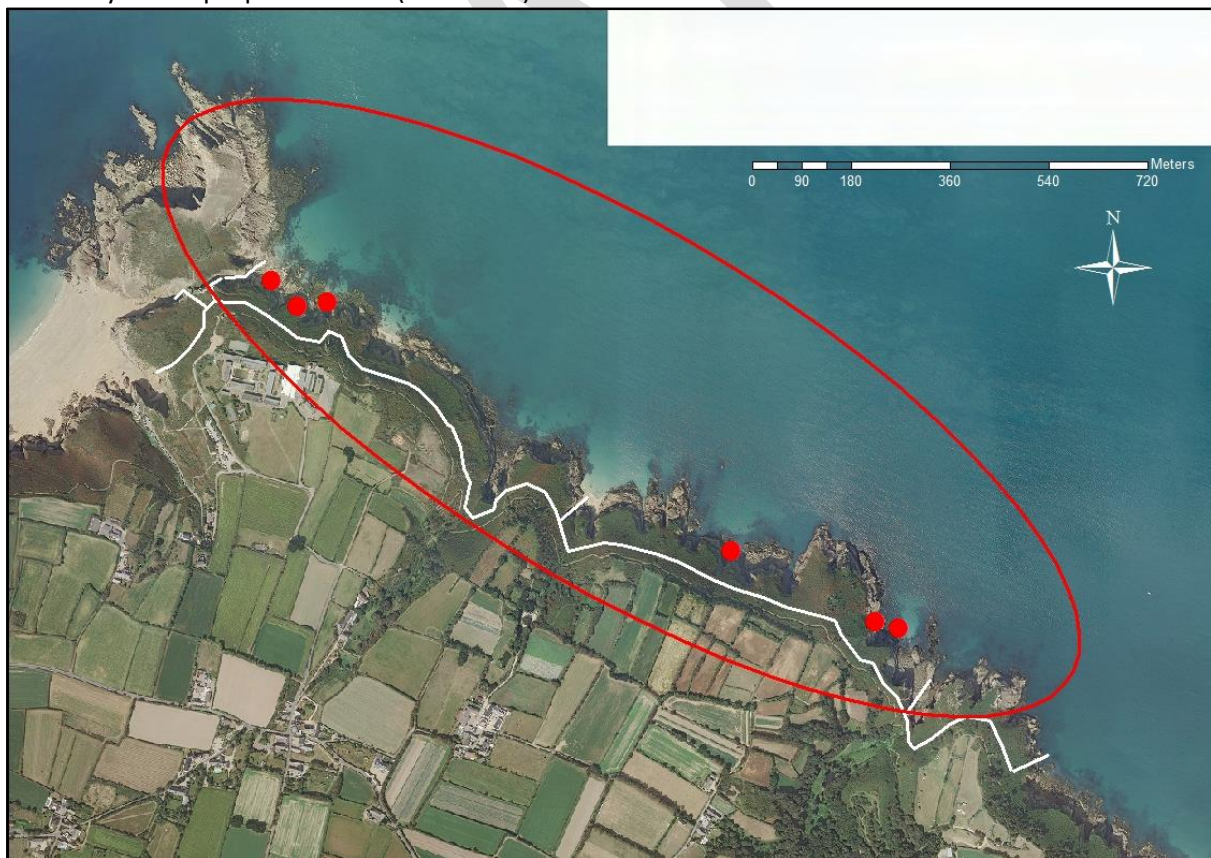
NEST	DESCRIPTION	2017	2018	2019	2020
1	On rock face, entrance below rocky overhang on ledge	✓	✓✓	✓	✓
2	On same rock face as 1, entrance at the end of the ledge behind large boulder	✓✓	✓✓	✓✓	✓✓
3	Burrow on ground amongst rocky outcrops, below vegetation line	✓✓	✓✓	✓✓	✓✓
4	Entrance out of sight, above large boulder below vegetation line		✓✓	✓✓	✓✓
5	Entrance out of sight, above grassy ledge facing east on rocky outcrop				✓
6	Entrance below rocky outcrop, at lower edge of vegetation line				✓✓

On 2017 three nests were located and one more was suspected but not located, with food being brought to at least two of them<sup>32</sup>. As food is only brought to the nest for the purpose of feeding the chick, it is accepted as a sign of the chick having hatched. On 2018 the three nests found in 2017 were active with the fourth nest located. Much time was devoted to watching the nests. Food deliveries were observed for all of them, confirming that four chicks had hatched in their respective nests. In what was a very rare occasion, the puffling of one of the nests was seen one morning stepping outside of the nest, standing for a few minutes at the entrance of the burrow, before going back inside. This was probably the first sighting of a puffling in Jersey on recorded history.

In 2019 the same four nests were active again, and a fifth nest was suspected, but not confirmed. Three out of the four nests had food delivered, confirming that there were at least three chicks being raised. The pair on the fourth nest were seen visiting the nest often all throughout the season, therefore it was highly probable that they that been successful too in raising a chick.

On 2020 the time dedicated to watches was reduced to less than half from 2019, mainly due to the Covid-19 pandemic restrictions. All the four known nests were active again, and two new nests were identified in the vicinity (Image 30). Of these two new nests, one had signs of a chick being reared, with multiple food deliveries by the adults, and the other nest had puffins present throughout the season. Three of the other four nests had fish brought in, with the remaining nest seeing continuous activity by the breeding pair, even though a food delivery was not observed.

Image 28. Location of puffin nests in 2020 (red dots), former breeding range (red outline) and boundary of the proposed fence (white line).



The lack of sightings of fish being taken to two of the six nests could be attributed to the significant decrease in time dedicated to watches, whereas the fact that puffins were seen at the nest throughout the breeding season probably indicates that they had a chick there too – especially in a

nest which has had a breeding pair consistently for the last four years. The conclusion from the 2020 season was that there were six active puffin nests, of which four were very likely to have produced pufflings, and with the other two nests quite likely to have had pufflings too.

Image 29. A puffin on its way to take several fish to a nest in the study area.



© Wilson de la Haye

The puffin is probably the most endangered seabird in Jersey, and one of the most vulnerable animals in Jersey. The review of its status and realization of the precarious situation of its population, highly vulnerable to extinction, prompted the start of the Plémont Seabird Reserve Project. The puffin is not only this project's flagship species, thanks to its charisma and iconic presence in the cultural heritage of Jersey, but it also acts as an ecological 'umbrella' species, meaning that by protecting it and restoring its habitats, many other native species of birds, reptiles, mammals and invertebrates will also be protected as a result.

The precarity of the puffin population in Jersey cannot be understated, and to reflect this it has been placed in the Red List of Jersey Birds of Conservation Concern<sup>33</sup>. A population as small as four to six breeding pairs is highly vulnerable to extinction in a very short time – as short as a single year, if a particularly negative random event were to cause all individuals to perish, such as very bad storm or a disease, for example.

On a hopeful note, the fact that there appeared to be two new pairs in 2020, one of them successful on their attempt to breed, and the second one prospecting a nest at the very least, if not successful too, might be interpreted as an increase in the local population, either by recruits from other colonies, or by individuals born in Jersey reaching maturity and coming back to their natal cliffs to establish a nest. As social breeders, prospecting puffins are more attracted to sites which already have puffins breeding on, and the larger the existing colony, the bigger the attraction to new recruits. It is to be hoped that this small increase in numbers in 2020 might increase the interest to potential pairs which might have visited the area at the end of the breeding season.

All the existing nests are, however, in sub-optimal locations, namely on rocky slopes and cliffs, below the vegetation line. Some seem to be hollow spaces between rocks and ledges, and others appear to be excavated holes between large boulders and rocky outcrops. This distance from the slopes above,

where puffins would usually make their burrows, might be due to the presence of invasive predators on these slopes, and to the widespread and dense bracken cover.

Image 30. A puffin stands outside its nest (bottom left quarter) below fulmar nests in a rocky vertical cliff in the study area.



This study concluded that if the coastal slopes above the sea cliffs were made safe from predators and their coastal grasslands restored, there would be a very high possibility of prospecting puffin pairs exploiting this suitable habitat, which would lead to the population expanding its breeding grounds and increasing in numbers.

### **Razorbills**

Image 31. A small raft of razorbills below their breeding site in the study area.

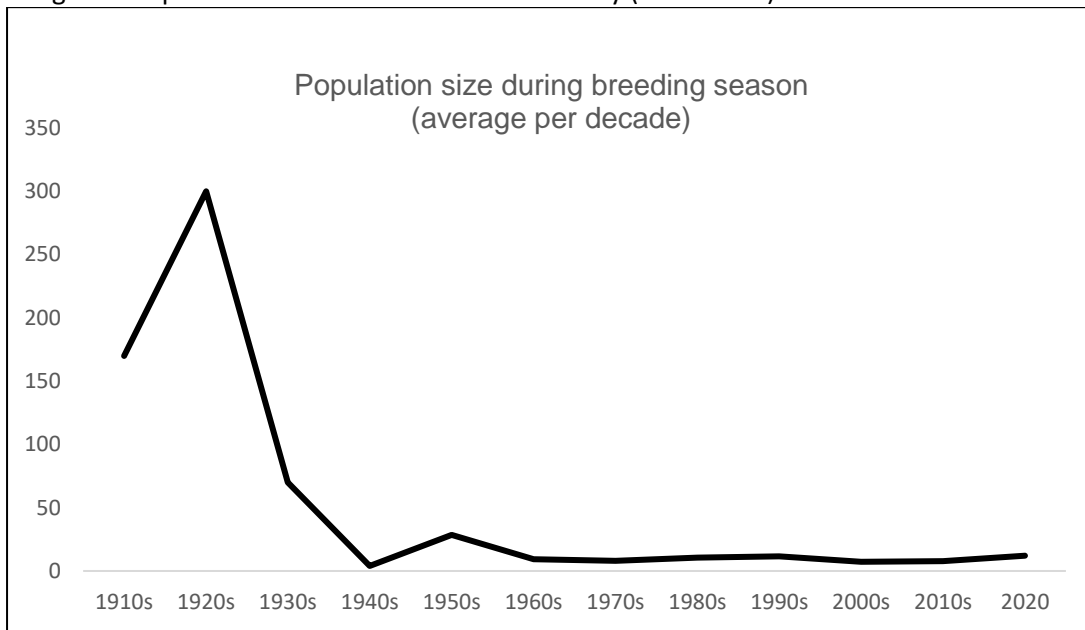


Like the puffin, the razorbill belongs to the auk family, spending most of its life at sea and only coming to land in order to breed, where it nests in ledges and rock crevices in cliff faces. The only two breeding locations for razorbills in the Channel Islands are Jersey and Alderney's islets, although they might have bred occasionally in Sark - and potentially as recently as 2017, when a group of approximately 40 individuals was seen in the vicinity. In Herm there might have also breed

occasionally, with nine individuals on the last count in 2015. In Alderney there were 14 individuals in 1970 and again in 2015, although numbers did fluctuate up to 80 during that period<sup>31</sup>.

In Jersey the razorbill is considered a rare breeding species. The earliest records, from 1911, estimate a breeding population of 100 individuals, with a maximum estimate of 300 individuals in 1924. After that the species started a steep decline with numbers fluctuating between 20 and 50 individuals for most years, until 1990. Since then the population has numbered less than 20 birds at the breeding site, with a present estimate of eight to ten breeding pairs (Image 32).

Image 32. Population trend of the razorbill in Jersey (1911-2020).



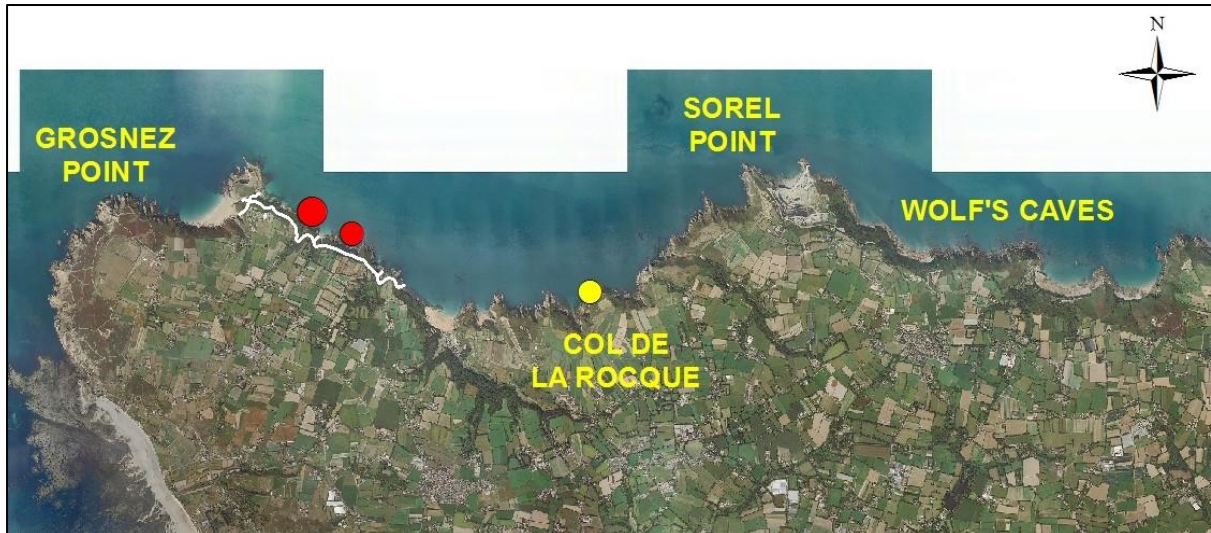
The vast majority of razorbill nests known in Jersey have been and still are within the study area, and would also fall inside the boundaries of the proposed seabird reserve (Image 33).

Image 33. Location of razorbill nests in 2020 (red dots) and boundary of the proposed fence (white).



Records point to a small colony of five to ten pairs near Col de la Rocque between 1934 and 1937. Individuals have also been observed during the breeding season at Grosnez, Sorel point and Wolves' Caves, although breeding has never been confirmed at these sites (Image 34).

Image 34. Location of razorbill nests in 2020 (red dots), previous known breeding site in the 1930s (yellow dot) and other sites with sporadic presence during the breeding season.



Like the puffins, the present population of razorbills is too small to be considered safe from extinction, and is also in the Jersey Red List<sup>33</sup>.

Records have shown that Jersey held the largest colony of razorbills in the Channel Islands, and that the cliffs of the north coast were a suitable nesting habitat. The large numbers seen less than a century ago, in the spot where the small breeding colony still remains, point to the ability of this species to utilise the cliffs for nesting and raising their young (Images 37 and 38).

Images 35 and 36. Razorbills standing outside their nests and a razorbill carrying fish to one of the nests in the study area (Image 36 © Romano da Costa).



It is believed that the razorbill is well placed to increase in numbers and expand across its former breeding range if the right conditions were met; these would include the removal and long-term exclusion of predators from the area, as well as a degree of management of the disturbance at sea from humans and watercraft.

## Guillemots

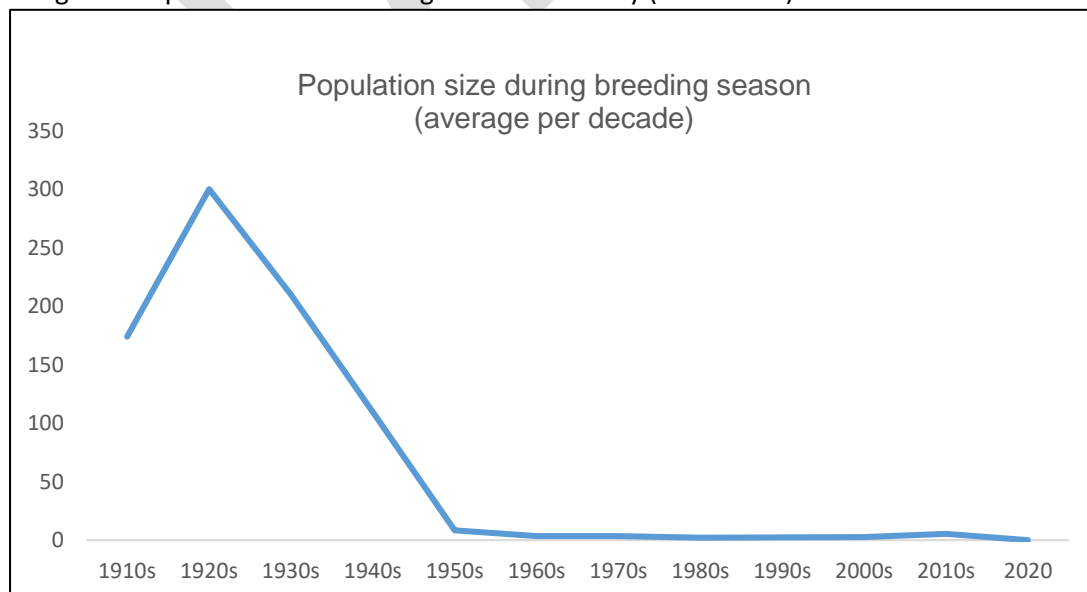
Image 37. A guillemot nearby the razorbill and puffin nesting area.



Another member of the auk family, the common guillemot breeds on ledges and crevices between rocks on steep cliff faces, sometimes sharing the sites with razorbills. Guillemots are a common visitor across the Channel Islands, but presently only breed at two main locations: Sark and Alderney. In Alderney breeding pairs use offshore islets such as Cocque Lihou, with numbers fluctuating between 40 individuals in 1970 to 60 individuals in the 2015 census. Sark holds the largest colony in the Channel Islands, with over 200 individuals counted in 2013 - however recent records might indicate a slow decline, down to an estimated 60-70 individuals in 2020. In Herm numbers have fluctuated between 24 and 90 birds since the 1970 survey, with the last count in 2015 totalling 30 individuals<sup>31</sup>.

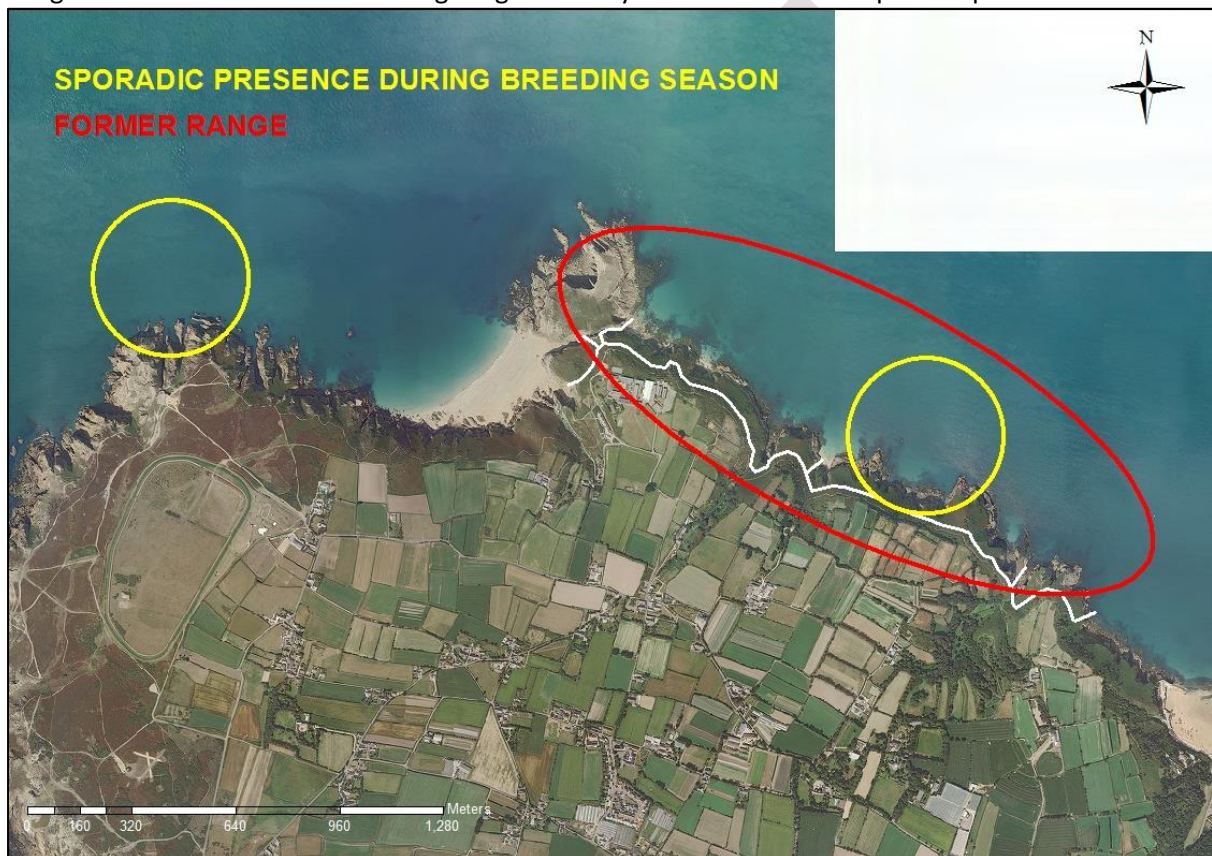
The guillemot is considered extinct in Jersey, but in 1911 it was home to a colony of 200 individuals, which increased to an estimated 300 by 1924. Sadly, the following available record is from ten years later, 1934, and by then the population had dropped to 50 individuals. What followed was a steady decline, with numbers in single figures from the 1950s onwards, and with the last sighting of a guillemot sitting on a nest ledge in Jersey in 1974 (Image 40).

Image 38. Population trend of the guillemot in Jersey (1911-2020).



Like razorbills, all the historical colonies and clusters of guillemot nests in Jersey were found between Plémont and Douet Du Mer, therefore would have fallen within the study area and also inside the boundaries of the proposed reserve. A recent breeding site might have appeared in the last two decades to the west of the study area in Grosnez – that is unless it had been undetected until then. Since the mid 1990s between 1-7 birds have been seen below Grosnez point during the breeding season almost every year, and sometimes observed flying into the cliffs below the castle and lighthouse. Despite the lack of more substantial evidence, it is possible that one or more pairs of guillemots breed in that area unnoticed. Birds are also seen below the cliffs of their former colonies in the study area, where razorbills and puffins still breed, however they have not been observed flying into those cliffs in recent times (Image 41).

Image 39. Guillemot former breeding range in Jersey and locations with sporadic presence in 2020.



The guillemot is a species that could return to its former breeding sites in the proposed reserve. All the management that might be needed is a removal and long-term exclusion of predators from the area, and management of the disturbance at sea from humans and watercraft. The occasional visits to the area by guillemots that approach the cliffs whilst swimming, and the nearby colonies in Sark and Alderney, offer hope that the species can have a future in Jersey.



## European storm petrels

Image 40. A storm petrel released after being ringed.



This small petrel, belonging to the Procelariidae family (albatrosses, petrels and shearwaters) is a nocturnal burrowing bird which nests in little holes under grass tufts or crevices between rocks and boulders.

There are no records of storm petrels breeding in the study area, but the species have bred in Jersey in the past, namely in the east coast and the offshore reefs. The first record is from 1951 and it appears that between 1951 and 1961 the species bred regularly, albeit in very small numbers (1 or 2 nests) at Icho Tower, at La Rousse (near Icho Tower), and at Maitre Ile (Les Minquiers). During the '70s and '80s individuals were observed or trapped (for ringing) occasionally at these locations during breeding season (May to August), however there was not enough evidence to confirm breeding.

Storm petrels have also been heard and trapped by bird ringers at the Plémont headland, during nocturnal survey sessions in July, with a total of 206 birds caught since 1999 (Image 40). Some of these birds might be young individuals from nearby populations in France or other Channel Islands, which might be prospecting for a breeding site to nest.

Ringing records support this statement, with some of the birds trapped at Plémont having been ringed originally at French or British colonies. Birds ringed at Plémont have also been re-caught in Burhou, between 1 and 9 years after their original ringing. The population of Burhou has seen a fast and dramatic increase in the past 20 years, growing from an estimate 60 pairs in 2000 to over 1,000 pairs in 2015. In regards to the other Channel Islands, the species is known to have bred in the north of Herm (Guernsey) up until 1946, and is also considered a former breeder in Sark.

Despite storm petrels being heard and trapped at the Plémont headland, it is believed that the site does not hold any breeding pairs, due to the presence of rats, ferrets and hedgehogs. The playback survey carried out throughout the headland did not detect any burrowing birds either. The fact that the area is visited regularly during the breeding season gives hope that if the site was made secure from predators, the area would have the nesting requirements for the species, and that there would be a surplus of young individuals from nearby colonies ready to colonize the area.

Image 41. Two storm petrels in suitable breeding habitat.



The apparent and rapid increase in the Burhou population over the past 20 years might indicate that this species is very skilled in utilising a suitable area, and that the chances of it breeding in the Reserve soon after it is made safe from predators are probably very high. Conservation projects aimed at helping storm petrels have succeeded with techniques such as rat eradication and by increasing nesting opportunities, for example by creating dry-stone walls<sup>34</sup> and placing ice-cream tubs inside that the birds can use as a burrow.

#### **Manx shearwaters**

Image 42. A Manx shearwater released after being caught and ringed.



This is another member of the albatross family, a small shearwater with a dark back and pale belly, and pointy wings adapted for flying close to the surface of the sea. They are also nocturnal and nest mainly on grassland sea-cliffs, where they use rabbit warrens or dig their own burrows (Image 43). If natural cavities are in short supply, they can use artificial burrows connected to the exterior by a short pipe.

Image 43. Manx shearwater breeding habitat in Lundy Island.



The UK has an important role in the conservation of this species, as it is the home of 80% of the world population. It is also the longest living bird in Britain, with the oldest shearwater known, from ringing research, being ten days short of 51 years old.

In the Channel Islands there are scattered records of breeding pairs of Manx shearwater. Records from nation-wide surveys indicate that there were 55 breeding pairs around 1985-88, and 10 in the 1998-2002 survey<sup>31</sup>. The Guernsey islet of Jethou has a traditional site in which possible burrows were identified in 2015. In Sark there is also evidence of breeding, as in 1977 five shearwaters were killed by a ferret, suggesting they were on their breeding grounds<sup>35</sup>. In 2015 a shearwater was found on the ground of a tunnel in April, and in 2016 and 2017 birds were seen offshore close to the island during the breeding season. It is possible that Manx shearwaters also breed in Burhou in small numbers, as some were heard in July 2015 during a nocturnal ringing expedition to the island.

There are no breeding records of Manx shearwaters in Jersey, but there have been sightings in Jersey waters during the breeding season since at least the 1950s. The playback survey in the Plémont headland did not produce any positive results, and the audio recorder set in the headland did not record any individuals in the vicinity; although having only been set up at the end of the breeding season, the chances of birds in the vicinity were very low to start with.

It is very likely that Manx shearwaters do not breed in the study area, mainly due to the presence of invasive predators, however 17 individual birds have been caught during the night-time ringing sessions at the headland since the year 2000. Birds have also been heard calling on these ringing sessions even before a tape was played, which is typically used to entice them to land near the nets. It is possible that these birds were immature individuals searching for suitable land to nest, which is an encouraging sign if future conservation works can remove and exclude invasive predators from the area.

## Other Seabirds

The seabird census surveyed all the cliffs between Plémont point and Douet Du Mer cove, carrying out visual checks of nesting pairs early in the season, and of chicks in nests later, to determine productivity. Early checks found nests of fulmar, herring gull, lesser black-backed gull, greater black-backed gull and oystercatcher. Productivity checks found chicks on nests or on land of fulmar, herring gull, lesser black-backed gull, oystercatcher, and European shag (Table 5).

Table 5. Results of the seabird census and productivity survey in the study area in 2020.

	Fulmar	Herring gull	Lesser black-backed gull	Greater black-backed gull	Oyster-catcher	European shag
April: nests	46	15	1	2	2	0
June: nests	77	15	4	0	1	1
July: chicks	10	5	5	0	0	2
<b>Productivity</b>	<b>0.2</b>	<b>0.3</b>	<b>5</b>	<b>0</b>	<b>0</b>	<b>2</b>

The species productivity has been calculated as number of chicks per nests counted in April. It is likely that some of the results are under-estimations, in particular the fulmars, as one of the main breeding areas was visited too soon for chicks to be visible. Only repeated surveys over the long-term will show if the 2020 productivity is close to the annual average for these sub-populations, nevertheless it is likely that methodology needs adjusting, in particular the schedule of visits, in order to obtain more accurate estimates for each species.

Image 44. A lesser black-backed gull with its recently fledged chick.



#### 4.2.2 Land birds

The main source of data on land birds at the study area was the Breeding Bird Survey, which was adapted from the BTO protocols for this purpose. The early survey in May produced 143 birds of 17 species, and the late survey produced 220 individual birds of 23 species. The most abundant species during the early survey was swift, followed by goldfinch, linnet and wren, and on the second survey the most abundant species was wren, followed by swift, jackdaw and linnet (Table 6).

Some of the birds observed were believed to be nesting inside the proposed area of the reserve, particularly cliff-nesters and those strongly associated with coastal habitats, such as swifts, kestrels, jackdaws and rock pipits. Based on their movements and behaviour, it is believed that all the other species observed during the surveys could also be holding territories within or overlapping the reserve area.

Some of the motion-activated cameras in the study area also captured birds on occasion. All were species expected to be in the vicinity of the cameras. The majority of the records belonged to pheasants; in smaller proportion were wrens, robins, blackbirds, magpies, song thrushes, stonechats, linnets, goldfinch, meadow pipits, kestrels, and marsh harriers (Images 45 and 46). With the present location and set up of cameras, this was not considered a reliable method for the long-term monitoring of land birds, but it could be for certain species at certain times.

Table 6. Land birds recorded during the breeding bird survey in 2020.

Species	May	June
Common stonechat	0	2
Meadow Pipit	2	0
Rock Pipit	3	6
Linnet	25	18
Chaffinch	0	9
Goldfinch	30	1
Barn Swallow	6	11
Chiffchaff	0	10
Dartford warbler	1	1
Great tit	1	0
Wren	10	42
Dunnock	5	18
Robin	0	12
Blackbird	1	7
Common kestrel	2	1
Common buzzard	0	1
Magpie	1	4
Western jackdaw	0	23
Carrion crow	8	5
Common raven	0	6
Stock dove	0	1
Rock Dove	1	11
Wood pigeon	3	3
Red billed chough	1	2
Swift	43	26
<b>TOTAL</b>	<b>143</b>	<b>220</b>

Images 45 and 46. A wren and a pheasant on a motion-triggered camera.



#### 4.2.3 Native reptiles and amphibians

There are four species of reptile in Jersey (slow worm, wall lizard, green lizard and grass snake), and three amphibians (common toad, agile frog and palmate newt). Thanks to research carried out by other local projects, the distribution of some of these is well documented, in particular species with a limited range, or species which benefit from targeted conservation management. This is the case of the wall lizard, the grass snake or the agile frog. Potentially, there could be suitable habitats for these three species within the study area, however none of them are known to be present, and were not expected to be found during the survey work.

Green lizards, slow worms, European toads and palmate newts were recorded in the survey area during various types of fieldwork (Image 47). Toads and palmate newts were found breeding at the ponds in the Trust land with hundreds of tadpoles developing in them. A large toad also appeared on camera on one occasion, in a location near two of the ponds.

Slow worms have been directly observed near or on footpaths on the western portion of the study area, including on the open ground near the bunker in the Trust land. Two slow worms have been found freshly dead also in this area, their bodies showing bites and puncture marks, which might indicate a casual attack from a domestic animal such as cat or dog. A 2012 local slow worm survey also detected this species on the eastern quadrant of the study area.

Green lizards are regularly seen near or on the coastal footpath as well as across the Plémont headland. They have also appeared on camera on three occasions at a location near the 'Puffin Village', on the public land below the shooting range.

Image 47. Records and recent observations of reptiles and amphibians in the study area.



Although the Refugia Monitoring Programme could not start in 2020, a total of 20 mats were laid out and much was learnt from the process of surveying suitable spots for them. It is believed that this type of monitoring scheme, already implemented in other sites across the Island, can produce data adequate for long-term monitoring of reptiles, amphibians and even small mammals. It is hoped that this monitoring scheme can start in 2021.

#### 4.2.4 Non-target mammals

Jersey has 11 species of terrestrial mammals, four of which are considered invasive and which are potential predators of native wildlife; those are the target species of this proposed eradication and exclusion. The other eight species are non-target mammals: red squirrel, Jersey bank vole, lesser white-toothed shrew, Millet's shrew, rabbit, wood mouse, house mouse, and mole.

The motion triggered cameras set across the study area captured images of all four invasive predators as well as at least six and a maximum eight of the non-target mammals (Table 7). This uncertainty is due to the difficulty in discerning between the two shrew species and the two mouse species from the images.

The main taxa on camera was mammals, and rabbit was the most abundant species of mammals to appear on camera (Images 48 and 49). Some of the rabbits on camera showed physical signs of the Myxomatosis disease.

All the other mammal species combined added up as many records as birds, with nearly half of the birds on camera being pheasants (Table 8). Dogs have been included in the list, as they appeared slightly more than cats. All dogs were thought to have been pets being walked at the time of the image, whereas it was not possible to ascertain whether cats were feral or pets from a nearby home.

Table 7. Terrestrial mammals in Jersey.

Species	Non-target	On camera
Red squirrel	✓	✓
Jersey bank vole	✓	✓
Shrew spp.	✓	✓
Hedgehog		✓
Rabbit	✓	✓
Brown rat		✓
Feral ferret		✓
Mole	✓	
Mouse spp.	✓	✓
Cat		✓

Images 48 and 49. Rabbits on a motion-triggered camera.

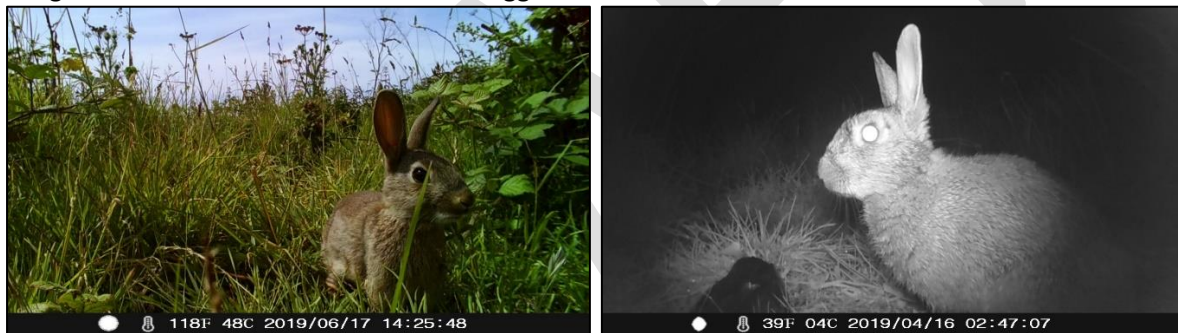


Table 8. Appearances on cameras 1-6 between 2018 and 2020.

TAXA		MAMMALS BREAKDOWN	
<b>Mammals</b>	1273	Rabbit	892
<b>Birds</b>	340	Mouse spp.	141
<i>(of which pheasants)</i>	155	Jersey bank vole	121
<b>Reptiles</b>	3	Shrew spp.	28
<b>Amphibians</b>	1	Brown rat	12
		Feral ferret	12
		Unidentified small mammal	12
		European hedgehog	6
		Dog	6
		Cat	4
		Squirrel	1



The distribution of species at the camera sites was not even nor equal (Table 9). Whilst mice and shrews appeared on most cameras, the larger species showed more clear differences in preferences for certain areas, and were absent from others, at least on camera. This could be due, in part, to the exact positioning of each camera, which was not conceived with one particular species in mind, and which might have been more suitable for one species at one site and for another species at another site. However, the differences between presence and abundance of invasive predators at least seemed partially supported by the live-trapping results (see section 4.5 Target species: invasive mammalian predators). The cameras did not capture moles at any of the sites, despite visual evidence of their presence, in particular between camera sites 1 and 2 (Image 51).

Image 50. Jersey bank vole on a motion-triggered camera.



Image 51. Evidence of a mole tunnel near the NTJ pond at Plémont.



Table 9. Distribution of species at the camera sites.

Species	Camera site					
	1	2	3 (trial)	4 (trial)	5	6
Rabbit	✓	✓	✓	✓		✓
Mouse spp.	✓	✓	✓	✓	✓	✓
Shrew spp.	✓				✓	✓
Jersey bank vole	✓	✓		✓	✓	✓
Squirrel				✓		

Image 52. Location and breakdown of the motion-triggered cameras for bait trials and monitoring.



The dusk surveys produced a steady income of data (Table 10). Most of the surveys were completed in three nights, but some had to be repeated a fourth night due to bad weather or disturbance from humans or dogs. Rabbits were usually easy to identify (Image 53), but various other mammals and birds were observed too. Some smaller mammals appeared to be rats or mice, although most were recorded as unidentified. The number of rabbits recorded with this method, and the consistency between survey nights and areas surveyed, is significant enough to warrant the continuation of this technique as a means for monitoring this species in the long term.

Table 10. Number of rabbits at dusk surveys.

Month	Plémont headland	Plémont NTJ land
June	17	34
July	13	35
September	35	26
October	2	9
November	31	15
December	38	42

Image 53. Two rabbits on the thermal imager.



### 4.3 Conservation value

Islands are disproportionately important for global biodiversity, their ecosystems holding between 15% and 20% of all plant, reptile and bird species. Jersey's biodiversity is unique to the Island, and holds an interesting combination of species from both continental Europe and British Islands. Its proximity to the continent and past physical connection to it has prevented the evolution of true endemic species, although some unique subspecies have emerged such as the Jersey bank vole.

Jersey has many natural areas which are protected in a variety of ways. According to the revised 2011 Island Plan, "The Minister for Planning and Environment has obligations under the Planning and Building (Jersey) Law 2002 and various international conventions to conserve and enhance the Island's biodiversity and to develop public awareness and involvement in conserving it"<sup>36</sup>. These obligations are, in part, met by the identification, designation and protection of sites of wildlife value. These protected sites include the Island's four Ramsar sites, Environmentally Sensitive Areas (ESA) and Sites of Special Interest (SSI)<sup>37</sup>. Jersey has 28 designated Ecological SSI, one proposed Ecological SSI and 13 Geological SSI. NO part of the study area is within an Ecological or Geological SSI at present. Eight Environmentally Sensitive Areas (ESA) have been defined to represent the main areas of the Island's key habitats<sup>38</sup>. The study area would fall into the "North Coast habitats; especially heathlands, coastal grassland, maritime cliff vegetation and interconnecting habitats".

The study area and proposed reserve site are protected from certain developments by the Island Plan and by being part of the Jersey's Coastal Park. The Government's Natural Environment Department recognises its ecological value with a management regime that aims to enhance its biodiversity, preventing the encroachment of dominant plants such as bracken and bramble, and encouraging a mosaic of coastal grassland, heathland and gorseland, amongst others.

The National Trust for Jersey also regards the area as one of its main conservation priorities. Its Lands team has worked to create a semi-natural habitat on the grounds where the old holiday camp stood, by planting sections with gorse and other native shrubs, creating wildlife ponds, and fencing areas vulnerable to human disturbance. Only two years after the clearing of the site, the Trust has reported that palmate newts and toads have been found in the ponds, whilst around 50 plant species have so far been recorded including wild carrot, toad rush, birds-foot-trefoil and small-flowered catchfly. Due to the origin of the top soil used on the restoration work, maritime duneland plants such as fragrant evening-primrose, sea beet and sea radish have also appeared<sup>39</sup>. Wild birds are also being monitored in Trust land with a bi-monthly transect undertaken by the BOTE project officer, as part of a local farmland birds monitoring scheme<sup>40</sup>.

In total, the study area is the home for all main groups of native mammals in Jersey, at least two out of the four native reptiles, and two out of the three native amphibians. A total of 66 bird species have been recorded on the ground, of which 63 are considered regular breeders in the Island (the total in Jersey is 71 species). Most vertebrates in Jersey are protected by the Conservation of Wildlife (Jersey) Law 2000<sup>41</sup>, and a select group of animal and plant species have individual 'biodiversity action plans'<sup>42</sup> in accordance with Jersey's Biodiversity Strategy<sup>43</sup> (see Appendix 1).

In addition to the ecological value of the habitats and natural communities found across the study area, this site is of particular conservation value as the only place in the Island where puffins and razorbills still breed. As both species breed in colonies and respond to visual and vocal social cues, this area has the highest potential to attract new breeding pairs to the Island.

## 5. THE TARGET SPECIES, IMPACTS AND BENEFITS OF EXCLUSION

### 5.1 The Target Species

The following are the species of invasive predators found within the study area, all of which represent a threat to puffins and other native wildlife (Table 11).

Table 11. Number and source of records of invasive predators in study area 2018-2020.

	<b>Brown rat</b>	<b>European hedgehog</b>	<b>Feral ferret</b>	<b>Domestic cat</b>
Images from monitoring cameras	12	85	12	2
Images from bait trial cameras	28	6	9	4
Trapping	13	200	36	0
Direct observation	0	0	0	1
<b>TOTAL</b>	<b>53</b>	<b>291</b>	<b>57</b>	<b>7</b>
<b>N of individuals (minimum known)</b>	<b>13</b>	<b>32</b>	<b>17</b>	<b>4</b>

#### Brown rats

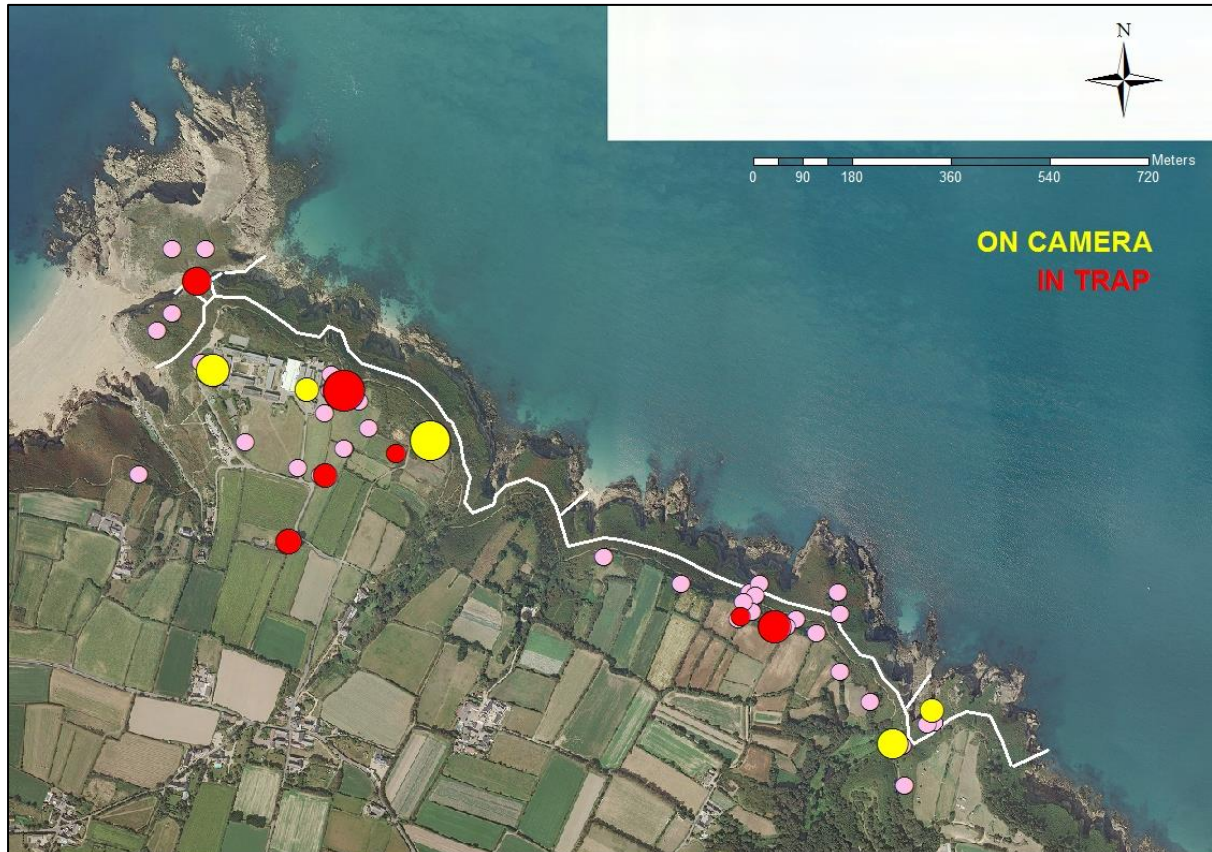
Brown rats were found throughout the study area, with individuals caught on camera, trapped, and observed directly during visits to carry out various different works.

There is a total of 53 records of rats in the study area, with 28 from images during the bait trials, 12 images captured during regular monitoring, and 13 from trappings. Records show a widespread distribution across the study area with one trapping at the entrance to the headland, inside what would be the reserve (Image 55). Rats in the headland have also been often reported by hobbyist anglers that fish there at night, as well as seen directly by the project officer on the roads leading to the Plémont car parks, during dusk surveys.

Image 54. A brown rat on a motion-triggered camera



Image 55. Distribution of brown rats in the study area (yellow and red), location of traps (pink).



Marking the rats was not possible due to logistical limitations, therefore it is not possible to ascertain how many different individuals the records represent. It is possible that the same rat appeared more than once on camera, especially during the bait trials. It is less likely that the same rat got caught more than once in a trap, as they are considered very intelligent animals with great capacity to learn. This would suggest that each of the 13 times a rat was trapped, it was a new rat. This belief is supported by the fact that on some days up to four rats were trapped in different traps, and that rats trapped in consecutive days in the same trap appeared very different from each other, in size as well as behaviour. The methods did not allow for a clear determination of the sex of each rat, but one was known to be female, as it was found in the trap with eight deceased rat fetuses. It is believed that the rat had suffered a natural abortion while it was inside the trap, and appeared otherwise unharmed.

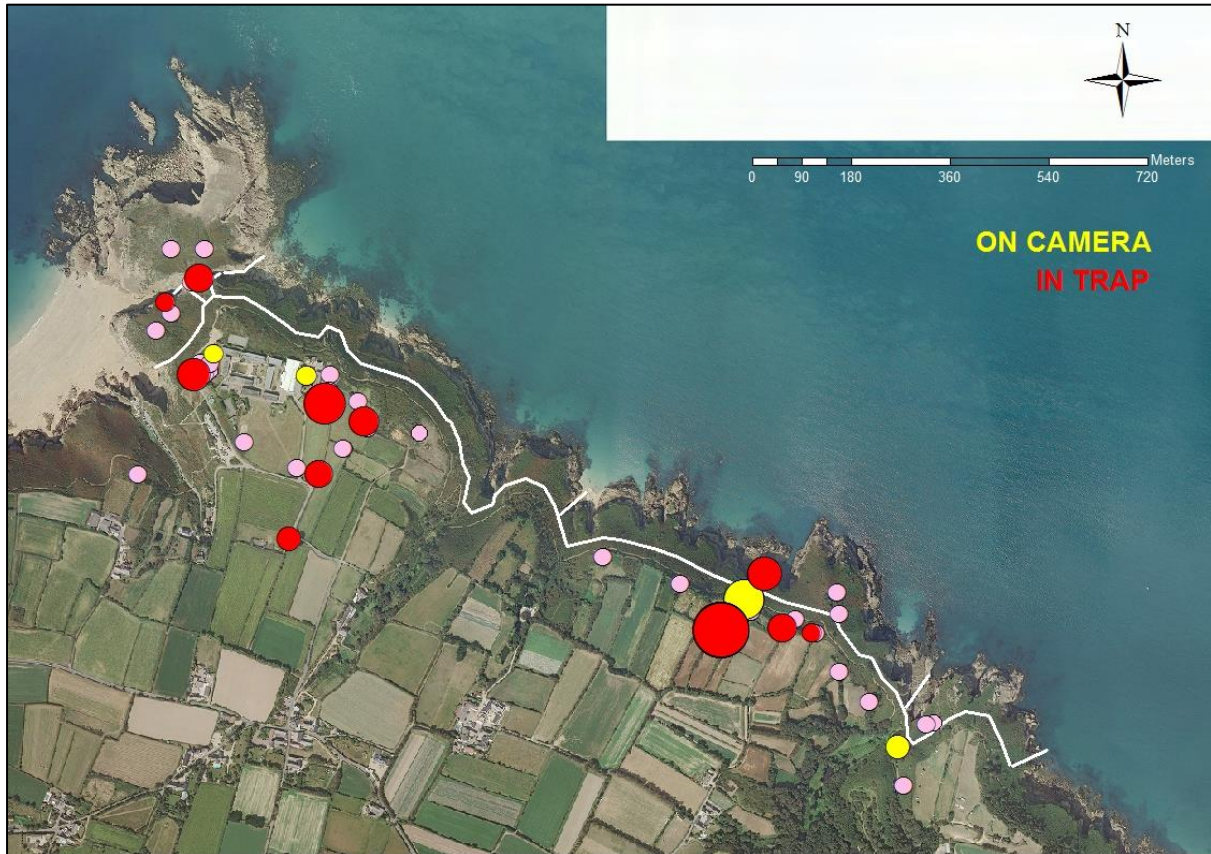
Image 56. A brown rat in a live trap.



## European hedgehogs

The hedgehog was the invasive mammal most recorded in the study area, with a total of 291 records. Of these, 85 were images captured during the bait tests, 6 from regular monitoring and 200 from trapping sessions. Hedgehogs appeared to be widely distributed throughout the study area with some inside the proposed reserve (Image 57).

Image 57. Distribution of European hedgehogs in the study area (yellow and red), location of traps (pink).



Hedgehogs were not marked at the first six of the total 200 trappings; therefore, it is not possible to know if they were six different individuals or less. Unmarked hedgehogs were marked with an individual code at all the trappings afterwards. Thanks to this technique most trapped hedgehogs were identified and their longevity and movements between traps was tracked during the time of the study. Results indicated at least 31 different individuals, most of which were trapped more than once. Adding the first six trappings which were not marked, this would give a population estimate of 32-38 individuals. The hedgehog with the longest running record at the time of writing was H6, marked for the first time in November 2018 and trapped for the last time in September 2020 (Images 58 and 59).

Image 58. Hedgehog H6 being weighed before release.



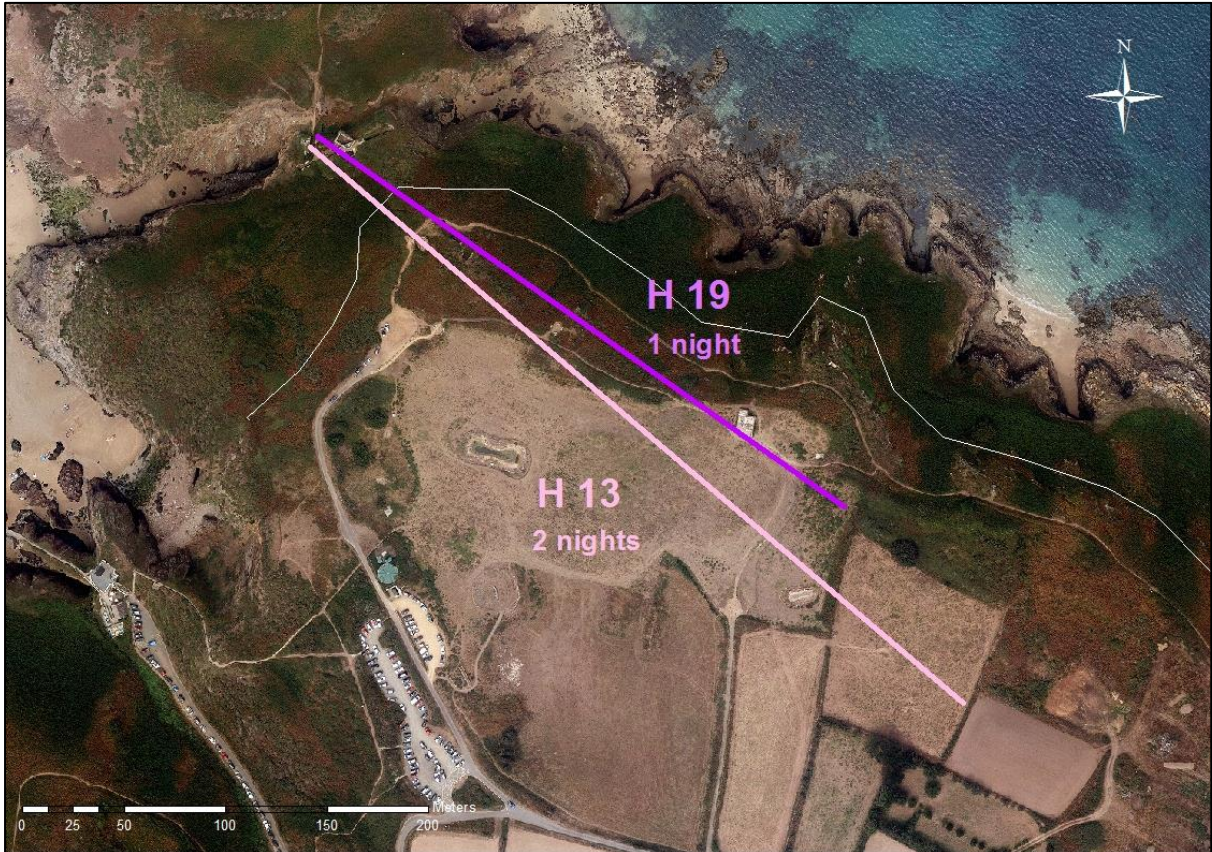
All hedgehogs were examined to determine their sex, which in some cases was possible. Behaviour and injuries were also noted, which mostly consisted of sections of damaged spines consistent with hedge strimmer cuts (Image 59).

Image 59. Hedgehog H6 showing its colour mark (deep purple) as well as a strimmer injury on the centre spikes.



The re-trapping of marked individuals also allowed to determine the movements of the hedgehogs between traps during the 5-day trapping sessions and between sessions. The longest distance covered in one night was 316m by H19 (on a straight line), 102 meters short of the distance covered by H13 in two nights (Image 60). Both of these hedgehogs were found inside the reserve.

Image 60. Minimum distances travelled by hedgehogs H19 and H13 based on start and end location.



Another hedgehog, H17, might not have travelled very far, but it visited almost every trap in a particular area within a period of six months (Image 61).

Image 61. Locations of hedgehog H17 within a period of six months.





## Feral ferrets

A total of 58 records of feral ferrets were produced from research involving bait trials, long-term monitoring cameras and trapping sessions. An even larger body of information was collected by radio-tracking some of the ferrets that were captured in the live traps. The first ferret to be trapped was taken to the JSPCA, as it was caught during the trial period and there was no research protocol in place. All the ferrets trapped afterwards were marked and released (as were found in good health). The ferrets were marked with numbered ear tags or with a radio-collar; or both. A total of 17 different ferrets was found in the study area, including the first one which was taken to the JSPCA. This ferret was a male. The other 16 ferrets trapped during the long-term monitoring were eight females and eight males. Males were often found with bite marks and small wounds around their necks and backs, especially during the mating season, which were consistent with territorial fights between males. Two females were found at different times with swollen nipples and vulvas, as well as bites on the neck – all potential signs of sexual maturity and breeding status. At different times during the research period two of the marked ferrets, upon re-trapping, appeared to be in poor health, and were taken to the JSPCA as per established protocols. None of the ferrets taken to the JSPCA were to be released back into the wild.

Ferrets were mainly trapped in the western half of the study area, where most of the images from motion-triggered cameras also came from. However, two cameras caught ferrets on the eastern edge of the study area and within the proposed seabird reserve. On two of these occasions, a ferret was photographed at the very edge of the first 'Puffin Village', where artificial burrows and decoys have been installed in an area of cleared bracken (Image 62).

Image 62. Distribution of feral ferrets in the study area (yellow and red), location of traps (pink).



Image 63. Two ferrets trapped on the same day before release.



Twelve ferrets were radio-tracked for a period between 1 and 360 days. Radio-tracking revealed, in some cases, large ranges of movement outside the study area (Image 65), with ferrets travelling over 1,000 meters from the edge of the reserve in various directions. A total of 53 dens were identified, with 11 of them found inside the proposed reserve, which was visited by several of the tracked ferrets. In some cases, an individual appeared to be resting in the same den where another ferret had been the night before, an occasionally two ferrets appeared to be in the same den at the same time.

Image 64. Ferret wearing a radio-collar before release.



Image 65. Locations of the dens identified (red dots) and proposed fence boundary (white).



Due to the radio signals being faint or disappearing on many occasions, probably due to the terrain, walls and buildings, it was not possible to locate each ferret every day, therefore it is not possible to know for certain their exact home range. However, when the signal permitted, the movements of some ferrets could be mapped in the space of months, weeks and even consecutive days. The ferret which appears to travel further and faster, according to the locations that could be determined, was ferret F9, a male who seemingly covered long stretches in short spaces of time and who held the largest radio-tracking range overall (1,7400 meters on a straight line) (Image 66). Ferret F13 also travelled far from the reserve (over 1,000 meters from the edge), spending most of her time in a human-built environment, in dens on field banks by roads and stables, and under wood piles in private gardens (Image 69).

Other ferrets, in particular F15 and F16, were often found using a very large burrow in the Plémont headland, in between stays at the NTJ restored land and in the small valley opposite, above the road to Plémont Bay (Image 6770).

Image 66. A selection of locations of ferrets F9 and F13 in relation to the proposed fence boundary.

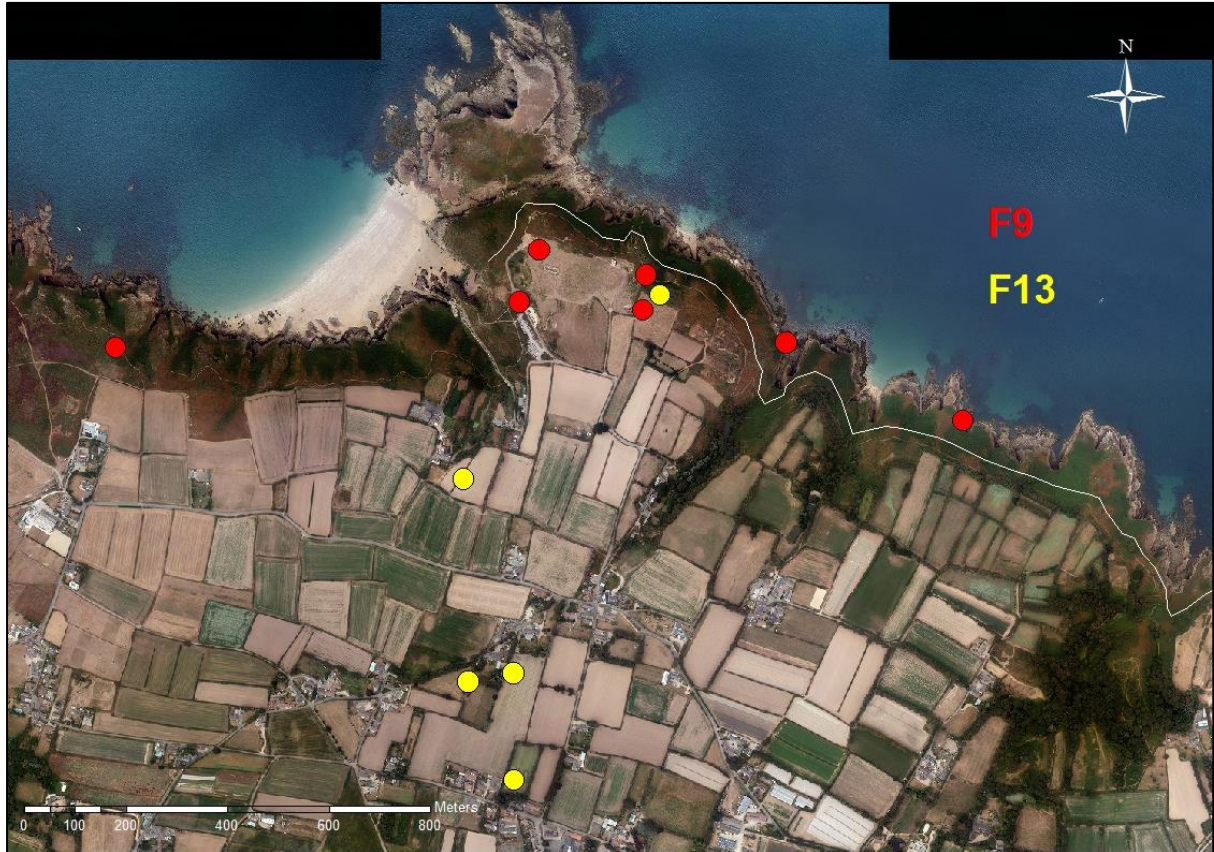


Image 67. Locations of ferrets F15 and F16 including visits to the proposed reserve area.



All dens were photographed and notes were taken on the vegetation, habitat type and tracks present (Images 68 and 69). Some dens appeared to be part of large rabbit warrens, and rabbit droppings were found often at the entrances of the burrows.

Image 68. A large rabbit warren with multiple entrances where various ferrets were located at different times.



Image 69. A den located in a bank on the edge of a horse field in a built-up area.



## Cats

No cats were caught during trapping sessions, although this was expected as the live-traps were not intended for a species of this particular size and habits. However, cats were detected across the study area in sites far from each other, in low densities. Most of the records came from the motion-triggered cameras, and in another instance a cat was directly observed by the project officer whilst working in the area. Two of the images were captured during the bait trials phase, and the other four images were taken during the non-baited long-term monitoring phase (Image 70).

Image 70. Location and identification of cats recorded in the study area (red) and location of motion-triggered cameras (yellow).



A visual check of the markings and coloration of the cats' coats indicated that there were at least four different cats (Image 71). It is not known if the cats caught on camera were feral or pets from a household in the vicinity. However, the cat which was encountered by the project officer seemed quite tame and relaxed, and allowed the officer to pet it (Image 72).

Images 71 and 72. A cat on a motion-triggered camera and a cat encountered by the project officer.



## 5.2 Impacts of the Invasive Mammalian Predators

There is a vast and varied body of evidence documenting the negative impact of invasive species on wildlife communities and their habitats<sup>44, 45</sup>, which are disproportionately larger on islands<sup>46</sup>. This is due to the fact that island species have evolved without the presence of mammals, and therefore have no natural defences against them.

The arrival of invasive mammals to islands happened intentionally or accidentally when humans took them to newly-discovered islands around the globe. These introductions have had devastating consequences, particularly on ground-nesting birds<sup>47</sup>. Globally, invasive predators have caused two thirds of all recorded animal extinctions<sup>48</sup> and approximately 42% of all known bird extinctions<sup>49</sup>.

There are many invasive species present in the UK, the exact number not known with certainty as estimates vary according to definitions used. However, their impacts and the effectiveness of control measures are known and have been documented extensively<sup>50</sup>.

### Brown rat

The brown rat, also known as the Norway rat, is a rodent with brown fur on the back and pale grey fur on its belly. In the UK, male rats have a mean home range of 678m, with that of females being smaller<sup>51</sup>. The adults weigh 150 - 300g, although they may reach up to 500g. In Europe, the brown rat exists primarily in close relationship with humans, but there are also 'wild' populations along water edges. The brown rat is considered to be territorial throughout most of the year, but they will spread when food is scarce, and migrations have been observed. Rats rarely climb trees, but can cross large bodies of open water comfortably such as mudflats, intermediate rocky islets and tidal flows<sup>52</sup>. Females are polyestrous and ovulate spontaneously, with breeding determined by food availability. Litter size is 6 - 11, and females can be sexually active in the season of their birth.

Their diet is omnivorous and they are opportunistic feeders - including raw or cooked meat, vegetable matter, grains and other seeds, berries, roots, a wide variety of vertebrates (fish, shellfish, reptiles, birds and birds' eggs), and invertebrates (beetles, spiders, and flies). Brown rats have also been known to attack and kill young rabbits<sup>53,54</sup>.

Rats are generally considered as the most successful invasive animal in modern times, having spread across the globe alongside humans. There are three rat species who have dominated this expansion: the brown rat, the black or ship rat, and the Pacific rat. Brown rats originated in China and spread across Europe during the 15<sup>th</sup> Century. From there, the species reached the Americas, Australia and Africa, as well as to island groups, via boats - being accidentally transported in cargo or stowing away on vessels. Once an island was colonised, the species could move throughout an archipelago by hitch-hiking boats as well as via natural dispersal, as they are capable of swimming up to 2.5km<sup>56</sup>. The latest data indicates that the brown rat is found in 36% of the world's island groups<sup>5</sup>.

The arrival of rats to islands have caused extinctions and breakdowns of natural communities, disrupting ecosystem functions through predation of animals and plants, causing direct or indirect knock-on effects through interruption of pollination, nutrient pathways, and seed predation, causing forest collapse in the most extreme cases<sup>8</sup>.

Extinction, or severe threat of extinction by rats, has been catalogued from the tropics to sub-Arctic and sub-Antarctic islands<sup>5,9,57</sup>. The brown rat and the black rat are believed to have caused 40-60% of all bird and reptile extinctions on islands, with the brown rat alone responsible for the decline or extinction of the largest number of indigenous vertebrates: 60 species so far.

## European hedgehog

This species is native to western Europe and can be found from Italy and Spain to southern Scandinavia. An adult can vary in weight between 600-700g in the spring to up to 1500g in the autumn, before their winter hibernation. They have few natural predators thanks to the over 7000 spines on their back and crown of the head, protecting them from threats when the animal rolls into a ball. They are nocturnal and solitary, with males often fighting if in close proximity of each other.

The females can have up to two litters of 4-5 young each year, with the female raising the litter alone. Litters can be larger in colder climates. The average lifespan is of 3-4 years although some individuals have been known to live up to ten years. Hedgehogs are omnivorous, preferring a diet of invertebrates such as beetles, caterpillars, earthworms, slugs and earwigs, but they can switch if other prey becomes more accessible, such as frogs, snakes, mice and shrews. The eggs and chicks of ground nesting birds are also an attractive source of food when available<sup>58</sup>.

Humans have introduced hedgehog populations outside of their natural range for a variety of reasons, from controlling slugs and other crop pests (in UK offshore islands) to act as a reminder of the settlers' homeland (in New Zealand in the 1870s).

The extent to which hedgehogs impact upon the New Zealand environment is only beginning to be understood. They are now known to be a major predator on eggs of riverbed breeding birds as well as of a variety of ground-nesting birds, and in some areas are responsible for one in five predator attacks on nests. The species is not only threatening native birds but also many endemic invertebrates, reptiles and amphibians. For these reasons the species is considered a nation-wide pest and is part of many ground-trapping programmes across the country<sup>59</sup>.

In the UK hedgehogs are listed as 'priority species' and 'species of principal importance' on the national Biodiversity Action Plan, as the species has declined over the last 60 years from 30 million to less than one million. However, in the UK's offshore islands they have been an invasive species since their introduction by humans in the 1970s. A single pair brought to North Ronaldsay grew to a population of 1000 hedgehogs in only 12 years. Similar introductions have resulted in hedgehogs being found at in the Orkney Islands, Shetlands and the Isles of Scilly, where they mostly survive at densities higher than in their native habitats. This is probably due to the lack of some of their natural predators, like the European badgers, and the fact that climate change is leading to more favourable conditions for hedgehogs on islands.

This lack of predators and competitors has led to very high levels of predation by hedgehogs on the eggs of ground-nesting birds: on one island hedgehogs were found to cause a 39% decline in breeding shorebirds over ten years, and were responsible for up to 50% of all breeding failures.

In another island (South Uist, Scotland) hedgehogs were introduced to control slugs and snails in gardens. A group of four individuals released in 1974 had grown to a population which averaged 2750 individuals per year by 2006. A study found that hedgehogs predated up to 60% of the nests of some waders, and their density did not appear regulated or diminished by the subsequent wader declines. The study concluded that local extinctions of susceptible wader species were likely if no action was taken to reduce hedgehog predation<sup>60</sup>. The effects of the removal of hedgehogs was tested using fenced enclosures, showing that the breeding success of waders inside the plots (where hedgehog densities were zero or low) was approximately 2.4 times that of birds nesting in adjacent control areas (where hedgehog densities were high)<sup>61</sup>.



## Feral Ferret

The ferret is the domestic form of the polecat, a member of the weasel family (Mustelidae). Depending on the interpretation of its origins, it is referred to as *Mustela putorius* (European polecat), *Mustela putorius furo* (subspecies or descendant of the European polecat), or simply *Mustela furo*. Feral ferrets and European polecats hybridise where their ranges overlap, and the hybrids are often indistinguishable from the wild polecat<sup>62,63</sup>.

Ferrets were first known in Palestine some 1000 years BC, where they were used for hunting rodents and hunting rabbits. They have been known in Europe since at least the Middle Ages<sup>64</sup> and in the 1970s begun to gain popularity as pets in many parts of the world<sup>65</sup>.

Ferrets are sexually dimorphic, with males being much larger than females (between 1000-2000g in weight, 600-900g for females)<sup>66</sup>. They have a long and slender body, which measures 48cm-60cm long including the tail. They have large canine teeth, and each paw has a set of five non-retractable claws<sup>67</sup>.

Ferrets are largely nocturnal, and spend the day resting in crevices, hollow logs, other animal's burrows or dens that they dig themselves. Their home ranges vary in size depending on many environmental conditions, such as terrain and prey density. In New Zealand's high country, home ranges are 100–120 hectares for males and 80–100 hectares for females. In the lowlands, males have home ranges of 30 hectares and females of 12 hectares<sup>68</sup>. In Europe feral ferrets are found in many habitats, such as dune systems which are the home to large rabbit populations. Den sites can be found in gorse, dense scrub, rabbit holes, buildings, rubbish piles, and hay barns<sup>69,70</sup>.

In Europe, rabbits are the main prey of feral ferrets, but they also eat birds, even when rabbits are plentiful<sup>70</sup>. Ferrets are capable of switching to other prey when rabbit numbers decrease<sup>71</sup> and will prey on hares, possums, bird eggs, lizards, hedgehogs, frogs, eels and invertebrates. They are known to scavenge the carcasses of other ferrets, hedgehogs, cats and possums, and even lambs in the Scottish Isles. Some seasonal variation in diet has also been observed: rabbits and hares are preferred in summer, and rodents in autumn and winter. Birds are eaten all year round, but more in spring and summer<sup>71</sup>.

Ferrets have been introduced accidentally in the wild as escapees from ferreters and pet owners, but have also been introduced intentionally to control rabbits. This invasive predator threatens a wide variety of native wildlife in any place where it has been introduced: Europe, North America, Australia and New Zealand amongst others. Native wildlife threatened by ferrets includes ground nesting and flightless birds in New Zealand<sup>72,73</sup>, seabird populations on the Azores<sup>74</sup>, and bird populations in the Scottish Isles, to name a few<sup>75</sup>.

It is believed that ferrets have been in Jersey since at least the 1970s, originally brought by people with the purpose of hunting rabbits. Nowadays ferrets are kept in Jersey as pets, as well as still used for hunting in some parts of the Island. The present feral population is believed to have originated from escaped or released animals, and it is probably self-sustaining or increasing; recruitment happening from animals breeding in the wild, as well as from captive ferrets escaping or going missing in small numbers. Feral ferrets are known to attack poultry and domestic waterfowl, and even pet rabbits on occasion<sup>76</sup>. Local pest controllers are licensed to trap and humanely destroy them if required.

## Cat

The cat is the domesticated descendant of the African wildcat, and thanks to its association with humans it has established itself in almost every environment that people have colonised. Feral cats are usually smaller than their domestic counterparts, averaging 1.5-3.0kg, although can grow up to 5kg. Despite the various breeds generated by human selection, feral cats tend to revert to physical types closer to wild ancestors, such as black, tabby or tortoiseshell.

Cats are intensive breeders. A female cat reaches reproductive maturity between 7 to 12 months of age, and can reproduce any month of the year if food and habitat is adequate. A female may produce three litters per year<sup>77</sup>, with an average litter of four to six kittens<sup>78</sup>.

Feral cats in the wild have adapted to a variety of habitats, such as forests and woodlands, fields, grassland, riparian habitats and even tussock grassland in a sub-Antarctic island<sup>79</sup>. Their mean home range is usually larger for males than females, and dependant on prey availability. When comparing domestic cats to feral cats the differences are significant: an Australian study determined a mean home range of 7 to 28 hectares for domestic cats, whilst it was just short of 250 hectares for feral cats. In New Zealand some home ranges of feral cats were as large as 985 hectares.

Cats arrived to all continents via ships, being carried intentionally to control rat infestations. They were also taken by humans as pets and subsequently escaped or were left behind in islands. The diet of feral cats on islands may be different to that of their mainland counterparts, with many cats often taking advantage of alternative food sources. Besides many species of seabirds, they have been recorded to prey on indigenous invertebrates, hatchlings of green turtle, black rats, flying foxes, and even birds as large as themselves, such as frigate birds, pelicans and flightless cormorants<sup>80,81</sup>.

Cats are considered among the top ten worlds' most invasive species<sup>82</sup>. Their effects on native species are widespread, especially on islands, where they have caused at least 14% of all the bird, mammal and reptile extinctions<sup>83</sup>. Of the 63 species of birds, mammals and reptiles that cats have brought to extinction, 33 were found in islands. In a well-known case, a single cat caused the extinction of a whole species, the endemic Steven's Island wren<sup>83</sup>. On islands native predators are few or non-existent, and potential prey is abundant, so cats can increase to form large and dense populations. The estimated mortality caused by cats in Marion Island, for example, stands at 450,000 seabirds per year<sup>84</sup>, and at Kerguelen Island at 1.2 million seabirds per year<sup>85</sup>.

### 5.3 Benefits of Removal and Exclusion of Mammalian Invasive Predators

Given the well-documented impact of invasive predators on seabirds and other native wildlife<sup>9</sup>, conservation efforts that have focused on the eradication of these threats from islands worldwide have produced significant and positive results. New techniques developed have prompted renewed efforts to control or eradicate introduced mammals<sup>86</sup>, and the application of good ecological principles and practice has become paramount to maximise the success of such intense management<sup>87</sup>.

Conservationists in New Zealand, Australia, America and the Pacific have a great tradition of mammal eradications from islands and have established a vast pool of techniques and expertise. By 2010 New Zealand had seen 147 populations of 13 species of invasive vertebrates removed from a least 95 islands with a total area of 32,000 ha. The benefits to biodiversity from those eradications included improved prospects for 16 species of invertebrates, two species of frogs, three taxa of tuatara, 23 species of lizards, 32 taxa of terrestrial birds and 16 taxa of seabirds. In Western Mexico, the eradication of black rats from 5 islands resulted in the protection of 46 seabird populations<sup>88</sup>. At Midway Atoll National Wildlife Refuge, Bonin petrel populations increased from less than 5000 pairs in the 1980s to over 135,000 pairs in 2008 after the eradications of rats in 1997<sup>89</sup>.

Globally, a review found that 251 eradications of invasive mammals on 181 islands has benefitted 236 native species<sup>90</sup>. A later study identified and ranked a further 292 of the most important islands where eradicating invasive mammals would benefit highly threatened vertebrates. It was determined that eradication of invasives was socio-politically feasible in 169 of these islands, and that it would improve the survival prospects of 9.4% of the Earth's most highly threatened insular vertebrates (111 of 1,184 species)<sup>90</sup>.

Closer to Jersey, various island eradications have been completed successfully in British Islands, with St Agnes and Gugh of the Isles of Scilly being declared rat-free in 2016, and the Shiant Islands in Scotland in 2018. The benefits of eradications have already been documented in other islands such as Ramsay, Wales, where rats were eradicated during the winter of 1999/2000. In Ramsay populations of Manx Shearwater increased from 849 pairs before the eradication to 4,796 pairs in 2016, and Storm Petrels, which had been previously absent, established a small colony which in 2016 had at least 12 pairs. In the cliffs of Lundy Island about 3,500 puffins used to breed in the 1930's, but had been reduced to only 13 puffins by 2001 due to rat predation. Manx Shearwaters had also reached a low 297 individuals. The eradication of rats from Lundy, carried out from 2002 to 2004, allowed both species to recover, and by 2019 there were 375 puffins and 5,504 Manx shearwaters in Lundy.

The benefits of removing and excluding invasive predators from around the Jersey puffin colony are expected to be positive both ecologically and socio-economically. Even though the direct impact of predation is not easily observed at present, historical records point to most of the damage being done in the past decades, with the remnant population of puffins and other seabirds being physically limited to the cliffs most inaccessible to the terrestrial predators. This, paired with their low densities, makes them a difficult prey to access, and therefore of low priority to potential predators.

There is a small possibility that this status quo, of small and stable populations, will be maintained at perpetuity. However, evidence has shown that small populations are disproportionately more vulnerable to random chance events which precipitate extinction. This is reflected in the fact that most species that go extinct do so by suffering a string of unfortunate events, such as fire, storms, heavy rainfalls, extreme temperature changes, diseases, or even an unusual increase in predation,

which exterminate the last individuals of a colony. Large populations, in contrast, usually hold a wider range of individual variations which prevent such random events from killing all individuals at once, allowing for some to survive and re-build the population. Moreover, the Jersey puffin has been placed in the Jersey Red List of birds of conservation concern, which echoes locally the IUCN classification and which recognises a 'small but stable population' (of less than 50 individuals) to be Critically Endangered.

The Jersey puffins would benefit from a chance to increase in numbers in order to make the population more resilient to random events. This situation, paired with the negative effect that the invasive predators are known to have on terrestrial native wildlife, makes a case for removing and excluding all invasive non-native terrestrial mammal predators from the area.

If we establish that the presence of predators in the area is impeding the expansion of the small colony, and the re-colonisation of other extirpated species, to otherwise suitable habitat, we would conclude that the removal and exclusion of these predators will result in the expansion and increase of puffins, as well as other native seabirds, land birds, reptiles, amphibians, mammals and invertebrates.

Key benefits of removing invasive predators include:

- Returning this area in the north coast partly towards its natural state, free of all introduced mammals. The gross result is expected to be positive, especially if all four target species are removed simultaneously.
- Providing safe and suitable breeding habitat for the surviving colonies of puffins and razorbills, giving them room to expand safely.
- Protecting all other seabird species in the area.
- Allowing the recovery of locally extinct species such as the guillemot and the storm petrel.
- Allowing the natural colonization of species that breed in the English Channel such as the Manx shearwater, or that could potentially breed in the Channel such as the Balearic shearwater.
- Allowing the recovery of a wide range of terrestrial vertebrates and invertebrates, particularly those associated with coastal grassland, heathland and gorseland.
- Restoring ecological food chains and webs, allowing native plant communities to recover free of unnatural influences such as grazing from rats.
- Protecting terrestrial bird species in the area, allowing their expansion and increase.
- Protecting many other native vertebrates such as green lizard, bank vole and shrews.
- Management of dominant plant species such as bracken and control of its expansion with grazing flocks.
- Creating a safe biological reservoir where other native species can be translocated into. This could be useful to create populations of species with limited range in the Island, such as the wall lizard, or to translocate and head-start populations which might become endangered at their present ranges. It can also be used, if necessary, to generate a Myxomatosis-free sub-population of rabbits or to study the disease in an enclosed population.

Social and economical benefits:

- A significant step in acquiring local expertise and experience in managing vulnerable native species and their threats, and as a means of furthering local skills and capacity-building.
- Supporting economic activity in the Island with the use of local contractors, producers and suppliers, including a shepherd for the grazing flock.
- Offering training and apprenticeships, via the main contractor’s own trainee and student bursary scheme, and the Government’s Back To Work scheme.
- Providing learning opportunities for students of institutions such as JICAS.
- Creating internships in the field of ecological monitoring and management for local students.
- Promoting collaboration with local businesses such as cafes, shops, birdwatching guides and other leisure operators via our awareness campaigns and interpretation materials.

The hospitality, leisure and tourism sectors will also directly benefit from an increase of biodiversity and public interest in the area.

It is not possible to foresee all potential negative effects of the project effects of the project, however it will be paramount to prepare a list as comprehensive as possible, to assess the potential effects and to prepare mitigation measures. Table 12 summarizes a preliminary list of effects and their potential mitigations.

Table 12. Potential negative effects, assessment and mitigation.

Potential negative effect	Assessment and mitigation
Risk of partial failure with one or more species surviving eradication	Best practice and adequate resourcing will maximize prospects for success, repeated attempts at eradication until completed.
Risk of re-invasion	Biosecurity measures in place with ongoing feedback and adjusting of protocols.
Risk to some non-target species through use of toxins or traps	Scenarios are largely predictable and avoidable based on experience and best practice protocols. Effects will be short-termed and monitored, and most likely off-set by long-term gains.
Potential isolation of small mammal and reptile populations inside the fenced area	Monitoring long-term, with potential management options in place such as translocations in either direction, but likely that populations will be fine as area is large and ecological balance will re-set itself. Possible increase of natural predators of small vertebrates such as kestrels, buzzards, marsh harriers, crows and ravens.
Lack of predation results in an increase of rabbits and subsequent increase of grazing pressure leads to changes in plant community	Some changes in vegetations will be welcome due to the dominant nature of species such as bracken, bramble and gorse. Increase in rabbits and rabbit kits pups will be capped by increase in predation from their natural predators such as buzzards, marsh harriers and ravens. Monitoring long term in place, which will advice if and when reduction of rabbits is needed. These techniques might involve non-lethal and lethal methods, and should be agreed with government and licencing authorities.

## 6. FEASIBILITY ANALYSIS

### 6.1 Technical feasibility

The basic principles to achieving a successful eradication of an invasive mammal are:

- All individuals of the target species must be removed or put at risk by the methods used.
- All target species must be removed at a rate faster than they can reproduce.
- The risk of reinvasion must be zero or must be managed effectively.

The proposed methods to achieve this are, in order of implementation:

Phase 1. Construction of a predator-exclusion fence around the puffin colony.

Phase 2. Removal of any invasive mammalian predators present within the fenced area.

Phase 3. Establishment of a biosecurity strategy to prevent reinvasion (See *Section 6.2 Sustainable*).

Additionally, all monitoring of target and non-target species is to be continued as per the methodologies and recommendations from section 4.1.

#### 6.1.1 Construction of a predator-exclusion fence

Pest-proof and predator-exclusion fences have been used in conservation projects in New Zealand, Australia, Hawaii, the Azores and many other islands, generating a wealth of experience and technical knowledge on their costs, installation and long-term management.

##### Requirements for an effective fence

Fences that have succeeded in excluding invasive mammals have the following specifications in common:

- 1.9 metres high (as cats can jump over 1.7m unassisted).
- A hood or cap (to prevent mammals from climbing over the fence).
- A mesh square no wider than 7 mm (to exclude juvenile rats or mice).
- An underground skirt extending at least 350mm from the base of the fence (to prevent burrowing under the fence).
- To sit on a level platform (to prevent water run-off).
- Four metres clear of vegetation or other structures outside of the fence. No trees should overhang the fence.

The general design of this type of fence is shown in Image 73. This design uses either anodised aluminium or timber posts; aluminium posts can prevent damage by termites or seawater.

Image 73. General pest and predator-proof fence design. © Dep. of Conservation, NZ, Bell 2014.



Whilst it is generally possible to build a fence anywhere, the exact location and design will depend on the target species, terrain, waterways and access requirements. Structurally speaking, the fence must be able to withstand the environmental conditions of Jersey, which will include heavy rainfall and storms. The average rainfall in Jersey is 875mm/year (Jersey Met Climate Statistics, 1981-2010), with the wetter season being from September to March. The prevalent winds are westerly, with an annual average speed of 15.2mph (data from Jersey Airport and Weatherspark statistics, 1980-2016). The Island can occasionally experience violent storms (64-72mph, Beaufort scale) which can be accompanied by gusts of hurricane-strength winds (73mph and over). Structural integrity of the fence has been a particular problem for fences exposed to strong winds and hurricanes, such as the case of the Corvo seabird project in the Azores (T.Pipa *pers. comm.*). It is believed that the weather conditions and the type of exposure are more favourable at the Jersey site.

Within the proposed reserve, there are two small seasonal waterways that led onto the sea from coastal valleys. In order for the fence to be effective, waterways must be sealed to prevent invasive species from obtaining access into the reserve, especially rats who often use waterways to move between areas using the cover these areas provide. At the point of the streams, the fence can be fitted with swinging water gates, so that as water pressure builds, the gates open to allow water through without adversely affecting the fence. These gates can be fitted with alarms to automatically indicate that they have been opened allowing staff to check the fence. Alternatively, culverts with fixed screens can be placed in small waterways through or adjacent to the fence. One of the most important aspects is the continued water management at the site. It is important to assess the issues of water flow at the site to ensure that the excess water does not undermine the integrity of the fence. It may be necessary to test the local flood conditions at the site to determine maximum flow and timing to ensure the fence design has scope to deal with these events. Any breach by water (or vegetation) will allow the reinvasion of invasive species, particularly rodents.

Ideally, the ground would be flat or with a gentle slope, to allow the fence to sit on top of a platform at least four metres wide to provide space for the fence, vehicle access (for monitoring) and drainage (for surface run-off). The strip at either side of the fence would need to be kept clear of vegetation and any other structures.

It might not be possible to have a four-metre platform along the whole fence due to the terrain and slope, but at these difficult spots it will still be possible to install and maintain the fence by other means. Issues arising from the management of stream gates and damage from fallen trees can be pre-emptively mitigated by choosing a route that reduces the potential impact of these issues.

Access into the area protected by a pest-proof fence requires at least one entry gate. This is either a half-gate that people step over while the base portion of the fence remains intact, or a lockable double-gate system that allows vehicle, pedestrian or livestock access. An alarm system can be established if both gates are opened at the same time. The type of gates used at different points of the fence will depend on the requirements of that particular access point.

The integrity of the fence can be compromised by the effects of animals, vegetation, inanimate objects, and accidental or deliberate human action. There will need to be a damage detection system; this can include physical checks by staff, alarms installed which will trigger automatically when there is a breach, and even security cameras if vandalism or theft are a potential problem. Community engagement work will hopefully instil a sense of pride and ownership in the population; this will be of great benefit for the long-term sustainability of the project (see Section 6.3).

### **Location and route options**

The proposed location of the fence would place it below the public footpath and above most of the cliffs and headlands between the shooting range and Plémont headland. The approximate length of the fence would be 2,938 metres and the protected area would total 32.3 hectares (Image 74).

Image 74. Proposed location of the predator-exclusion fence at the Plémont Seabird Reserve.





If the terrain allows it, the reserve would have four sections, to match topographic features as well as to separate land by legal ownership. The sections will also give the advantage to slow down re-invasions to core areas, as they are more likely to happen from either end of the fence, and will allow the management team to deal with incursions in a more effective and quick manner. The partition of the fenced area based on topography and ownership also opens the possibility to build the fence in stages, protecting one or more sections at a time (Image 75). However, it will be more cost-effective to build the totality of the fence at once. Table 13 summarizes the characteristics and features of each section.

Image 75. Proposed location of the predator-proof fence and partitions for four optional sections.



Table 13. Reserve sections (from west to east).

	<b>Plémont headland</b>	<b>Plémont to Creux Gabourel</b>	<b>Vinchelez common</b>	<b>Le Rocquerel</b>
<b>Puffins:</b> Presence / Potential if protected	Absent Very high	<b>Present</b> Very high	<b>Present</b> Very high	Absent Good
<b>Razorbills</b> potential	Absent High	<b>Present</b> Very high	<b>Present</b> Very high	Absent Good
<b>Guillemots</b> potential	Absent High	Absent Very high	Absent Very high	Absent Good
<b>Manx Shearwaters</b> potential	Absent Very high	Absent Very high	Absent Very high	Absent Good
<b>Storm Petrels</b> potential	Absent Very high	Absent High	Absent High	Absent Good

<b>Habitat management</b>	Ongoing	No	Ongoing at 'puffin village'	Ongoing at 'puffin village'
<b>Human activity level (on land)</b>	High	Very low or none	Low	Very low or none
<b>Notes</b>	Habitat in good condition for most target species, their absence probably reflects the presence of predators and humans	Habitat needs mainly bracken management, high potential for many target species	Habitat needs mainly bracken management, high potential for many target species	Habitat needs mainly bracken management, high potential for many target species
<b>Fence length (m)</b>	526	1,069	915	428
<b>Protected area (m<sup>2</sup>)</b>	141,570	69,129	83,359	29,229
<b>Ratio: m<sup>2</sup> protected per metre of fence</b>	269.1	64.6	91.1	69.2

Pest-proof fences are expensive to construct, requiring a high initial investment and a long-term commitment to ensure integrity and biosecurity. For every hectare conserved, a larger fenced area is generally more cost-effective than a smaller one<sup>91</sup>; and for every meter of fence built, a headland is more cost-effective to protect than a length of coastline. This makes the Plémont headland the section of the reserve most cost effective to protect (Image 76). However, even a relatively small fenced area will be more cost effective than ongoing control of invasive animals in an unfenced area.

Image 76. Plémont headland as seen from land.



Although it is difficult to assess the lifespan of the pest-proof fence in Jersey, evidence from other fences around the world suggests that a well-maintained fence in a coastal or high salt-prone area could last up to 30 years, depending on the materials used. Although the proposed site has sections of steep cliffs and rugged terrain, it is feasible to construct on it a predator-exclusion fence (Image 77). It would then be possible, using current eradication techniques and supporting technology, to remove the invasive species from within the newly created seabird reserve.

Image 77. Example of a predator-exclusion fence in a coastal area.



### 6.2.2 Removal of invasive mammalian predators

There are increasing numbers of successful eradications, especially of introduced mammals. The range of species eradicated increases as new techniques are perfected, with for example 21 different species of introduced mammals eradicated alone from 17 islands in the Galapagos archipelago off Ecuador<sup>92</sup>. Species eradicated at other sites include foxes<sup>93</sup>, mice, ship rats, Pacific rats, cats, rabbits, goats, and pigs.

As determined by the field studies, all target species have been detected in the study area, and all except the cat have been detected inside of what would be the proposed reserve area. All four target species have been the subject of successful eradications in other islands. The operational plans, methodologies and biosecurity strategies from such projects are available for consultation and can be adapted to local populations and environmental conditions. It is likely that by the time the fence is installed the density of invasive predators will be lower than at present, as the construction works, likely to be in place for 4-6 months, might drive away from the area some of the target animals, making them retreat to other parts of their territories. This will reduce numbers inside the fenced area and will also reduce the costs of eradication.

Monitoring of both target species and native wildlife will continue throughout eradication works, to help ascertain changes in the density of target species as the removal progresses. In regards to live trapping, Jersey has an established protocol on how to process feral cats and ferrets. In the case of cats, it involves trapping the animal humanely and handing it over to licenced organisations such as the Jersey Society for the Prevention of Cruelty to Animals (JSPCA) and the Cat Action Trust Jersey. These organisations are professionally recognized to find the owner of an escaped pet, or in the case of a feral animal, to evaluate its health and character in order to make a decision on its outcome; be it euthanasia or re-homing. Ferrets are classified as pests in Jersey and, although they can be taken

to the JSPCA for evaluation and possible re-homing, they can also be humanely euthanized by a licenced pest controller. The guidelines for the removal protocols will be detailed in the operational plan for the eradication and relevant risk assessments that will be produced in preparation of this work.

### **Brown rats**

Over fifty decades of rat eradications have made New Zealand the home of global experts on this matter. There, a string of local extinctions of island endemics caused by rats in the 1960s' prompted the start of formal work to eradicate rats. Since then, teams of wildlife managers, ecologists and scientists have designed, tested and perfected a wide array of methods to successfully eradicate rats from islands; techniques and knowledge which has been useful all around the world ever since<sup>108,109</sup>.

It is surprising and perhaps unexpected then, that the first documented successful rodent eradication, in 1951, happened not far from Jersey at all: it was in Rouzic Island (Île Rouzic), in the Sept-Îles archipelago off Brittany and less than 60 miles from Jersey<sup>94</sup>. This, like many other early eradications, were generally an unintentional effect of normal rodent control; however, their unanticipated results prompted efforts to investigate and develop eradication methods.

Nowadays there is a vast reservoir of literature and technical resources on the subject, as much has been learnt from decades of control programmes around the world. Success in removing rats from islands has been achieved from the Mediterranean to the Antarctic, from the smallest islands, such as Maria Island, New Zealand, with 1 hectare, eradicated in 1960, to what once was thought to be an impossible, unsurmountable task: South Georgia, in the South Atlantic, spanning over 352,800ha, where the largest eradication project to date took place, and which was declared rat-free in 2018.

Worldwide, there have been 332 successful rodent eradications to date, with invasive rodents eradicated from 284 islands. With the exception of two islands, all the successful campaigns have used rodenticides as the main or only method of eradication.

For the task of eradication, the type and scale of techniques has evolved from the most traditional trapping (with live or kill traps, which generally fails to remove all individuals due to the species avoidance behaviour), to the most common and cost-effective method: rodenticide-poisoned bait. The efficiency, secondary effects and environmental effects of various types of bait have been evaluated in several field trials, as well as the behavioural responses to different designs of bait station. Besides these two methods of control, other avenues of research are exploring the effectiveness of contraceptive methods, and there are several formulations in development which may make one day oral immunisation a possibility<sup>95</sup>.

As it is the case for most traditional eradications, the cost and effort involved increases rapidly as the target population decreases, and the most difficult individual to be trapped or killed is usually the last one. A single animal, especially in the case of a rat, is likely to display atypical behaviour, due to the lack of conspecifics which are a source of reference behaviour and competition for resources, amongst others. Therefore, the last individual of a population might be less likely to investigate a trap or bait station, and might also have less need to, since it has no competition for other more available and natural sources of food<sup>96,97</sup>.

The recommended strategy for rat eradication involves a combination of techniques that will complement each other in terms of removal, monitoring of population changes, and detecting remaining individuals.

Recommended strategy for rat removal:

- a) Determining and monitoring the presence of brown rats within the fenced area via a combination of baited camera traps, rodent detector cards (Goodnature®), dusk surveys and searches for tracks and signs.
- b) Deploying feeding stations with live traps in areas where rats have been trapped in the past, in other locations where they have been detected by surveys and in additional locations which are deemed suitable for rats.
- c) Deploying bait stations in areas where rats have been trapped or detected, as well as on a grid of 25-50m squares.
- d) Deploying kill traps (Goodnature® A24-type) specific for rats, where rats have been trapped or detected, as well as on a grid of 25-50m squares.

### European hedgehogs

Hedgehogs have been successfully eradicated from the Uists Islands, Scotland, in what is so far the only known attempt at eradicating this species from an island group. After determining that the hedgehogs were present in both islands, and were responsible for 52% of all predation in South Uist, The Uist Wader Project was set up in 2000 to remove all the hedgehogs from the Uists. Various methods for monitoring and trapping were trialed and improved, including lamping, motion-activated cameras, footprint tunnels, sniffer dogs and live traps. Furthermore, a trial eradication was carried out to estimate the costs and efforts of scaling up the work to encompass both islands, and an Index of Abundance (IOA) was developed using footprint tunnel data, in order to monitor the impact of the removal activities on the hedgehog population<sup>61</sup>. At the early stages of the project the hedgehogs that were caught were humanely euthanized, but this drew criticism from some sectors of the general public, due to the fact that the species is native in mainland and is popular amongst gardeners and nature enthusiasts. An agreement was reached with stakeholder organizations and authorities that established a protocol whereby the trapped hedgehogs were released in the mainland by a licensed organisation. The number of trapped hedgehogs declined steadily until only one was trapped in 2017; this rate suggested that eradication could be confirmed by end of 2020.

The success in the Uists gives us reason to believe that it is also possible to remove all hedgehogs from the seabird reserve once the fence is installed. The area is much smaller and the number of hedgehogs in it is very small and potentially null. Even though we might not need to implement all the techniques developed by the Uist Wader Project, the lessons learnt by this project, technically and socially, are helping plan and guide the implementation of the hedgehog removal in the seabird reserve.

Recommended strategy for hedgehog removal:

- e) Determining and monitoring the presence of hedgehogs within the fenced area via a combination of baited camera traps, dusk surveys and searches for tracks and signs.
- f) Deploying feeding stations with humane traps in areas where hedgehogs have been trapped in the past, in other locations where they have been detected by surveys and in additional locations which are deemed suitable for hedgehogs.

- g) All hedgehogs caught will have a health check (weight, shape, breeding status). Hedgehogs which appear in good health will be released at the closest suitable point outside of the reserve. If the individual is marked, it will be released at the point closest to its last known location outside of the reserve.

### **Feral ferrets**

The impact of feral ferrets on native wildlife is a problem for conservationists worldwide. Many techniques have been tried to curb their effect, from changing their behaviour to reducing their numbers or modifying their habitat. In New Zealand, the traditional method has been by trapping and culling, although poison has also been found to be a suitable alternative<sup>98</sup>. A bait station that is used by ferrets but which excludes dogs and cats has been developed, alongside various palatable baits which can integrate the poisonous chemicals<sup>99</sup>. In the Orkney Islands there is presently an ambitious project underway which aims to eradicate all stoats from the archipelago. The stoats is a relative of the polecat and ferret, and having been introduced in the islands by humans, the result has been a decline of many native species and in particular ground-nesting birds. The Orkney Native Wildlife Project is using a team of 20-25 staff and over 7000 lethal traps to achieve complete eradication. The skills and field techniques gained by this project can potentially be of great help to the Seabird Reserve Project<sup>100</sup>.

Other methods that have been tried include conditioning ferrets to avoid certain prey species<sup>101</sup>, or modifying the habitat surrounding bird breeding sites, by planting buffer zones of long grass around breeding sites on coastal grassland. These techniques have had mixed results, as the present view is that habitat modification has a limited applicability for farmland areas<sup>102</sup>. The potential suitable habitats are near or adjacent to agricultural fields, therefore it would not be possible to implement this technique at the site.

Many lessons learnt from projects in New Zealand can be applied to the removal of feral ferrets from the reserve, including the fact that bait is most effective when laid in late summer, autumn and early winter, and that ferrets are overall more difficult to trap in the spring<sup>103,104</sup>. The experience gained by the author during the field work carried out from 2018 to 2020, in particular the work involving trapping and radio-tracking feral ferrets in the study area, will also help implement this eradication plan in an effective and safe manner.

Recommended strategy for feral ferret removal:

- a) Determining and monitoring the presence of feral ferrets within the fenced area via a combination of baited camera traps, dusk surveys and searches for tracks and signs.
- b) Deploying feeding stations with humane traps in areas where ferrets have been trapped in the past, in other locations where they have been detected by surveys, and in additional locations which are deemed suitable for ferrets.
- c) Handing over any trapped ferrets to a licenced organisation in the understanding that should the animal be re-homed, it would also be neutered and micro-chipped, and would not be re-homed in the north-west part of the island (parishes of St Ouen and St Mary).

## Cats

The status of cats as pets and domestic animals in most countries raises moral dilemmas on how to manage their impact when they threaten native wildlife, despite the approximately 275 million animals killed by 9 million cats in Britain alone<sup>105</sup>. Even so, cats have been successfully eradicated from many islands with the help of various methods.

Feral cats have been successfully eradicated from islands in all the world's oceans, including some large, remote and challenging places such as Faure (Australia, 5,241ha) and Marion Island (South African sub Antarctic, 29,000ha). The Database of Island Invasive Species Eradications<sup>106</sup> lists 130 cat eradication projects, of which 106 were successful on first attempt and many more have been successful after being re-attempted. These eradications used a variety of methods, sometimes implemented in combination, such as leg-hold traps, hunting, poisoning, cage traps, and dogs. In New Zealand all cat eradications have eventually been successful, and the country's Department of Conservation has produced a 'best practice' draft document for eradication of feral cats from islands which will improve the prospects and efficiency of any future eradications.

Recommended strategy for feral cat removal:

- a) Determining and monitoring the presence of cats within the fenced area via a combination baited camera traps, dusk surveys and searches for tracks and signs.
- b) Deploying feeding stations with humane cat traps in areas where cats have been detected as well as a grid of potential cat travel routes.
- c) Handing over any trapped cat to a licenced organisation in the understanding that should the animal be re-homed, it would also be neutered and micro-chipped, and would not be re-homed in the north-west part of the island (parishes of St Ouen and St Mary).

These three main actions are to be implemented simultaneously and for the period that is deemed necessary at the time. The number of cats likely to be in the area is small or null, as the species has not been detected in the proposed reserve area. Their presence is also likely to be further reduced during the construction works, hence no cats will be expected to be found inside the reserve. However, monitoring will be carried out as per the operational plan, and subsequent biosecurity protocols will guide the response if a cat were to be encountered.

### **Overall likelihood of eradication success**

Standing on the body of reference from the cumulative experience from hundreds of successful eradications worldwide, combined with local and national expertise from similar British projects, there is no reason to believe that a full eradication of the four target species would not be successful, especially if undertaken in a properly resourced and professional manner. As with all eradications, success cannot be guaranteed, but if the best practice is applied and scaled locally, and the adequate resources are made available, a positive outcome would be expected.

## 6.2 Sustainability

The concept of sustainability is understood, in this context, as the ability to maintain the seabird reserve free of invasive mammalian predators after the installation of the fence and eradication.

This will entail minimizing the risk of re-invasion by the species that have been removed from the reserve as well as of new invasive species which might become a threat in the future. Managing these risks requires putting in place effective biosecurity measures to cover not only aspects relating to the physical environment of the reserve, but also the socio-cultural and political environment. All these measures will be detailed in the biosecurity strategy, but in brief they will aim to:

- A. Prevent invasion or re-invasion by long-term passive means.
- B. Detect any invasion via ongoing monitoring and surveillance.
- C. React to any invasion with a response to ensure the removal of the invader.

The biosecurity strategy will look into detail at all potential risks of re-invasion, by species and pathways, and will lay out a plan to prevent, detect and eradicate each of them with feasible management options.

At present the land and sea that would comprise the Seabird Reserve have clear pathways of access; the main one is by land, but can also be accessed by any landing craft coming from the sea. Once the fence has been installed, the most likely mechanism for mammalian predators to re-invade will be through damage or malfunction of the fence, accidental transport via sea by humans, or via sea by directly swimming from land adjacent to the reserve (Table 14).

Table 14. Potential invasive pathways for the Seabird Reserve and summary of prevention strategies.

By Land	Risk (Low/Medium/High)	Prevention strategy
Damage in fence.	Medium, but highly dependent on size, location and origin of damage (i.e. weather, debris, accident). Rats and hedgehogs might be more prone to exploit damage near the ground, whilst ferrets likely to be wearier, and cats less likely to cross a newly-formed aperture.	Routine visual checks on fence integrity. Response protocol and contingency budget to repair damage in the fence as soon as it is detected.
Malfunction of fence gates.	Low, as design will comprise a double-gate security system which is unlikely to fail on both gates at the same time.	Routine visual checks on gates. Motion-activated cameras in the vicinity of gates to detect possible malfunctions and incursions from target species.
Accidental release via gates.	Low, as it would be unlikely that any of the target species would cross unnoticed the double-gate system when in use.	Awareness campaigns and clear signaling on proper operating of gates. Motion-activated cameras in the vicinity of gates to detect accidental incursions from target species.



Intentional release via gates.	Low, but depending on positive community and awareness work.	Awareness campaigns and positive community engagement. Motion-activated cameras in the vicinity of gates to detect intentional releases of target species.
<b>By Sea</b>	<b>Risk (Low/Medium/High)</b>	<b>Prevention strategy</b>
Accidental introduction from boats approaching, mooring or landing on the reserve.	Low for hedgehogs, ferrets and cats, as they are likely to be noticed in a boat and they are not likely to abandon a ship and swim to shore. Medium or high for rats, as they are less likely to be noticed on a ship and can swim on open water, but also dependent on the type of ship and the state of the sea at the time.	Awareness campaigns for leisure craft owners. Facilitation of checks and response protocols for commercial fisheries operators. Review of mooring and landing regulations in the Seabird Protection Zone.
Swimming from land adjacent to fenced reserve.	Low for hedgehogs, ferrets and cats, as they are less likely to swim across open water. Medium for rats, as they can swim confidently on open water.	Rodent bait and traps near the ends of the reserve, at both sides of the fence where it meets the sea. Motion-activated cameras at the ends of the fence to detect target species venturing into the sea.

The seabird reserve, like any other natural protected area, cannot be made totally secure from potential reinvasion, as some events will be beyond the control of managers and authorities. However, many pathways can be effectively managed with the implementation of a robust biosecurity strategy.

The advantages that the reserve will have towards its biosecurity is that there already exists a high level of public awareness on local nature conservation and that there is support of the local community to habitat and species restoration work, as well as a sense of local pride on the natural heritage of the island. The public and authorities are also aware of the health and economic benefits associated with the restoration of wild landscapes, and the local businesses that operate in the area (tour operators and commercial fisheries) support this project and have indicated their interest in collaborating on the production and implementation of the biosecurity strategy. Furthermore, the area features mainly steep cliffs, rocky headlands and prevalent swells, which lessens the interest and likelihood of recreational visitors to moor near or land on the reserve.

### 6.3 Social acceptability

The final goal to all conservation projects is to benefit the human population, whether directly or indirectly. But even though the result of a project will have an overall positive impact, the response of the local community can be very varied and will depend on how the project affects their economic, cultural and historical values. Furthermore, no conservation work is ever free from some negative impact, be it to non-target species or to aspects of the locals' way of life. A negative impact is often subjective and depends on the persons' or organisation values upon the project site and its natural community.

The support and engagement of the local community is critical for the success of any project, from the planning and implementation phase to the long-term sustainability of this effort. In order to secure this support, it is paramount to involve the community at a very early stage – this will not only help understand, prepare, minimise and mitigate any negative impacts of the community, but might also help identify any impacts not foreseen by the management of the project. It is indeed the importance of this local knowledge and positive engagement which will conduit towards a real participation and sense of ownership of the project from the local community.

In the case of eradication initiatives, there is usually a positive reaction from the local community, as it has likely suffered the direct effects of the invasive species, especially if it is a rodent. The impacts might have been direct, via depletion of natural resources, attacks on crops or vectors of disease; or indirect, via loss of income from tourist and hospitality venues due to predation on key species and degradation of habitats. In those cases, the public response to the eradication initiative tends to be of engagement and support. In the case of predator-proof fences, there is a smaller body of tradition and expertise on their implementation. Some of the issues of concern might be matters of visual impact, access, and long-term management. Most projects involving a predator-proof fence have succeeded by engaging a strong base of community support, with public consultations, education campaigns and ongoing activity programmes.

In Jersey there might be, too, concerns in regards to the visual impact of the fence. These will be addressed firstly by landscaping the fence within its natural environment and placing as much of it as possible below the line of sight from the public footpath. This issue will also be addressed during informal discussions that have already started as part of the public consultation process (Image 78). Another potential concern might be that of access during the installation works or even after the fence is completed. The planned reserve has few spots which are visited by the public: the Plémont headland (for a variety of leisure activities) and two other smaller, less accessible sites (mainly for the purpose of angling). Access to all these spots might be reduced during certain times of the installation process, but it will be regained in the long-term via access gates, which will hopefully address the concerns on this matter (Image 79).

Image 78. Example of a predator-exclusion fence below the line of sight.



© John Cooper

Image 78. Example of an access gate on a predator-proof fence.



© John Cooper

The eradication, translocation or removal of the invasive species is likely to be met with mixed responses. There is the potential for some emotive reactions amongst sectors of the wider public to the suggestion of using lethal methods for eradication, or to restricting hedgehogs from parts of their ranges. Either of these options have not been formally decided and other courses of action remain open, although it is important to educate on the different costs and time scales involved on each option.

It is almost certain that there would be some negative effects on some of the non-target species. If bait is used as one of the methods to remove rats from the reserve, its effects on the community will be short-lived. The effects of a fence will also impact on populations of non-target mammals and reptiles. The implications and mitigations of this impact are reviewed in detail in Section 6.5. Managers and field professionals involved in ground-level conservation work have to be pragmatic

about such matters, whilst making informed decisions that balance the goals of a project against its negative impacts. Members of the general public, if they are more detached from this process, might express concern about the effects on non-target species. There might even be a broader debate over the relative merit and priority of creating this seabird reserve, however preliminary responses show an inclination by most consulted to support the protection of Jersey's puffins and other local seabirds, based mainly on a sense of pride for the Island's natural heritage as well as enjoyment of an iconic and charismatic species at such accessible spot.

The consultation process for the Seabird Reserve has to focus on the installation of the fence and subsequent eradication of invasives from within. It needs to prioritise a transversal inclusion of all groups and interested parties, and present in a clear manner what to expect of the process. The process will collect and clarify the concerns of the community, form a platform where solutions can be debated and ensure that all parties are informed on how their expectations will be met.

The consultation process for the reserve has already been initiated in various ways:

- During casual conversations between the project officer and neighbours, site users and passers-by during various tasks in the field.
- During discussions with local leisure operators and hospitality businesses which are invested in the improvement of the natural environment of the area.
- In meetings with stakeholders and land owners to discuss management options and feasibility of the goals and aims.
- In meetings with Parish officials and Government politicians to potentiate the involvement of the general public on this project

There is scope to increase the involvement of the local community in the project as well as the general awareness on the problems facing Jersey's puffins and other seabirds, and on the solutions that are feasible for the Island. Besides the consultation, two additional lines of work are addressing these matters at present: An awareness strategy and a community engagement plan.

The awareness strategy plans to increase public knowledge of the state of Jersey's puffins and the threats they face, as well as aims and objectives of this project. It will work on various communication fronts such as:

- Press releases and articles on local media, newspapers, websites and magazines.
- Updates and links to partners and other relevant projects on social media outlets.
- Public talks at schools and colleges, parishes, local cafes and other community centres.

The community engagement plan consists of:

- Public activities which will offer the participation of the different sectors of the community, such as children and parents, during public and online events.
- Provide the opportunity for local students to participate in an ambassadors' program to take part in various aspects of the ecological work and to represent the project in their schools and communities.
- Learning opportunities for students of institutions such as JICAS, and internships in the field of ecological monitoring and management for local students of STEM disciplines.
- Training and volunteering opportunities for the general public in a variety of research, management and outreach work of the project.

The project will also benefit the Island's society by promoting local economy and cultural values. The project will directly employ the services of local companies, contractors, suppliers and professionals,

and some parts of the project will be undertaken in liaison with apprenticeship programs, education institutions and back to work schemes, in order to offer training in new skills to the local workforce. It is also believed that an increase of biodiversity and species richness in the area will promote the value of the site as a tourism asset, creating new opportunities for local travel agencies, outdoor leisure operators, hospitality businesses and small commerce. The project officer has already engaged the collaboration of various local businesses such as cafes, shops, birdwatching and other leisure operators via our awareness campaigns and interpretation materials. It is believed that Islanders place in high value the puffin as a charismatic and iconic species, and have a sense of pride on its presence in Jersey. The species is also strongly tied to the cultural history of the Island; present in Islanders' daily lives with abundant depictions on local art, household items and souvenir artifacts.

Furthermore, this project intersects with various common themes that have arisen from the Government's Common Strategy Policy, such as enabling Islanders to lead active lives, promote Jersey's profile and reputation internationally, work in partnership with parishes, community groups, volunteers and key stakeholders, and to nurture a diverse and inclusive society. Various stakeholders for this project have been engaged at the early stages and have been working in different fronts to offer their input and expertise, as well as to incorporate in this process the concerns and issues encountered during the consultation. The stakeholders will play a crucial role in strengthening ties with the local community via their roles in policy, education, employment and outreach (Table 15).

It is believed that the ecological and economical benefits of this project will outweigh the negative impacts and inconveniences on the local community, and that a straight and thorough process of consultation which engages, listens and acts upon the public's concerns can result in an overall support from the community.

The importance of the community's support for the long-term viability of this project cannot be overstated. One of the most successful predator-fence projects to date is a community-led sanctuary project which managed to raise \$1.9 million in order to build 8.2km of fence, to create a predator-free wildlife reserve in Taranaki, New Zealand<sup>107</sup>. Since the completion of the fence the Rotokare Scenic Reserve Trust has launched several habitat restoration and wildlife translocation projects as well as producing a large body of educational material and events for schools and visitors.

Image 80. Members of the Rotokare Scenic Reserve Trust maintaining the predator-proof fence that they manage.



© [www.rotokare.org.nz/Gallery](http://www.rotokare.org.nz/Gallery)

Table 15. List of stakeholders identified.

Stakeholder	Representative/s	Notes
Natural Environment Department	Tim Liddiard, Senior Natural Environment Officer Willy Peggie, Director	Partner of Birds On The Edge. Landowner of two thirds of Reserve's area. Undertakes habitat management in area.
Durrell Wildlife Conservation Trust	Glyn Young, Bird Dept Director	Partner of Birds On The Edge.
The National Trust for Jersey	Charles Alluto, CEO	Partner of Birds On The Edge. Funding and employing the Project Officer.
Vinchelez Estate		Private owner of one third of Reserve's area.
Housing, Infrastructure and Environment Department	John Young, Minister Gregory Guida, Deputy Minister	Government representation of environmental interests and policies. Have expressed support for the project.
Economic Development, Tourism, Sport and Culture Department	Deputy Kirsten Morel	Supporting cultural heritage and promoting community engagement and artistic opportunities.
St Ouen Parish	Richard Buchanan, Connectable Richard Renouf, Deputy	Local authorities and management of inhabited area, community liaison and regulations. Supportive of project proposed. Offered to moderate local platform for public consultation.
Geomarine Ltd	Iain Barclay, Operations Director	Main contractor on installation works. Local expertise, employment and apprenticeship opportunities.
Wildlife Management International Limited	Biz Bell, Director	Technical consultant on overall project but primarily on fence (design, installation, maintenance) and eradication work (planning, implementation, biosecurity).
Office of the States Vet	Dr. Alistair Breed	Compliance and legislation, technical assistance and licensing of trapping and other animal management work.
Animal welfare NGOs: JSPCA, Cat Action Trust, and Jersey Hedgehog Preservation Group	Dr. Jo McAllister Jean Falk Dru Burdon	Training and technical assistance on capture and handling of target species. Ultimate depositories of feral ferrets and cats removed from the Reserve.
Jersey Coastal Park	Mike Stentiford	Reserve is within the Park boundaries, positive impact on overall aims and goals of Park designation.

Société Jersiaise and Ornithology Section	Nicky Westwood, Stuart Fell, Mick Dryden, Tony Paintin	Local expertise on terrestrial and marine habitats, natural communities, bird populations, technical assistance on best practice, habitat management and timings of work.
Conservation NGOs and groups such as JAFW, JARG, Marine Conservation, Jersey Bat Group, JCV and various sections of the Société Jersiaise.		Local expertise and technical assistance with components of the ecological work, developing broader aspects such as sustainability and environmental responsibility, assistance in planning long-term ecological monitoring, volunteer activities and other local engagement.
Fisheries and Marine Resources Jersey Government	Francis Binney, Marine scientist Paul Chambers, Marine & Coastal manager	Technical adviser on ecological monitoring and liaison with other marine organisations.
Commercial fisheries and Association of Boat Owners (Grève de Lecq)	Royston Rimeur Avril Rimeur	Working or visiting area by sea, with implications in biosecurity. Will commercially benefit from regulations on mooring and landing in the area. Ongoing communications and have expressed support for the aims of the project.
Outdoor and leisure operators	Jersey Seafaris Jersey Kayak Tours	Visiting the area by land or sea. Will benefit from increase in seabird colonies and habitat restoration. Participation in consultation regarding access and regulation of traffic in the area.
Visit Jersey	Amanda Burns Rebecca Collins	Major beneficiary of the project. Consultation on promoting the area and educational campaigns of proposed work and future access regulations.
Jersey Hospitality Association Local establishments	Jeremy Swetenham, Chair Emma Machon, Cafe owner Paul Mc Dermott, Cafe owner	Major beneficiaries of the project. Have expressed support for project and there is ongoing collaboration in awareness, education and community events in area.
Jersey International Centre of Advanced Studies (JICAS)	Dr Amy Hall, Director	Opportunities to collaborate on long-term ecological research and practical conservation management.
Local community		Ultimate beneficiaries of project. Employment opportunities both direct and indirect. Improved quality of life.

## 6.4 Political and legally acceptability

Fences are built in Jersey for a variety of reasons: to protect equipment and property, to protect habitats and humans from hazards, and to contain pets or livestock, to name a few. Some animals are legally regarded as pests, such as brown rats, and rodenticides are commonly used to eradicate them by the general public, professional pest controllers, and many sectors of the industry. Rats and feral ferrets can also be humanely killed by licenced professionals after being trapped. Other times, feral ferrets, like feral cats, are trapped and assessed for the purpose of re-homing as pets. Wild hedgehogs found in poor health are rescued and cared for local specialists, with many being released back into the wild at or near their territories.

There are many local precedents for the various works involved in this project, and no legal impediments to the installation of a predator-proof fence, nor to the removal and management of the target species based on the options described in Section 5.1.

However, there will be a need for a consultation with various bodies in the Government to understand all legal compliance requirements and processes, and a detailed list of all permits and other authorisations needed as well as a description of the application process. Typically, this would include, but not be limited to: licenses to disturb wildlife and habitats during installation works, trapping and handling, rodenticide application, use of lethal traps, protection of native species, protection of natural resources, and protection of any historic or cultural resources on the area of the reserve. It is envisaged that memorandums of understanding will be agreed between the project management and the licensed animal organisations which will take final custody of any trapped feral ferrets or cats.

The fence in particular would require written permission from the two landowners involved, which have already verbally agreed, as well as the procurement of planning permission. Other legal requirements will likely be the presentation of an environmental impact statement and the present feasibility study.

In political terms, this project will yield benefits for the Island's nature, economy and community, by supporting the following pledges of the Government's Common Strategic Policy (CSP)<sup>108</sup>:

### 1. To protect and value our environment

This project embraces environmental innovation and ambition, aiming to protect a natural environment through conservation, management and sustainable use of resources. It will enhance local biodiversity, protect the Island's natural heritage and retain the character of this particular landscape. With the puffin colony safe from predators, many seabirds and land birds in the area will potentially have a chance to recover and thrive, including razorbills, guillemots, Dartford warblers, stonechats and linnets. The land adjacent to the proposed fenced area, recently restored by the National Trust, is a preferred feeding and resting spot for wintering skylarks and other farmland birds. All of them will benefit from predator-free foraging opportunities on the other side of the fence. On ecological terms the puffin is an 'umbrella species', which means that its protection and that of its habitats will also benefit other endangered key species native to Jersey, such as the green lizard, the slow worm and the Jersey bank vole. This project is compatible with the current Planning Law, with the Jersey Biodiversity Strategy, and with the Conservation of Wildlife (Jersey) Law. Furthermore, this project will demonstrate to global partners that the Island takes its environmental responsibilities seriously, as it falls in line with international conventions that Jersey is signatory of, such as the Convention on Biological Diversity, as well as the Bonn and OSPAR Conventions<sup>109</sup>.



Furthermore, this project aligns with the Government of Jersey in its aims to:

- Protect areas of coast to improve our stewardship of these areas;
- Improve abundance of key indicator species;
- Protect and increase biodiversity;
- Improve access to open and green space for a variety of users, with an intervention that will have benefits over generations, and not just for the short term.

## 2. To put children first

As part of its community engagement strategy, the project will launch events and activities aimed at children which will improve their educational opportunities and involve them in various community-based aspects of the project.

## 3. To improve Islanders' wellbeing and mental and physical health

The restoration of this coastal area and the increase of biodiversity that the project hopes to generate will attract people to this area, engaging visitors with the outdoors and promoting a variety of physical activities such as walking, swimming or simply bird-watching. The public recognizes and values the environment, as evidenced by the responses to the Future Jersey public consultation. Time spent in this natural space will likely contribute to the mental and the physical health of many Islanders.

## 4. To create a sustainable and vibrant economy and skilled local workforce for the future.

The project will directly employ the services of local companies, contractors, suppliers and professionals. Some parts of the project will be undertaken in liaison with apprenticeship programs, education institutions and back to work schemes, in order to offer training in new skills to the local workforce. An increase of biodiversity and species richness in the area will promote the value of the site as a tourism asset, creating new opportunities for local travel agencies, outdoor leisure operators, hospitality businesses and small commerce.

Whilst this project is consistent with the CSPs and other policies put in place before the Covid-19 pandemic, it is also compatible with the Government's pursue of an Island-wide 'green recovery' and new post-Covid strategic aims.

## 6.5 Environmental acceptability

A risk-assessment for the installation of the predator fence will be required to evaluate the potential harm of this feature to wildlife and the environment. The assessment should evaluate direct effects such as disturbance, displacement and habitat degradation, as well as indirect effects such as litter, debris and chemical spills. It should encompass all aspects related to the process of creating the fence, from early stages of transportation of materials and ground preparation, to installation works and long-term effects on its structural integrity, such as degradation, leeching and accidental breakage.

Once installed, the fence itself will create a physical barrier to stop invasive predators from accessing the seabird breeding grounds of the reserve, as well as the many coastal habitats within it. At a mesh square of 7mm, the fence will exclude the smallest threat envisaged - a brown rat at time of weaning, and anything above that in size, encompassing all the target species. The fence will have the unwelcome effect of stopping the movements of some non-target species which are too big for the mesh. It is likely that this will include a portion of the populations of field mouse, bank vole, shrews, green lizards, slow worms and possibly all the rabbits. It is believed that the overall result of the eradication of the invasive predators will be a positive effect on most non-target species by eliminating this unnatural source of predation, and that their populations might recover to higher densities close to pre-invasives time. Any potential demographic imbalances due to the lack of movement across the fence might be naturally readdressed by an increase of pressure from native predators, such as buzzards, kestrels, marsh harriers, barn owls, short-eared owls, long-eared owls, crows, ravens, and gulls of various species. A long-term monitoring and management strategy will be required to ensure that any possible demographic changes and ecological imbalances in these non-target species are recorded, and that protocols are in place should the need for management arise.

The eradication of invasives from the reserve after the fence is installed will involve trapping, handling and potentially the death of some target species, particularly of rats. Humane traps have already been used in the area to conduct surveys with regular success on some of the target species (ferrets, hedgehogs and rats) and a very low record of incidents. The only non-targets that have been caught in live traps were rabbits that had most likely wandered inside the traps. Lethal methods such as rodenticide bait and lethal traps have not been used in the area, but their use will be necessary in order to counter the ability of some rats to evade humane traps.

A non-target risk assessment will be required to evaluate the potential harm to native species from the lethal traps and from exposure to the proposed rodent bait, and to detail measures to mitigate potential effects. The effect of the rodenticide on the marine environment will also have to be considered, although it is unlikely to have any, as bait will not be distributed into, or towards, the sea, and will be kept inside bait stations. It is possible that some bait pellets might enter the sea by being dragged outside the bait stations by rodents and then rolling off or being swept off coastal cliff edges, but at such low density it is expected to be below trace amounts. The transportation, storage, broadcasting and disposal of bait and associated equipment should be covered in the operational plan as well as the non-target risk assessment (Table 16).

Mortality of native vertebrates due to non-target poisoning has been documented, but evidence shows that affected species quickly recover to pre-eradication population levels or higher. A variety of methods have been developed to mitigate those impacts, and applied research can further aid in minimizing them. In the non-target risk assessment, the benefits of long-term species protection, recovery, and habitat restoration should be considered, along with any potential negative impacts through mortality, displacement or demographic changes. Mitigation options should be listed along

its potential benefits and costs. Actions to reduce the risks to non-targets might involve various methods, or combinations of methods, such as timing of eradication works, captive-holding, aversion training or antidotes. Some actions will be potentially expensive and will increase the complexity and costs of the project, therefore the potential negative impacts on the target species will have to be balanced against the costs and the biological, cultural or socio-political reasons for implementing certain mitigation techniques.

Table 16. Preliminary list of non-target species that might require mitigation actions to avoid negative impacts.

Non-target taxa	Species*	Potential impacts and mitigations
Small mammals	Field mouse, bank vole, shrew spp.	Portion of the populations not able to move across the fence, potentially isolating the populations. Monitoring of population, demographic structure and genetic variability recommended. Rodents might be at risk from exposure to bait; baiting should happen during winter when small rodents are hibernating.
Large mammals	European rabbit	Population will be isolated genetically inside the reserve. Monitoring of population size and genetic variability recommended. If density increases above carrying capacity, potential measures include one-way flaps in fence to allow individuals to exit reserve, translocations or culls.
Reptiles	Green lizard, slow worm	Portion of the populations not able to move across the fence, potentially isolating the populations. Monitoring of population, demographic structure and genetic variability recommended. If genetic variability is at risk potential measures include translocating individuals from more diverse sub-populations.
Breeding birds: land birds and seabirds	European stonechat, linnet, dunnock, wren, Dartford warbler, rock pipit, swift, Eurasian kestrel, razorbill, puffin, lesser black-backed gull, herring gull, great black-backed gull, fulmar	Disturbance during breeding season, breeding habitat damaged or destroyed. Most small land birds build a different nest each year within a territory. Seasonal nests might fall on the path of the fence; however, seabirds and cliff-nesters would not be on the path of the fence. Number of annual nests on the path of the fence is believed to be small due to the presence of bracken which is not suitable for nesting for most bird species. Mitigation measures should involve scheduling the installation works outside of the bird breeding season, and the restoration of habitats, which is in itself a main goal of this project.
Birds: raptors and scavengers	Common buzzard, kestrel, marsh harrier, barn owl, short-eared owl, crow, raven, great black-backed gull	Potential secondary poisoning from scavenging target species. The formulation of the latest modern rodenticides reduces this risk greatly however mitigation measures might include regular checks for carcasses with observers and trained dogs.

\* List not comprehensive. Only a selection of species which have been detected within the area of the proposed reserve are included at this stage.

## 6.6 Capacity

The BOTE partnership has led and coordinated the work involved in this project to present date. The habitat management of the Plémont headland and adjacent coastal cliffs is undertaken by the Government of Jersey Natural Environment team, and involves tasks such as monitoring vegetation, removing bracken and opening clearings from encroached vegetation. The Lands Team of the National Trust for Jersey maintains and manages the restored land directly above the headland, reducing the cover of dominant species and encouraging a mosaic of coastal habitats. In 2017 Durrell Wildlife Conservation Trust funded the visit of a specialist that carried out a preliminary study of invasive predators in the study area, and in 2018 the NTJ commissioned a local consultant to monitor the puffins and other seabirds within the proposed reserved.

The National Trust for Jersey employs the full-time project officer, who has carried out most of the recent fieldwork, ecological surveys, fundraising and public communications, and who has initiated the consultation process. Managers at all three organisations have provided expert advice and steering on the various tasks accomplished, and teams of rangers from NE and the NTJ have assisted the project officer in particular tasks where many hands were required, such as clearing bracken and installing the puffin nest boxes underground.

The BOTE project officer will be responsible for the Seabird Reserve's operational plan as well as various risk assessments, biosecurity strategy, community engagement, awareness campaigns and any additional fundraising required. The officer is also likely to undertake management duties during the fence installation process, ecological monitoring and invasive predator eradication works.

A local contractor has been approached to carry out the fence installation works, and the project has outsourced an international predator-proof and eradication specialist consultancy, to assist with the technical aspects of the project such as feasibility assessments, fence design, and biosecurity plans. The project will also benefit from the experience of a similar project in the Azores Islands, as its management staff has been invited to visit Jersey and share their insights. A local consultant will carry out an environmental impact audit, whilst specific aspects of the wildlife assessment will be carried out in coordination with local experts and organisations.

If the option of using a herd of herbivores to undertake conservation grazing is explored, the landowners will negotiate a lease agreement with a local shepherd and all the administrative and practical responsibilities related to the grazing herd will be taken on by the shepherd or its representative.

The project officer is the only full-time member of staff dedicated to the seabird reserve project, but the BOTE partners' technical, practical and legal expertise is available to the project officer at all times. It is expected that specific manual works, as well as ongoing and long-term habitat works, will be led by the NE team and carried out in collaboration with the project officer and the rangers of the NTJ Lands Team, in continuation of the long-established collaboration between these organisations.

## 6.7 Affordability

A preliminary estimate of the cost of the project, comprising the installation of the fence and removal of invasive predators, is 879,921 GBP. This is a conservative estimate based on quotes sourced to contractors as well as projects of similar scale, which includes a 20% contingency to cover all unexpected costs and variations in matters such as fence route, eradication methods and mitigation actions. The cost for a predator-proof fence can vary a lot depending on many factors, but the usual estimates lies between 200-400 GBP per metre – it is estimated that the proposed fence would work out at 220-250 GBP/m.

The following itemised estimate covers the installation of the fence and removal of invasive predators which will effectively create the Plémont Seabird Reserve. The cost of ongoing ecological monitoring and long-term biosecurity measures will be determined at the completion of the biosecurity plan. Funding has been secured for some parts of the project or portions of them, however the bulk of the work is still to be fundraised for.

Table 17. Indicative costs for the fence installation and eradication works (in GBP)

Planning and Bye-laws	76.55
Fence design	13,500
Fence installation	650,000
Eradication of target species	2,600
Consultant fees	4,000
Nest boxes for puffins, Manx shearwaters and storm petrels	2,000
Ecological monitoring equipment and analysis software	3,500
Visit and training from Azores project representatives	3,700
Puffin decoys	2,100
Project management	31,791
QS Quantity Surveyor	20,000
Contingency (20%)	146653.51
<b>TOTAL</b>	<b>879921.06</b>

## 6.8 Project contingencies

After assessing the feasibility of establishing a Seabird Reserve via the installation of a predator-exclusion fence and the eradication of invasive predators from inside it, it was determined that the following contingencies should be met prior to initiating the project:

- The project is socially acceptable to the Islands' community.
- All licences and agreements are in place for the removal and management of the target species.
- All individuals of the target species can be put at risk by the removal or eradication techniques.
- The target species can be removed at a rate exceeding their rate of increase.
- The biosecurity measures can be deployed effectively and with a long-term continuity plan.
- The environmental risk assessment suggests that the benefits will outweigh the costs.
- Sufficient funding is secured prior to initiate the construction of the fence.

## 7. ANALYSIS OF AVAILABLE OPTIONS: COSTS AND BENEFITS

There are three main options as how to proceed given the present status of the puffin population and other seabirds in the study area (Table 15).

### 7.1 'No Action' Option

If no action is taken and no new species are introduced in the study area, it is expected that whatever effect rats, ferrets, hedgehogs and cats may have on the wildlife and ecosystems will continue. Predation from rats is the most likely responsible for the decrease of puffins and razorbills, and the local extinction of guillemots, over the last century. All the four target invasive predators are likely to have caused reductions in populations of many native species and to be keeping them from increasing in numbers or expanding their ranges.

The present population levels of puffins, razorbills, other seabirds and most of the wildlife found in the study area have seemingly stabilized at very low levels. The best example of this is the puffin population, which stands at 4-6 breeding pairs, a number which has not changed much in the last ten years. This might indicate an adaptive response which sees the surviving individuals being pushed to areas out of reach of invasive predators. These sites out of reach are less suitable habitats with fewer breeding opportunities, which would explain the very small number of breeding pairs found there.

The costs of this option are very high in terms of local biodiversity. The populations of puffins and razorbills are at such low levels they are considered threatened with extinction, and could be completely extirpated by a single random environmental event, such as a storm or a disease, which might not leave enough numbers or genetic variation to survive after that. Furthermore, we do not know what other effects the invasive predators are having on terrestrial wildlife, from breeding songbirds to lizards and small mammals, or what might happen to their populations should the pressure from these predators continue.

Other costs could be the potential increase of invasive predator populations across the study site, the loss of tourism revenue if the puffins become locally extinct, and the negative reaction of the community to the inaction of conservation authorities towards protecting and restoring the natural heritage of Jersey.

The only net benefit of this option is that it has no cost involved. This option also leaves way for the research and development of new techniques for managing wildlife and invasive predators which may become available in the future, such as pathogens, immunocontraception or genetic engineering. These techniques have not been fully developed or field-tested, and may be many years away yet.

As for the relation between invasive predators and natural communities at the study area, it is possible that removing the pressure from invasive predators on some prey species, like the rabbit, might have a cascade effect upon the vegetable community if the rabbit population increases and with it the grazing pressure. However, there are natural predators of rabbits in the area, therefore it would be likely that over time their pressure and density would increase as a response to an increase in rabbit numbers. It is also likely that some intra-guild competition and even predation might be occurring between the invasive predators (such as ferrets or cats killing young rats), and that if only one species was to be removed, other invasive predators might increase. However, it is stated in this proposal that all four species need to be removed at once, hence eliminating the possibility that any of them would benefit from the removal of another.

## 7.2 'Sustained Control' Option

This option would involve ongoing work to control or remove the target invasive predators from the reserve area and buffer zones, effectively the whole of the study area at the very least. The methods involved would likely include a combination of lethal and non-lethal techniques, such as toxins and traps.

Whilst this option might appear to have lower costs than a fence, it is the least cost-effective option on the long term, mainly due to the need for control work at perpetuity. It requires the long-term commitment of financial resources, personnel, equipment and toxic chemicals.

The main aim of this work would be to achieve sufficiently low densities of the invasive predators to result in an increase of survival, range and population of puffins and other threatened native wildlife. These techniques do not have a 100% rate of success and their starting rate of success decreases the longer the techniques are used for.

The negative effects of the control measures on the environment and on native species would also be perpetual, and might increase in time. It is likely that the continued input of toxins in the environment over time will have a negative effect on many non-targeted native species as well as the soil and water table, eventually.

The live and kill traps used for control or removal would also present serious danger to non-target species, and are very costly to maintain, bait and check on a regular basis. It is also likely that the targeted invasive mammals will develop, over time, behavioural and physiological responses to avoid or survive the control methods, such as avoiding traps or becoming resistant to the effects of rodenticide<sup>110</sup>. This will mean that the return per effort of the control measures will diminish over time, and will make future eradications more difficult, longer, and more costly, environmentally and economically.

Furthermore, if control works were to be stopped or scaled down the conservation benefits would be lost, as the invasive predator populations would recover quickly through lack of control and from immigration from adjacent areas. This would mean that all the financial investment, as well as the hours, months and even years of work, as well as the negative impacts that the environment and that non-target species might have suffered, had all been wasted in a pointless exercise.

In regards to the intended beneficiaries of this control work - the puffins and other native species-, the scaling down or cease of control measures would mean not only that their populations would likely decrease back to pre-control levels, but more worryingly it would mean that individuals and breeding pairs which have colonised new areas of the reserve would be put immediately at a much higher risk from predators and would likely be the first casualties of the change of management. The result of endangering native species after having encouraged them to establish in the reserve, would make this option ethically questionable.

Maintaining some form of control and removal of invasive predators at the reserve would have some benefits over the 'Status Quo' option. Ongoing control may be sufficient for slowing the rate of decline of some species, and it can be applied in a flexible manner in response to changes in conditions, budgets, personnel etc. It might be used as a temporary measure to allow for development of new techniques or to plan and implement long-term options, and it usually costs much less than installing a fence and a full eradication – although ongoing costs will be sustained for perpetuity.

### 7.3 'Fence and Eradication' Option

As presented in section 6.1, this option involves installing a predator-exclusion fence and eradicating or removing all four invasive predator species found within its boundaries. A predator-exclusion fence would effectively create a so-called 'mainland island', where wildlife and habitats can be managed on the whole and more effectively. Once the island is created, or in this case the Plémont Seabird Reserve, the eradication of invasive predators from within could proceed.

Eradication is a process widely researched, trialled and perfected in many islands by conservation researchers and managers all over the world, and with such positive results that many habitats have been restored and their species recovered from the brink of extinction, or even re-colonised former ranges where they had become extinct<sup>8,111</sup>. This positive experience suggests that creating a mainland site by constructing and maintaining a pest-proof fence and removing all invasive mammals inside the fenced area could enhance the native species and habitats within the site. This is also supported by evidence from other pest-proof sites around the world<sup>112,113,114</sup>.

The costs and risks of this option have been detailed in Sections 5.3 and 6.5, as well as listed in Table 18. They involve mainly the high financial costs, need for long-term biosecurity, and risks to non-target species on the short term (especially during the eradication phase).

The benefits from creating the reserve cannot be overstated. All the native wildlife within it will be safe from the predation and competition of these non-indigenous invasive predators, and will have a chance to recover, increase and expand their range. All native birds breeding in the reserve will benefit: puffins, razorbills and other seabirds; but also Dartford warblers, stonechats, linnets, meadow pipits and many others. The puffin is an 'Umbrella species', meaning that its protection will also benefit other endangered key species native to Jersey, such as the green lizard, the slow worm and the Jersey bank vole. Being a protected habitat and safe from invasive predators, the reserve can be used as a creche or reservoir of native species which are endangered elsewhere. Across the fence, the land recently restored by the National Trust is a preferred feeding and resting spot for wintering skylarks, stonechats, pied wagtails and wheatears. All of them will benefit from predator-free foraging opportunities inside the reserve.

The improvement of the habitats and the increase of biodiversity and numbers of birds and other wildlife in the reserve will increase the value of the site as a tourism asset, creating new opportunities for travel agencies, outdoor leisure operators, hospitality businesses and local commerce. The creation of the reserve in Jersey is likely to raise the profile of the Island and attract visitors, generating opportunities to showcase Jersey's native species, conservation programmes and the use of new management technologies to visitors and the local community alike.

By engaging local visitors with the outdoors and promoting a variety of physical activities in the area, the reserve will contribute to the mental and physical well-being of the many Islanders. The project would also launch events and activities aimed at children which will improve their educational opportunities and involve them in various community-based aspects of the project. In the future, the local community might become even more involved in the management of the reserve. Many of the pest-proof fenced mainland islands around the world are being developed and managed by community groups, due to the popularity of these projects<sup>115,116</sup>.

Finally, the installation of the fence and eradication work will generate employment of local professionals, and will also offer apprenticeships and training programmes via their contractors. It will also establish working links with local educational institutions, such as Highlands College and the



Jersey International Centre of Advanced Studies, to offer research and internship opportunities to students in STEM and the fields of ecological research, management and conservation.

The Reserve itself will become a case study and management tool as it develops, generating local knowledge which can be transferred to future conservation projects in Jersey and in the other Channel Islands.

Table 18. Costs and benefits of the three main management options.

OPTION	COSTS	BENEFITS
<b>No Action</b>	<ul style="list-style-type: none"> <li>• Continued presence of invasive predators and their effects on native species</li> <li>• Potential extinction of native species with numbers below critical</li> </ul>	<ul style="list-style-type: none"> <li>• No economical cost</li> <li>• Research opportunities to monitor long-term effect of invasive predators</li> <li>• Gives more time for new techniques to be researched and developed to manage invasive predators</li> </ul>
<b>Sustained Control</b>	<ul style="list-style-type: none"> <li>• Continued presence (in low numbers) of invasive predators and their effects on native species</li> <li>• Ongoing financial costs of labour and materials (baiting, trapping and other methods)</li> <li>• Risk of alien species developing avoidance or physical resistance to control methods</li> <li>• Benefits will decrease over time if invasive species adapt</li> <li>• Will hinder future eradications if invasive species adapt or develop bait resistance</li> <li>• Benefits will cease if ongoing control stops and endangered populations might revert to before-control levels</li> <li>• Continued re-invasion from adjacent areas</li> <li>• Ongoing use of poison and long-term effects in ecosystem</li> <li>• Increase risk to non-target species being affected by control methods</li> <li>• Difficulty to reduce invasive predators to satisfactory levels</li> <li>• Cannot prevent dispersal of native species into unprotected area</li> <li>• Potential extinction of native species with numbers below critical</li> </ul>	<ul style="list-style-type: none"> <li>• Lower numbers of invasive predators within the reserve area</li> <li>• Some native species of flora and fauna will benefit to some degree, might increase breeding success and expand populations</li> <li>• Research opportunities to monitor ecological changes and adaptations to long-term control</li> </ul>

OPTION	COSTS	BENEFITS
<b>Fence and Eradication</b>	<ul style="list-style-type: none"> <li>Restricting the range of some native species inside the reserve</li> <li>High initial costs to install fence and to eradicate all invasive predators</li> <li>Ongoing costs of fence maintenance and biosecurity measures</li> <li>Refurbishment costs of fence every 25-30 years (predicted lifespan)</li> <li>Might not exclude new invasive species not present in Jersey at the time of this assessment</li> <li>Use of poison or other control methods during eradication phase will present a risk to non-target native species</li> <li>Cost of development of structures, education and awareness services for the general public to use</li> <li>Public access to area will be by fence gates</li> </ul>	<ul style="list-style-type: none"> <li>Most cost-effective option in the long term</li> <li>Permanent removal of invasive predators and immediate cease of their harmful effects on native wildlife</li> <li>Barrier to target invasive predators</li> <li>Enhanced survival, breeding, increase and expansion on Puffin population as well as many other native species</li> <li>Potential re-colonisation of previously extinct native wildlife</li> <li>Potential colonisation of other wildlife native to the English Channel and British Isles</li> <li>Regeneration of vegetation and natural communities</li> <li>Spill-over effect of some native species outside the fence as numbers increase</li> <li>Track alongside the fence and regular maintenance serves as firebreak, protecting biodiversity inside from fire</li> <li>Enhanced used of area by locals and tourists</li> <li>Enhanced tourist revenue for leisure operators and service industries</li> <li>Increased awareness of conservation and invasive species issues</li> <li>Increase in public and political interest in project, with sense of pride in protection of natural heritage</li> <li>Research opportunities to monitor habitat and population changes after eradication</li> <li>Positive impact on similar projects in other Channel Islands and Britain</li> </ul>

## 8. CONCLUSIONS

The proposed Plémont Seabird Reserve site is suitable for the construction of a predator-exclusion fence. The feasibility of this part of the project is moderate for the total length of the fence, but good when considering the different layout options proposed, as well as the potential flexibility of a staggered approach of fencing / eradicating by sections. All the technical requirements can be met with existing methodologies, but will require the involvement of predator fence experts in the final design and the input of local contractors. Water management, visual impact and public access points will be an important aspect of the final design.

The eradication or removal of all invasive predators recorded at the site is feasible by current technology and methodologies. The feasibility of preventing and managing re-invasions is moderate to good, depending on the robustness of the protocols implemented and the strength of the long-term commitments from the agencies involved.

The project is likely to gather enough support from the local community to enable it to succeed, however it is possible that the main concerns amongst the community revolve around the visual impact of the fence as well as the eradication methods and outcomes for the removed or excluded target animals. Community engagement during the consultation process and awareness campaigns will be paramount to ensure that the process is socially acceptable and manages the expectations of the local community adequately.

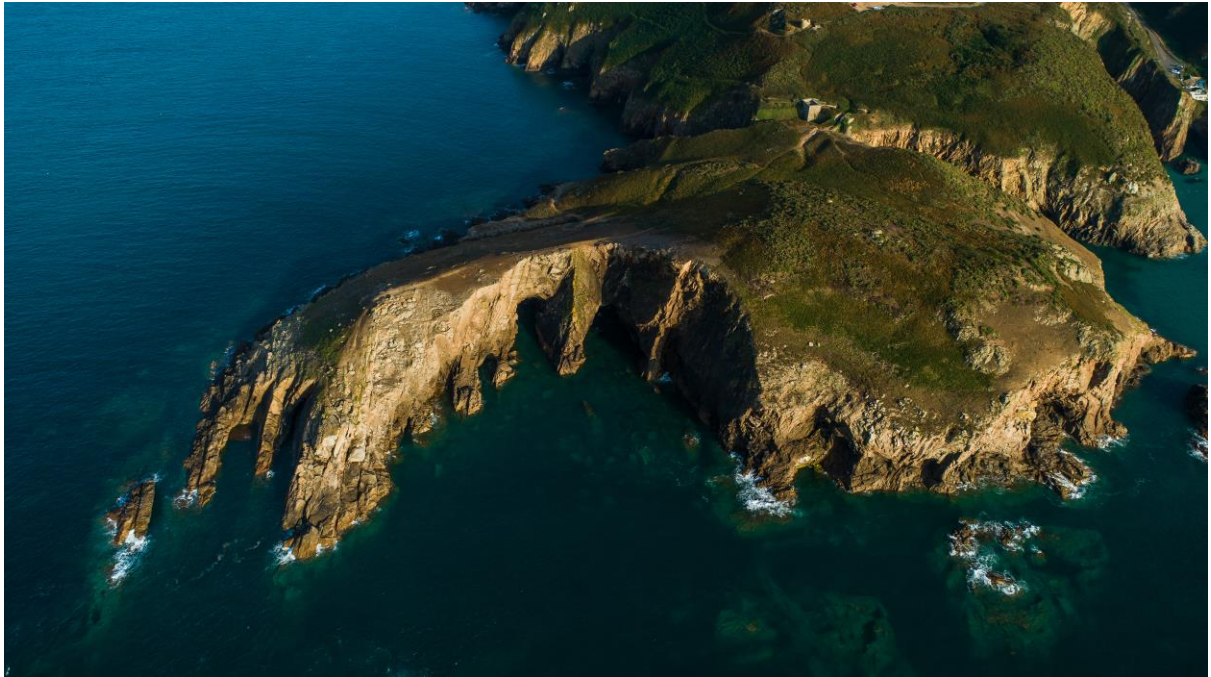
This assessment also states that the 'sustained control' option could only be considered acceptable in the very short-term. This option is more costly over time, its financial security cannot be guaranteed to perpetuity, has a cumulative negative impact on the environment and potentially on many non-target native species. It can hinder future eradications and, should the control cease or become relaxed, the impact to native species in the process of recovery would be negative in many fronts, such as the decline or extinction of the puffins, the loss of public confidence, and the ethical implications of such perceived waste of resources, funding, work and the lives of wild animals. This option would only be recommended in order to gain time whilst researching, planning and installing a predator-exclusion fence and committing to a full eradication.

The 'fence and eradication' is a high cost and high complexity option, but its potential reward is very high, and will likely result in higher, more effective conservation outcomes than the other options. The fence provides the chance to completely remove all the invasive predators from the reserve and to prevent them from re-invading, creating a safe and suitable habitat where the puffins and other endangered species can recover. The reserve created by the fence could also act as a source or nursery ground for other native species that can then disperse from the site to repopulate neighboring habitats and boost Island populations. It is likely that the reserve will become a tourist attraction, creating economic opportunities in the service and leisure industry, and it will also provide skill-building opportunities for many local professionals and students of related disciplines.

This feasibility study has found no obvious impediment or issue that would automatically preclude the successful construction of a predator-proof fence and eradication of all invasive predators from within the reserve. It also concludes that the environmental benefits will greatly outweigh the costs. In contrast to its many ecological and social benefits, this is a very expensive project, and it requires the long-term commitment of the BOTE partnership, the Jersey Government and its partner agencies. Maintenance of the fence, biosecurity measures and community engagement work will need to continue indefinitely to ensure the full benefits of this project.

It is recommended that:

1. Pending funding to create the Plémont Seabird Reserve, the consultation process continues with an education and advocacy programme to engage the local community in regards to the fence proposal. This should include the proposed design, location options, and the importance of protecting the puffins and endangered native species.
2. A socio-economic study is undertaken to assess the impacts of the predator-exclusion fence on the local industry and well-being of Jersey's community. This should also assess whether the selected site would be appropriate for certain community and tourist activities.
3. A fully-costed and detailed full-scale fence operational plan is developed by a pest-proof fencing specialist. The plan should include the input of local ecologists as well as contractors, in particular in regards to their experience on the expected longevity of different building materials.
4. A fully-costed eradication plan is developed for the target species within the fence site. This should include a comprehensive risk assessment to non-target species and habitats.
5. A fully-costed biosecurity plan is produced for the target species within the fence site. This should include a comprehensive risk assessment to non-target species and habitats.
6. A business plan is prepared covering all aspects of the project (including location, fence design, fence construction costs, eradication costs, maintenance costs, agency involvement, funding opportunities, predicted revenue and benefits to the endangered species). Options for financing the pest-proof fence project should be investigated (e.g. government funding, international grants, collaborative business partnerships).
7. Information and educational material is produced (i.e. leaflets, a presentation, a short video, posters) to educate decision-makers and other stakeholders about predator-exclusion fences, mainland islands, and the risks and benefits they provide.
8. Options for managing the site (government, conservation agencies, private business or community) are investigated, bearing in mind that the agency that constructs the fence does not necessarily have to be the same body that manages it thereafter.
9. A questionnaire survey is carried out among tourists and the tourism industry to assess the level of interest in establishing tours to the seabird reserve.
10. Ecological monitoring carries on during the planning process and the construction of the fence, and continues afterwards in the form of long-term monitoring schemes. This can be complemented or managed as part of ongoing collaborations with local research agencies and educational institutions.



## REFERENCES

1. Sellarés, C. 2020. 2019-2020 Winter Bird Crops Report. Birds On The Edge, Jersey.
2. Sellarés, C. 2020. Birds On The Edge 2017-2020 Hedge Restoration Scheme. Final Report.
3. Birds On The Edge: Event Archives <http://www.birdsontheedge.org/category/events/>
4. Sellarés, C. 2020 A shire for Jersey's Puffins <http://www.birdsontheedge.org/2020/03/16/a-shire-for-jerseys-puffins/>
5. Atkinson, I., 1985. The spread of commensal species of *Rattus* to oceanic islands and their effects on island avifaunas, In Conservation of Island Birds. ed. P.J. Moors, pp. 35-81. International Council for Bird Preservation, Cambridge, UK. Courchamp et al 2003.
6. Long, J.L. 2003. Introduced Mammals of the World: Their History, Distribution and Influence. CABI, Wallingford, UK.
7. Towns, D., Atkinson, I., Daugherty, C., 2006. Have the harmful effects of introduced rats on islands been exaggerated? *Biological Invasions* 8, 863-891.
8. Jones, H.P., Tershy, B.R., Zavaleta, E.S., Croll, D.A., Keitt, B.S., Finklestein, M.E., Howald, G.R., 2008. Severity of the effects of invasive rats on seabirds: A global review. *Conservation Biology* 22, 16-26.
9. Barun, A. Hanson, C.C. Campbell, K.J. & Simberloff, D. 2011. A review of small Indian mongoose management and eradications on islands. In *Island Invasives: Eradication and Management*. Proceedings of the International Conference on Island Invasives (eds C.R. Veitch, M.N. Clout & D.R. Towns), pp. 17–25. IUCN, Gland, Switzerland and Auckland, New Zealand.
10. The Pacific Invasives Initiative Resource Kit for Rodent and Cat Eradication [http://www.pacificinvasivesinitiative.org/rce/project/2\\_Feasibility\\_Study.html](http://www.pacificinvasivesinitiative.org/rce/project/2_Feasibility_Study.html)
11. 1987 IUCN Position Statement on Translocation of Living Organisms.
12. JNCC 1990 Handbook for Phase 1 Habitat Survey.
13. Statistics Jersey, Jersey Resident Population 2019 Estimate.  
[REDACTED]
14. Jersey National Park factsheet and map <https://jerseynationalpark.com/wheres-park/>
15. Fiona Fyfe Associates. 2020. Jersey Integrated Landscape and Seascape Character Assessment (ILSCA) Produced by the Strategic Policy, Performance and Population (Strategic Policy, Planning and Performance).
16. PAA, 1998. Jersey Phase I survey, Penny Anderson Associates. Unpublished report to Jersey Island Development Committee.
17. Seabird Protection Zone leaflet [REDACTED]  
[REDACTED]
18. Planning and Environment Department. Supplementary Planning Guidance. Archaeology and Planning. Policy Note 1. February 2008. Schedule of sites of archaeological interest.
19. Jersey Heritage factsheet [REDACTED]
20. Prehistoric Jersey Net Le [REDACTED]
21. Prehistoric Jersey Net La Roue de la Croute [REDACTED]
22. Walsh, P.M., Halley, D.J., Harris, M.P., del Nevo, A., Sim, I.M.W. & Tasker, M.L. 1995. Seabird monitoring handbook for Britain and Ireland. JNCC / RSPB / ITE / Seabird Group, Peterborough.

23. N. Ratcliffe, D. Vaughan, C. Whyte and M. Shepherd. 1998. Development of playback census methods for Storm Petrels *Hydrobates pelagicus*, *Bird Study*, 45:3, 302-312.
24. Shonfield, J., and E. M. Bayne. 2017. Autonomous recording units in avian ecological research: current use and future applications. *Avia Conservation and Ecology* 12(1):14.
25. Buxton, R.T. and Jones, I.L. 2012. Measuring nocturnal seabird activity and status using acoustic recording devices: applications for island restoration. *Journal of Field Ornithology*, 83:1.
26. BTO/JNCC/RSPB Breeding Bird Survey Instructions  
[REDACTED]
27. Mitchell, B. and Balogh, S. 2007. Monitoring techniques for Vertebrate pests. Bureau of rural Sciences, Canberra.
28. Barrio, I.C., Pelayo, A. and Tortosa, F. 2009. Assessment of methods for estimating wild rabbit population abundance in agricultural landscapes. *European Journal of Wildlife Research*. 56. 335-340.
29. Swinnerton K.J., Young, H.G. and Sangan, P. 2018. Options for the recovery of nesting seabirds on Jersey, Channel Islands. Report to the Durrell Wildlife Conservation Trust.  
[REDACTED]  
[REDACTED]
31. JNCC. 2020. Seabird Population Trends and Causes of Change: 1986-2018 Report (<https://jncc.gov.uk/our-work/smp-report-1986-2018>) Joint Nature Conservation Committee. Updated 10 March 2020.
32. Sangan, P. 2017. Seabird monitoring around Plémont. Sangan Island Conservation Ltd. Report to the National Trust for Jersey.
33. Young, H.G., Dryden, M. and Pinel, J. 2011. Conservation Status of Jersey's Birds: Jersey's Bird Populations in the 21st Century. Durrell Wildlife Conservation Trust, Jersey.
34. Liliana, Mendoza, C. and Pérez, J. 2003. Two new breeding localities for the Wedge-rumped Storm-Petrel *Oceanodroma tethys kelsalli* in Peru. *Marine Ornithology*.
35. Ed. From correspondent. 1977. "Rare seabirds die in attack". *Jersey Evening Post*, 1 Sept 1977.
36. States of Jersey 2014. Island Plan 2011: Revised (2014).
37. Government of Jersey SSIs search [REDACTED]
38. PAA, 2010. Protection of Ecologically Sensitive Areas (PESAP) Project. Jersey Environment.
39. Parkes, J. 2017. Plémont. How to create a species rich grassland. National Trust for Jersey Discover Magazine. Autumn 2017.
40. Young, G. and Young, R.P. 2012 Jersey Farmland Birds Monitoring Manual Version 1.1. Durrell Wildlife Conservation Trust.
41. Conservation of Wildlife (Jersey) Law 2000. Revised Edition 22.450 Showing the law as at 1 January 2017.
42. States of Jersey Planning and Environment Committee 2000 Biodiversity Action Plans for Jersey.
43. States of Jersey Planning and Environment Committee 2000 Biodiversity A Strategy for Jersey.
44. Mack, R., Simberloff, D., Lonsdale, W.M., Evans, H., Clout, M. and Bazzaz, F. 2000. Biotic invasions: Causes, epidemiology, global consequences and control. *Issu. Ecol.* 5. 2-20.
45. Simberloff, D. 2003. Confronting introduced species: a form of xenophobia?. *Biological Invasions* 5, 179–192 (2003). [REDACTED]
46. Paulay G. 1994. Biodiversity on oceanic islands: its origin and extinction. *Am Zool* 34:134–144.

47. Burger J, Gochfeld M, 1994. Predation and effects of humans on island-nesting seabirds. In: Nettleship DN, Burger J, Gochfeld M ed. *Seabirds on Islands: Threats, Case Studies, and Action Plans*. Cambridge, U.K.: BirdLife International (BirdLife Conservation Series no. 1), 39–67.
48. Cole, N. & Jones, C. & Harris, S. 2005. The need for enemy-free space: The impact of an invasive gecko on island endemics. *Biological Conservation*. 125. 467-474. 10.1016/j.biocon.2005.04.017.
49. King WB. 1985. *Island birds: Will the future repeat the past?* In: ICBP, editor. *Conservation of island birds*. Cambridge: International Council for Bird Preservation. 3–15. King 1985
50. Manchester, S. J. and Bullock, J.M. 2000. The impacts of non-native species on UK biodiversity and the effectiveness of control. *J. Appl. Ecol.* 37, 845-864.
51. Innes, J. 2001. Advances in New Zealand mammalogy 1990–2000: European rats, *Journal of the Royal Society of New Zealand*, 31:1, 111-125, DOI: 10.1080/03014223.2001.9517642
52. Russell, J., Towns, D., Anderson, S. 2005. Intercepting the first rat ashore. *Nature* 437, 1107. (Wittenberg, R. (ed.) 2005).
53. Russell, J. C., Towns, D.R. and Clout, M. N. 2008. Review of rat invasion biology. Implications for island biosecurity. *Science for Conservation* 286. Department of Conservation, Wellington New Zealand.
54. Imber, M. and Harrison, M. 2000. Interactions between petrels, rats and rabbits on Whale Island, and effects of rat and rabbit eradication. *New Zealand Journal of Ecology*. 24. M.
55. Song, Y., Lan, Z., Kohn, M.H. 2014. Mitochondrial DNA Phylogeography of the Norway Rat. *PLoS ONE* 9(2): e88425.
56. CABI.org datasheet [redacted] Accessed February 2021.
57. Kurle, C. M., Croll, D. A., Tershy B. R. 2008. Introduced rats indirectly change marine rocky intertidal communities from algae- to invertebrate-dominated. *Proceedings of the National Academy of Sciences* 105 (10) 3800-3804; DOI: 10.1073/pnas.0800570105.
58. Jones, C., and Sanders, M.D. 2005. European hedgehog. In 'The Handbook of New Zealand Mammals'. (Ed. C. M. King.) pp. 81–94. (Oxford University Press: Melbourne.)
59. Department of Conservation, 2005. *Animal Pests: Ferrets*. Wellington, New Zealand: Department of Conservation.
60. Jackson, D.B. and Green, R.E. 2000. The importance of the introduced hedgehog (*Erinaceus europaeus*) as a predator of the eggs of waders (*Charadrii*) on the machair in South Uist, Scotland. *Biological Conservation*, 93, 333–348.
61. Jackson, D. 2001. Experimental Removal of Introduced Hedgehogs Improves Wader Nest Success in the Western Isles, Scotland. *Journal of Applied Ecology*, 38(4), 802-812.
62. Corbett GB, Ovenden D, 1980. *The Mammals of Britain and Europe*. London, UK: Collins, 247 pp.
63. Howes CA, 1980. Aspects of the history and distribution of polecats and ferrets in Yorkshire and adjacent areas. *Naturalist*, 105:3-16.
64. Kowalski K, 1976. *Mammals, an outline of theriology*. Warsaw, Poland: Panstwowe Wydawnictwo, Naukowe.
65. Jurek RM, 1998. A review of national and California population estimates of pet ferrets. Sacramento, California, USA: Department of Fish and Game, Wildlife Management Division, 11 pp. [Bird and Mammal Conservation Program Report, 98-09.]
66. Wenker C, Christen C, 2002. Ferrets in veterinary practice. (Frettchen in der Tierarztpraxis.) *SAT, Schweizer Archiv für Tierheilkunde*, 144(11):575-584.
67. Duda J, 2003. *Mustela putorius furo: domestic ferret*. Animal Diversity Web. Ann Arbor, Michigan, USA: University of Michigan Museum of Zoology.



68. Landcare Research, 2008. Ferret and stoat research. Lincoln, New Zealand: Landcare Research.
69. Ragg J, 1998. The denning behaviour of feral ferrets (*Mustela furo*) in a pastoral habitat, South Island, New Zealand. *Journal of Zoology*, 246(4):443-486. [Special section: Communications from the Mammal Society - No. 77.]
70. Jeffares R, 1986. The feral ferret in New Zealand. *New Zealand Wildlife*, 10:43-46.
71. Department of Conservation, 2005. Animal Pests: Ferrets. Wellington, New Zealand: Department of Conservation.
72. King, C.M. (ed.) 1990. *The Handbook of New Zealand Mammals*. Oxford University Press. Auckland, New Zealand.
73. Clapperton, B.K. and Day, T.D. 2001. Cost-effectiveness of exclusion fencing for stoat and other pest control compared with conventional control. DOC Science Internal Series 14. Department of Conservation, Wellington, New Zealand.
74. Pitta Groz, M., Pereira, J.C., Silva, A.G. 2002. Invasive alien species as the main threat to Azores seabird populations. In: *Proceedings of Workshop on invasive alien species on European islands and evolutionary isolated ecosystems*, Horta, Azores, Portugal, 10 October 2002.
75. Lever C, 1985. *Naturalized mammals of the world*. London, UK: Longman, xvii + 487 pp.
76. Rive, R. 2009. "Ferret threat to our indigenous bird and animal populations". Letter published in the *Jersey Evening Post* Oct 15, 2009.
77. Brickner, I. 2003. The impact of domestic cat (*Felis catus*) on wildlife welfare and conservation: a literature review. With a situation summary from Israel.
78. Skira, I. and Gr Copson. "Biology of the Feral Cat, *Felis Catus* (L.), On Macquarie Island." *Wildlife Research* 12 (1985): 425-436.
79. Fitzgerald, B.M., Karl, B.J. and Veitch, C.R. 1991: The diet of feral cats (*Felis catus*) on Raoul Island, Kermadec group. - *New Zealand Journal of Ecology* 15: 123–129.
80. Dickman, C.R. 1996: Overview of the impact of feral cats on Australian native fauna. - Australian Nature Conservation Agency, Canberra, 92 pp.
81. P. Roberts, W.A. Seabrook. 1989. **A relationship between black rats (*Rattus rattus*), Seychelles fruit bats (*Pteropus sechellensis aldabrensis*) and the coccid (*Icerya sechellarum*) (Insecta, Homoptera) on Aldabra Atoll Seychelles.** *J. Zool.*, 218, pp. 332-334.
82. Lowe S. J., Browne M. and Boudjelas S. 2000. *100 of the World's Worst Invasive Alien Species*. Published by the IUCN/SSC Invasive Species Specialist Group (ISSG), Auckland, New Zealand.
83. Medina, F.M., Bonnaud, E., Vidal, E., Tershy, B.R., Zavaleta, E.S., Josh Donlan, C., Keitt, B.S., Le Corre, M., Horwath, S.V. and Nogales, M. 2011. A global review of the impacts of invasive cats on island endangered vertebrates. *Glob Change Biol*, 17: 3503-3510. <https://doi.org/10.1111/j.1365-2486.2011.02464.x>
84. Van Aarde. R.J. 1980. The diet and feeding behaviour of feral cats, *Felis catus* on Marion Island. *South Afr. Wildl. Res.*, 10:123-128.
85. Pascal M. 1980. Structure et dynamique de la population de chatsharets de l'archipel des Kerguelen. *Mammalia* 44:161–182
86. Towns, D. R., C. H. Daugherty and I. A. E. Atkinson (editors). 1989. *Ecological restoration of New Zealand islands: Papers presented at conference on ecological restoration of New Zealand islands*, University of Auckland, 20-24 November 1989, Auckland, New Zealand (Conservation sciences publication No. 2). Publisher: Wellington, Department of Conservation.
87. Veitch, C.R., Clout, M.N., Towns, D.R. 2011. *Island Invasives: Eradication and Management*. Proceedings of the International Conference on Island Invasives. Gland, Switzerland: IUCN and Auckland, New Zealand: CBB. xii + 542pp.

88. Aguirre-Muñoz, A. 2008. High-impact conservation: invasive mammal eradications from the islands of western Mexico. *Ambio* (2008): 101-107.
89. US Fish & Wildlife Service [https://www.fws.gov/refuge/Midway\\_Atoll/wildlife\\_and\\_habitat/Bonin\\_Petrel.html](https://www.fws.gov/refuge/Midway_Atoll/wildlife_and_habitat/Bonin_Petrel.html)
90. N.D. Holmes , B.S. Keitt<sup>1</sup> , D.R. Spatz<sup>1,2</sup>, D.J. Will<sup>1</sup> , S. Hein<sup>1</sup> , J.C. Russell<sup>3</sup> , P. Genovesi<sup>4</sup> , P.E. Cowan<sup>5</sup> and B.R. Tershy. 2019. Tracking invasive species eradications on islands at a global scale. In: C.R. Veitch, M.N. Clout, A.R. Martin, J.C. Russell and C.J. West (eds.) (2019). *Island invasives: scaling up to meet the challenge*, pp. 628–632. Occasional Paper SSC no. 62. Gland, Switzerland: IUCN.
91. Clapperton, B. and Day, T. 2001. Cost-effectiveness of exclusion fencing for stoat and other pest control compared to conventional control.
92. Donlan et al. 2003 Donlan, C. J., G. R. Howald, B. R. Tershy, and D. A. Croll. 2003a. Evaluating alternative rodenticides for island conservation: roof rat eradication from the San Jorge Islands, Mexico. *Biological Conservation* 14:29–34.
93. Ebbert, S. and Byrd, G. 2002. Eradications of invasive species to restore natural biological diversity on Alaska Maritime National Wildlife Refuge.
94. Lorvelec, O. and Pascal, M. 2005. French attempts to eradicate non-indigenous mammals and their consequences for native biota. *Biological Invasions*, 7, 135–140.
95. Cromarty, P.L., Broome, K.G., Cox, A., Empson, R.A., Hutchinson, W.M. and McFadden, I. 2002. Eradication planning for invasive alien animal species on islands— the approach developed by the New Zealand Department of Conservation. In: *Turning the tide: The Eradication of Invasive Species*. Veitch, C.R. and Clout, M.N. (eds). IUCN
96. Towns, D.R. and Broome, K.G. 2003. From small Maria to massive Campbell: forty years of rat eradications from New Zealand islands. *New Zealand Journal of Zoology*, 30, 377-398.
97. Russell, J.C., Towns, D.R. and Clout, M.N. 2008. Review of rat invasion biology: implication for island biosecurity. *Science for Conservation* 286. Department of Conservation, Wellington
98. Spurr E.B., Maitland, M.J., Taylor, G.E., Wright, G.R.G., Radford, C.D., Brown, L.E. 2005. Residues of brodifacoum and other anticoagulant pesticides in target and non-target species, Nelson Lakes National Park. *New Zealand Journal of Zoology* 32: 237-249.
99. Spurr, E.B., Ogilvie, S.C., Morse, C.W. and Young, J.B. 2005. Development of a toxic bait for control of ferrets (*Mustela furo*) in New Zealand, *New Zealand Journal of Zoology*, 32:2, 127-136.
100. Orkney Native Wildlife Project <https://www.onwp.org/>
101. Price, C.J., Banks, P.B., Brown, S., Latham, C., Latham, D.M., Pech, R.P., Norbury, G.L. 2020. Invasive mammalian predators habituate to and generalize avian prey cues: a mechanism for conserving native prey. *Ecological Applications*, 2020; DOI: 10.1002/eap.2200
102. Norbury, G. 2001: Conserving dryland lizards by reducing predator-mediated apparent competition and direct competition with introduced rabbits. *Journal of Applied Ecology* 38: 1350-1361.
103. Parkes, J. and Murphy, E. 2003. Management of introduced mammals in New Zealand. *New Zealand Journal of Zoology*, 30, 335–359.
104. Ragg, J.R. and Clapperton, B.K. 2004. *Ferret Control Manual*. (Prepared for: Animal Health Board, Wellington). [Accessed 18 March 2007, from: [REDACTED]
105. Woods, M., McDonald, R.A. and Harris, S. 2003. Predation of wildlife by domestic cats *Felis catus* in Great Britain. *Mammal Review*, 33: 174-188. [REDACTED]
106. Global Invasive Species Database [www.issg.org](http://www.issg.org)
107. Rotokare Scenic Reserve Trust [www.rotokare.org.nz](http://www.rotokare.org.nz)

108. Government of Jersey Common Strategic Policy 2018-2022

109. Jersey Legal Information Board: Treaties and Conventions

110. Alan Buckle, A., Jones, C., Talavera, M. and Prescott, C. 2020. Anticoagulant Resistance in Rats and Mice in the UK - Summary Report with new data for 2019-20. Report from the Campaign for Responsible Rodenticide Use (CRRU) UK for the Government Oversight Group. Vertebrate pests Unit University of Reading

111. Miskelly, C.M., Dowding, J.E., Elliott, G.P., Hitchmough, R.A., Powlesland, R.P., Robertson, H.A., Sagar, P.M., Scofield, R.P., Taylor, G.A., 2008. Conservation status of New Zealand birds, 2008. *Notornis* 55, 117–135.

112. Connolly, T.A.; Day, T.D.; King, C.M. 2009: Estimating the potential for reinvasion by mammalian pests through pest-exclusion fencing. *Wildlife Research* 36: 410–421.

113. Innes J, Saunders A 2011. Eradicating multiple pests: an overview. In Veitch CR, Clout MN, Towns DR eds *Island invasives: eradication and management: proceedings of the International Conference on Island Invasives*. Gland, Switzerland, IUCN. Pp. 177–181.

114. Burns, B.; Innes, J.; Day, T. 2012: The use and potential of pest-proof fencing for ecosystem restoration and fauna conservation in New Zealand. Pp. 65–90 in Somers, M.J.; Hayward, M.W. (Eds): *Fencing for conservation: restriction of evolutionary potential or a riposte to threatening processes?*, DOI 10.1007/978-1-4614-0902-1\_5, © Springer Science+Business Media, LLC 2012.

115. Campbell-Hunt, D.M., Freeman, C. and Dickinson, K.J.M. 2010. Community-based entrepreneurship and wildlife sanctuaries; case studies from New Zealand. *International J. of Innovation and Regional Development*, 2, 4–21.

116. Innes, J.; Lee, W.G.; Burns, B.; Campbell-Hunt, C.; Watts, C.; Phipps, H.; Stephens, T. 2012: Role of predator-proof fences in restoring New Zealand's biodiversity: a response to Scofield et al. (2011). *New Zealand Journal of Ecology* 36(2): 232–238. Innes et al. 2012). Corbett GB, Southern HN, 1977. *The Handbook of British Mammals*. London, UK: Blackwell.

## APPENDIX I. BIRDS OBSERVED OR ASSOCIATED TO THE STUDY AREA AND CONSERVATION STATUS

Common Name	Current Scientific Name	Conservation Status Jersey	Working List of Channel Island Birds	Red Data Categories (non-IUCN)	International Designations
Sparrowhawk	<i>Accipiter nisus</i>	Green	Scarce resident		
Common Sandpiper	<i>Actitis hypoleucos</i>			Amber	
Skylark	<i>Alauda arvensis</i>	Red	Scarce resident	Red	Jersey Biodiversity Action Plan
Razorbill	<i>Alca torda</i>	Red	Common migrant	Amber	Critically Endangered (France)
Mallard	<i>Anas platyrhynchos</i>	Green	Common resident	Amber	
Rock Pipit	<i>Anthus petrosus</i>	Green	Common resident		
Meadow pipit	<i>Anthus pratensis</i>	Amber	Common resident	Amber	
Swift	<i>Apus apus</i>	Amber	Common visitor	Amber	
Grey Heron	<i>Ardea cinerea</i>	Green	Common visitor		
Short-Eared Owl	<i>Asio flammeus</i>		Rare visitor	Amber	Vulnerable (France)
Long-Eared Owl	<i>Asio otus</i>	Amber	Rare breeder		
Common Buzzard	<i>Buteo buteo</i>	Amber	Rare resident		
Eurasian Linnet	<i>Carduelis cannabina</i>	Amber	Common breeder		
European Goldfinch	<i>Carduelis carduelis</i>	Green	Common resident		
Greenfinch	<i>Carduelis chloris</i>	Green	Common resident	Green	
Short-toed Treecreeper	<i>Certhia brachydactyla</i>	Amber	Common resident	Amber	
Ringed Plover	<i>Charadrius hiaticula</i>	Red	Rare resident	Red	
Marsh Harrier	<i>Circus aeruginosus</i>	Amber	Scarce resident	Amber	
Rock Dove	<i>Columba livia</i>		Common resident		Endangered (France)
Stock Pigeon	<i>Columba oenas</i>	Green	Common resident	Amber	
Woodpigeon	<i>Columba palumbus</i>		Common resident		
Northern Raven	<i>Corvus corax</i>	Amber	Rare resident		
Western Jackdaw	<i>Corvus monedula</i>	Green	Rare resident		
Carrion Crow	<i>Corvus corone</i>		Common resident		
Blue Tit	<i>Paridae</i>	Green	Common resident		
House Martin	<i>Delichon urbicum</i>	Amber	Common breeder	Amber	
Great Spotted Woodpecker	<i>Dendrocopos major</i>	Green	Common resident		
Common Reed Bunting	<i>Emberiza schoeniclus</i>		Common visitor	Amber	Priority Species (UK)
European Robin	<i>Erithacus rubecula</i>	Green	Common resident		
Peregrine Falcon	<i>Falco peregrinus</i>	Amber	Rare resident		
Common Kestrel	<i>Falco tinnunculus</i>	Amber	Common resident	Amber	
Atlantic Puffin	<i>Fratercula arctica</i>	Red	Rare migrant	Amber	Jersey Biodiversity Action Plan; Critically Endangered (France)
Chaffinch	<i>Fringilla coelebs</i>	Green	Common resident		
Brambling	<i>Fringilla montifringilla</i>	Green	Scarce visitor		
Northern Fulmar	<i>Fulmarus glacialis</i>	Green	Common resident	Amber	
Common Snipe	<i>Gallinago gallinago</i>	Amber	Common visitor	Amber	Endangered (France)
Eurasian Jay	<i>Garrulus glandarius</i>	Green	Common resident		
Oystercatcher	<i>Haematopus ostralegus</i>	Green	Common resident	Amber	
Barn Swallow	<i>Hirundo rustica</i>	Green	Common breeder	Amber	
European Storm-petrel	<i>Hydrobates pelagicus</i>	Amber	Scarce visitor	Amber	Near Threatened (France)
Herring gull	<i>Larus argentatus</i>	Amber	Common resident	Red	

Lesser Black-backed Gull	<i>Larus fuscus</i>	Green	Common resident	Amber	
Great Black-Backed Gull	<i>Larus marinus</i>	Green	Common resident	Amber	
Pied wagtail	<i>Motacillidae</i>	Amber	Common resident		
Grey Wagtail	<i>Motacilla cinerea</i>	Green	Occasional breeder	Amber	
Eurasian Curlew	<i>Numenius arquata</i>	Eed	Common visitor	Amber	Near Threatened (France)
Northern Wheatear	<i>Oenanthe oenanthe</i>	Red	Common visitor	Amber	Near Threatened (France)
Great Tit	<i>Parus major</i>	Green	Common resident		
House Sparrow	<i>Passer domesticus</i>	Amber	Common resident	Red	
European Shag	<i>Phalacrocorax aristotelis</i>	Red	Common resident	Amber	Jersey Biodiversity Action Plan
Great Cormorant	<i>Phalacrocorax carbo</i>	Red	Common resident		
Common Pheasant	<i>Phasianus colchicus</i>	Green	Abundant resident		
Common Redstart	<i>Phoenicurus phoenicurus</i>		Scarce visitor	Amber	
Common Chiffchaff	<i>Phylloscopus collybita</i>	Green	Common resident		
Willow warbler	<i>Phylloscopus trochilus</i>	Red	Common visitor	Amber	
Eurasian Magpie	<i>Pica pica</i>		Common resident		
Snow Bunting	<i>Plectrophenax nivalis</i>		Rare visitor	Amber	
Eurasian Golden Plover	<i>Pluvialis apricaria</i>	Amber	Common visitor	Amber	
Duncock	<i>Prunella modularis</i>	Green	Common resident	Amber	Vulnerable (France)
Manx Shearwater	<i>Puffinus puffinus</i>		Common visitor	Amber	
Balearic Shearwater	<i>Puffinus mauretanicus</i>	Red	Scarce visitor	Red	Critically endangered (Global Red List) Priority species (UK) Vulnerable (France)
Red-billed Cough	<i>Pyrhacorax pyrrhocorax</i>		Former resident	Amber	
Eurasian Bullfinch	<i>Pyrrhula pyrrhula</i>	Red	Scarce resident	Amber	
Firecrest	<i>Regulus ignicapillus</i>		Common visitor	Amber	
Goldcrest	<i>Regulus regulus</i>	Amber	Common resident		
Sand Martin	<i>Riparia riparia</i>	Amber	Common breeder	Amber	
Common Stonechat	<i>Saxicola torquatus</i>	Red	Rare resident		Jersey Biodiversity Action Plan
Woodcock	<i>Scolopax rusticola</i>	Amber	Common visitor	Amber	
Collared Dove	<i>Streptopelia decaocto</i>	Green	Common resident		
Common Starling	<i>Sturnus vulgaris</i>	Red	Common resident	Red	
Eurasian Blackcap	<i>Sylvia atricapilla</i>	Green	Common resident		
Common Whitethroat	<i>Sylvia communis</i>	Green	Common breeder	Amber	Near Threatened (France)
Dartford Warbler	<i>Sylvia undata</i>	Green	Scarce resident	Amber	Jersey Biodiversity Action Plan
Wren	<i>Troglodytes troglodytes</i>	Green	Common resident		
Redwing	<i>Turdus iliacus</i>	Green	Common visitor	Red	
Common Blackbird	<i>Turdus merula</i>	Green	Common resident		
Song thrush	<i>Turdus philomelos</i>	Amber	Common resident	Red	
Barn Owl	<i>Tyto alba</i>	Green	Scarce resident	Amber	
Guillemot	<i>Uria aalge</i>	Amber	Scarce visitor	Amber	
Northern Lapwing	<i>Vanellus vanellus</i>	Red	Common migrant	Red	

**APPENDIX II. MAMMALS AND REPTILES OBSERVED OR ASSOCIATED TO THE STUDY AREA AND CONSERVATION STATUS**

Common Name	Current Scientific Name	Conservation Category / Action Plans
<b>Mammals</b>		
Jersey Bank Vole	<i>Clethrionomys glareolus caesarius</i>	Own action plan - Jersey Biodiversity Action Plan
Lesser White-toothed Shrew	<i>Crocidura suaveolens</i>	
European Hedgehog	<i>Erinaceus europaeus</i>	Biodiversity Action Plan UK list of priority species
Red Squirrel	<i>Sciurus vulgaris</i>	Own action plan - Jersey Biodiversity Action Plan; Biodiversity Action Plan UK list of priority species
Millet's shrew	<i>Sorex coronatus</i>	Grouped plan - Jersey Biodiversity Action Plan
Mole	<i>Talpa europaea</i>	
European rabbit	<i>Oryctolagus cuniculus</i>	
Wood mouse	<i>Apodemus sylvaticus</i>	
Brown rat	<i>Rattus norvegicus</i>	
Feral ferret	<i>Mustela putorius furo</i>	
Cat (domestic or feral)	<i>Felis catus</i>	
Dog (domestic)	<i>Canis lupus familiaris</i>	
<b>Reptiles</b>		
Green Lizard	<i>Lacerta bilineata</i>	Own action plan - Jersey Biodiversity Action Plan
Slow-Worm	<i>Anguis fragilis</i>	Own action plan - Jersey Biodiversity Action Plan; Biodiversity Action Plan UK list of priority species
Common Wall Lizard	<i>Podarcis muralis</i>	Own action plan - Jersey Biodiversity Action Plan
Common Toad	<i>Bufo spinosus</i>	Own action plan - Jersey Biodiversity Action Plan; Biodiversity Action Plan UK list of priority species
Palmate Newt	<i>Lissotriton helveticus</i>	

### APPENDIX III. OTHER SPECIES MENTIONED IN THE REPORT

Common Name	Current Scientific Name
Black rat	<i>Rattus rattus</i>
Pacific rat	<i>Rattus exulans</i>
Domestic pig	<i>Sus scrofa</i>
Red fox	<i>Vulpes vulpes</i>
Flying fox	<i>Pteropus sp.</i>
Bonin petrel	<i>Pterodroma hypoleuca</i>
Flightless cormorant	<i>Phalacrocorax harrisi</i>
Frigate bird	<i>Fregata sp.</i>
Pelican	<i>Pelecanus sp.</i>
Green turtle	<i>Chelonia mydas</i>

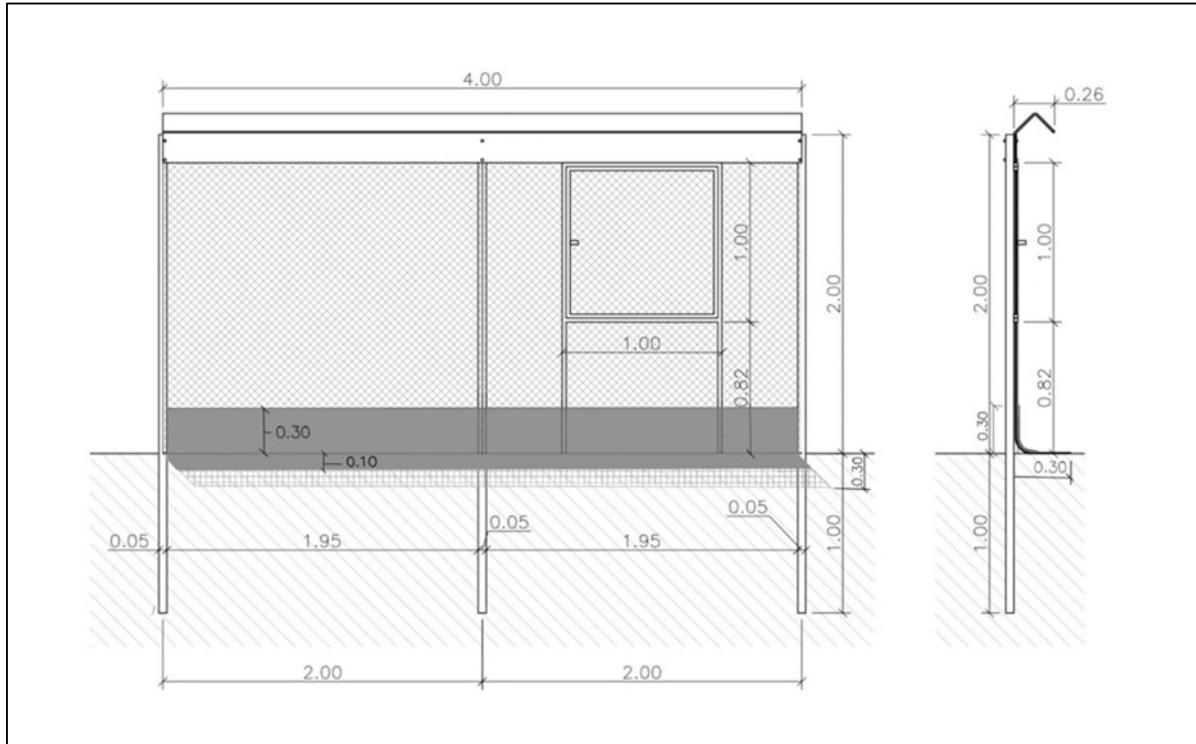
DRAFT

DRAFT



## **Annex 2 – Plémont Seabird Sanctuary: Design Statement**

# PLÉMONT SEABIRD SANCTUARY



## DESIGN STATEMENT



September 2024

Cristina Sellarés de Pedro MSc

BIRDS ON THE EDGE PARTNERSHIP

## **TABLE OF CONTENTS**

1. Proof of concept
2. Project partners
3. Fence specifications
4. Other fence elements
  - 4.1 Gates
  - 4.2 Culverts
5. Structural integrity and maintenance
6. Sustainability
7. Addressing concerns from public consultation
  - 7.1 Public access
  - 7.2 Visual impact
8. References

Appendix. Visualizations of the fence on the site

## 1. PROOF OF CONCEPT

In circumstances where an invasive species cannot be removed from an island, a conservation fence serves the purpose of creating an 'island within an island', thus separating biodiversity and habitats from its threats. The fence makes it feasible to remove and exclude all undesirable animals from within the protected area, thereby allowing conservation work to focus on restoration of habitats and the enhancement of breeding opportunities for the endangered species.

The use of conservation fences has evolved from earlier designs aimed at excluding large mammals such as wolves, foxes, goats and sheep, to more complex set-ups that can cope with smaller and more agile threats, such as mink, ferrets, cats, hedgehogs, rats and even juvenile mice and rats, preventing them from squeezing through the mesh, burrowing under, climbing, or jumping over it.

Predator fences have been perfected over the past 30 years to achieve outstanding results in conserving habitats and protecting threatened biodiversity. Safe habitats have been successfully restored and native species have rapidly increased their breeding success, whilst species that had become locally extinct have returned to the area and formed new colonies.

Globally, over 120 km of modern predator-exclusion fencing has been built to protect endangered wildlife. The size of predator-free ecosanctuaries ranges from less than one hectare to over 3,000 ha.

In New Zealand there are over 40 predator-exclusion fences, and their combined 109 km of fencing protects more than 10,000 hectares. Twelve predator exclusion fences have been built in Hawai'i to date, and another two are in the planning or building stage. The use of this technology is expanding to islands across the Pacific, the Indian Ocean, the Caribbean, and South Atlantic Ocean.

The first fence of this type was installed in 1999 in Wellington, NZ, to create the Zealandia Ecosanctuary, protecting 225 hectares of land. The 8.6km long fence was the result of years of testing and trialling designs against over 200 animals during the 1990s. By the year 2000, the reserve had been cleared of all 13 species of invasive predators that lived there.

The native species thriving in Zealandia Ecosanctuary include 17 species of birds and reptiles which have naturally returned to the area, along 20 species of birds, reptiles, amphibians and invertebrates that have been re-introduced, including tuataras, giant wetas, bellbirds, kakas, kiwis, moreporks and shining cuckoos. Such a large area of predator-free land in the middle of the nation's capital city is a triumph for conservation, as the reserve not only keeps increasing in biodiversity, but is also 'spilling' wildlife over the sides, with birds that were once very rare now repopulating the city gardens and backyards.

## 2. PROJECT PARTNERS

We are working with the world's leading organizations in design and implementation of predator-exclusion fences. Their experience encompasses wildlife research and conservation, testing of designs, implementation, and biosecurity techniques.

Our partners are New-Zealand based Wildlife Management International Limited (WMIL), who work alongside fence design and provider company Xcluder, and Pacific Rim Conservation, a not-for-profit organisation based in Hawai'i. Between them, these organisations have implemented over 30 predator-free ecosanctuaries.

## 3. FENCE SPECIFICATIONS

Fence design and construction techniques vary depending on the physical and ecological characteristics of the site, invasive predators, native species to be protected, and habitats. A significant amount of research and trialling has gone into the development of an effective design of predator-exclusion fences.

The species identified for exclusion from the proposed Plémont seabird sanctuary are brown rat, European hedgehog, feral ferret, and domestic cat. The fence specifications take into account the size and abilities of these target species in order to prevent them from accessing the reserve. These can be summarized as follows:

Species	Push	Chew	Dig	Climb	Jump (height mm)	Mesh hole size to contain (mm)
Brown rat	✓	✓	✓	✓	800	13
Ferret	✓	✓	✓	✓	800	50
Cat	✓	✓	-	✓	1800+	50
Hedgehog	✓	✓	-	✓	-	50

Fortunately, the current 'best practice' design already features all necessary specifications to account for the animals and behaviours listed above, and surpass the physical capabilities of all our four target species:

- A 1.9-2.0 m high 'base fence', to prevent all target species from jumping over the fence.
- A hood or cap to prevent climbing mammals from getting over the fence.
- Wire mesh attached posts, with a square no wider than 7mm to exclude juvenile rats. This is achieved with a galvanised mesh of 6x25mm, or our proposed tight-woven chain link of 1.2mm diameter wire, with internal aperture of 7x10mm.
- Posts 2.5m long, 100x100mm square galvanised posts, supported by cross beams of galvanised steel square posts (between 50x50mm and 100x100mm) set up at 1m high.
- An underground skirt extending 300-500mm horizontally facing the predator side, pinned 50-100mm underground, to prevent animals burrowing under the fence. In cases where a buried skirt is not feasible, such on a rocky substrate, concrete or a similar compound will be used to secure the mesh along the rock.

Image of the design of the main fence base.

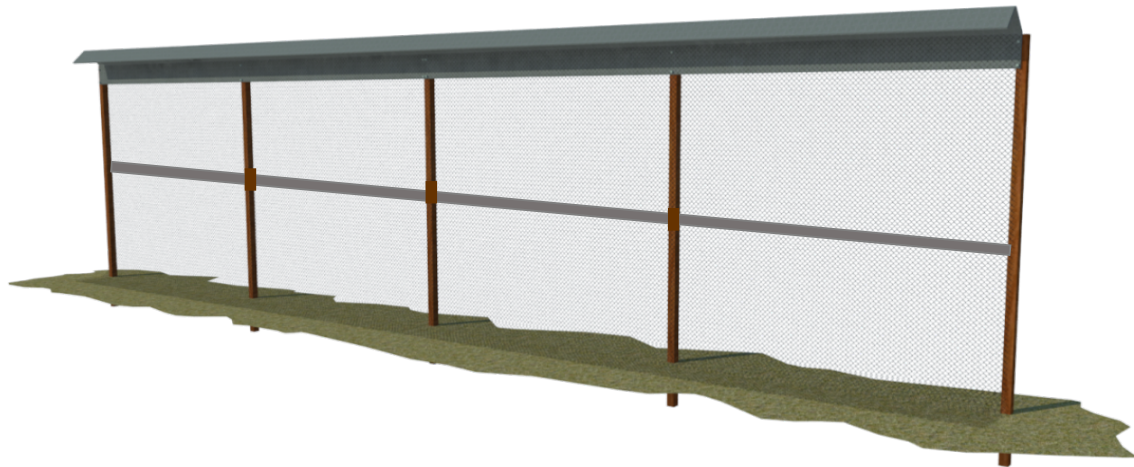
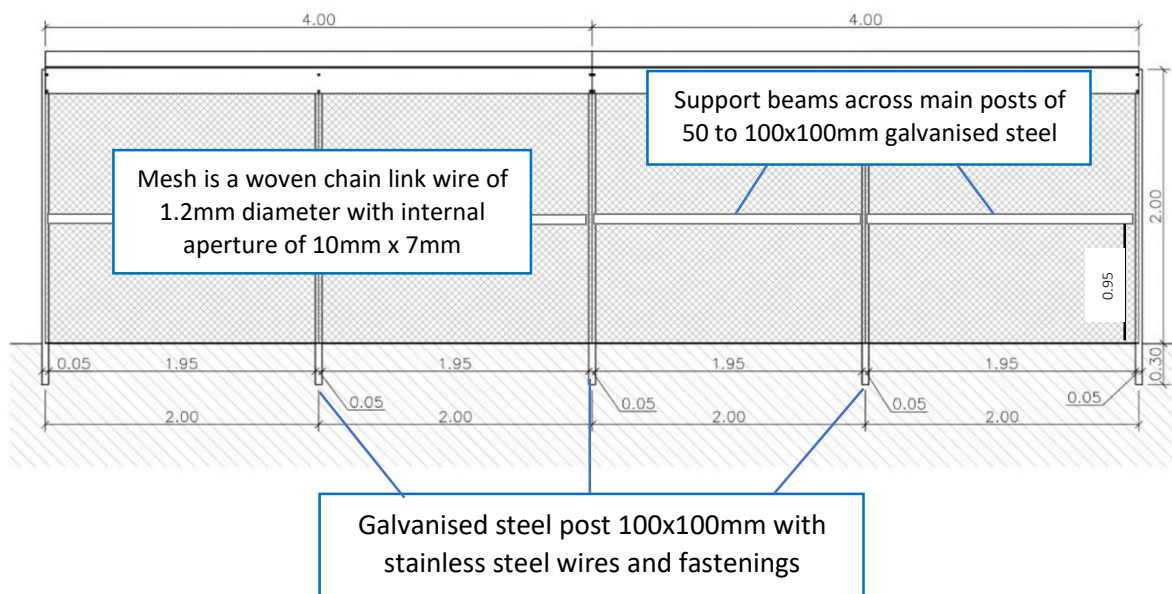


Image of the main fence (front). Distance between posts to be 5.00m in most cases, but as low as 2.00m subject to ground conditions.



Images 3&4: Coastal fences in Dunedin and Kaikoura, NZ (both untreated for colour).



Image of a cross-section of the main fence and an of a mouse attempting to penetrate.

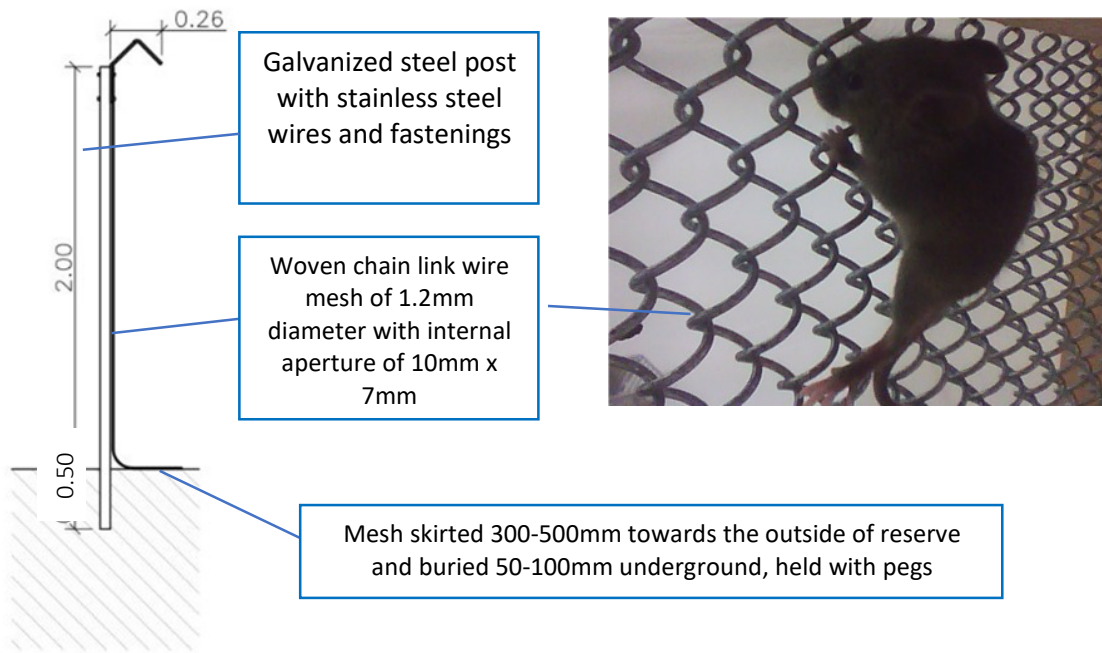


Image of the fence hood.

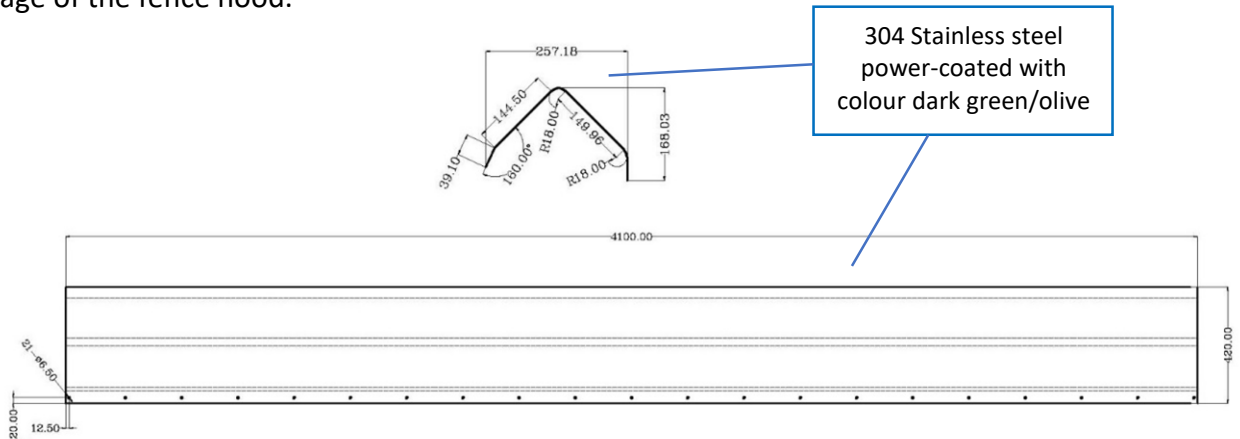
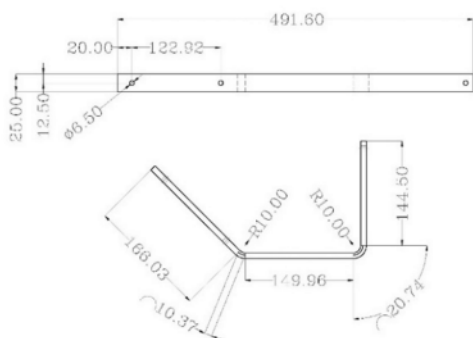


Image of the left- hood bracket specifications.



## 4. OTHER FENCE ELEMENTS

### 4.1 Gates

Image of the livestock and service gate. Maximum length on this design can be 3.66m, however real length will be 2.00m. Mesh and posts of same specifications as main fence.

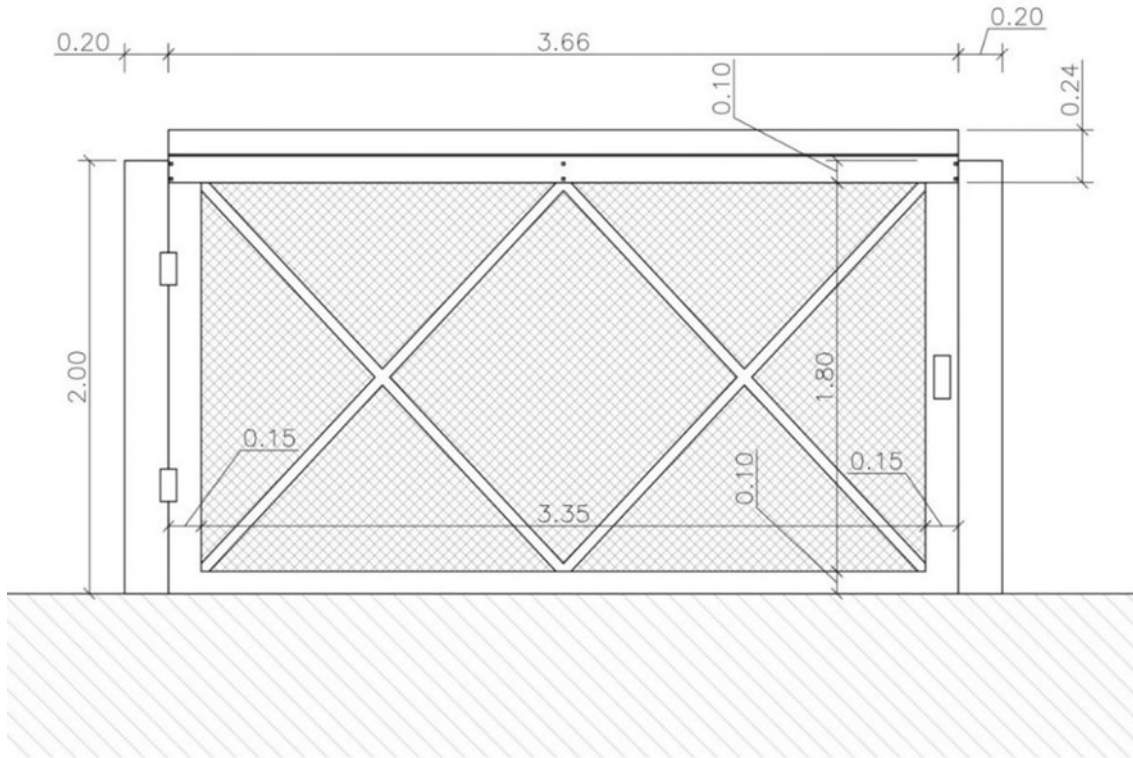
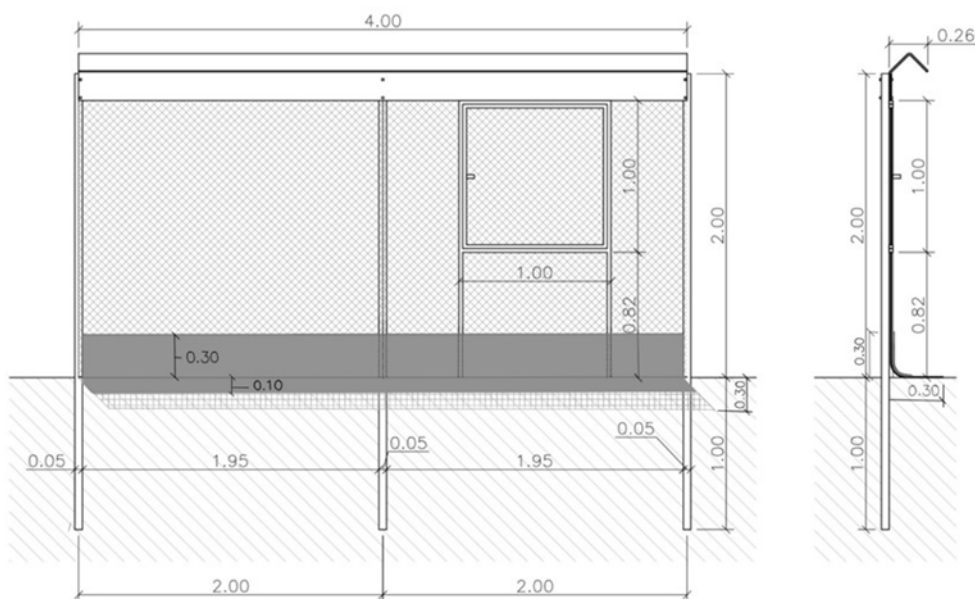


Image of pedestrian half-gate.





## 4.2 Culverts

Within the proposed ecosanctuary there is one small seasonal stream that leads into the sea from a coastal valley. In order for the fence to be effective, all entries will be sealed to prevent invasive species from getting inside the reserve, especially rats who often use waterways to move between areas.

The stream will be fitted with a specialised culvert at the point of intersection with the fence, and the fence will be installed on the ground placed over the culvert. The culvert will be held in place by bags of concrete, and covered by the mesh skirt of the fence, which will go over the top of the culvert.

The culvert will comprise a 600mm-diameter round precast concrete pipe with a swinging gate, encased in a mesh box with a square inset that is hinged at the top and locked at the bottom (a block and lock system), so that it can be opened for maintenance and to flush out debris. This mesh inset will be at both ends of the culvert for added biosecurity. The mesh inset will be bolted to the inside of the concrete culvert. The flap gate and metal cage will be bolted down to the concrete slab.

The pipe will be secured with concrete bags, placed over the face of the soil, which also directs the flow in the upstream end of the culvers and reduces risk of erosion when then stream is at full flow. To ensure that rats do not tunnel through the soil behind the bags or through the gaps between them, steel mesh will be added between the soil and the bags.

Image of predator-proof culvert under fence.

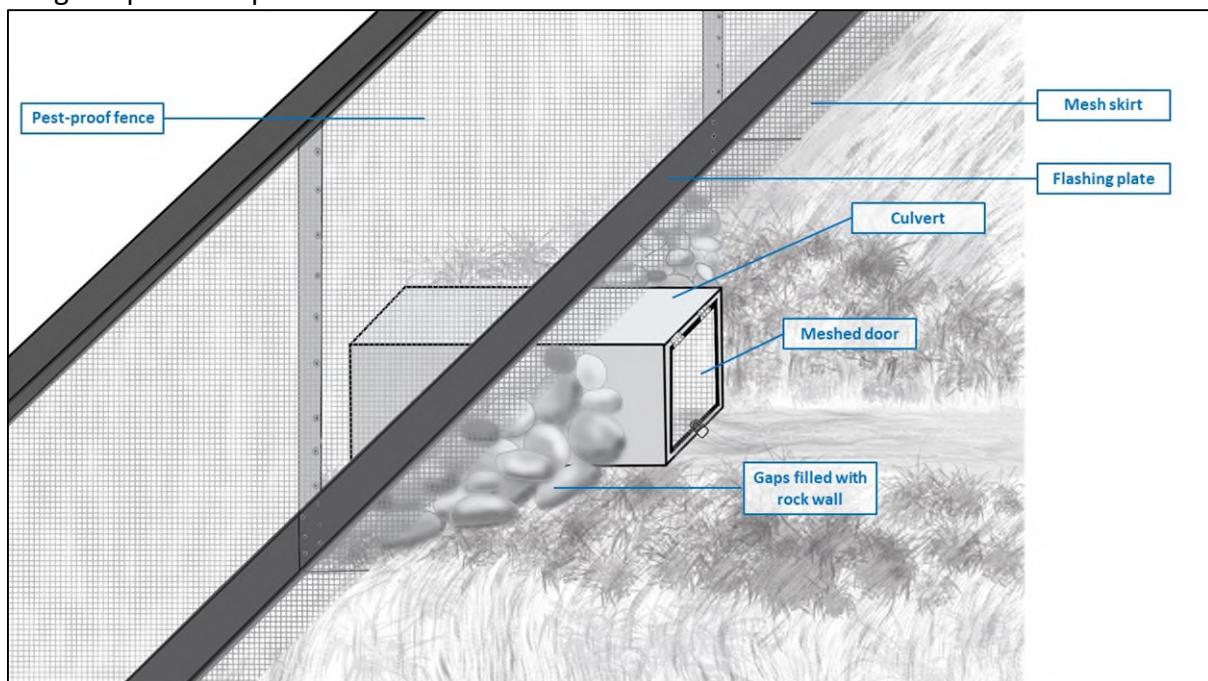
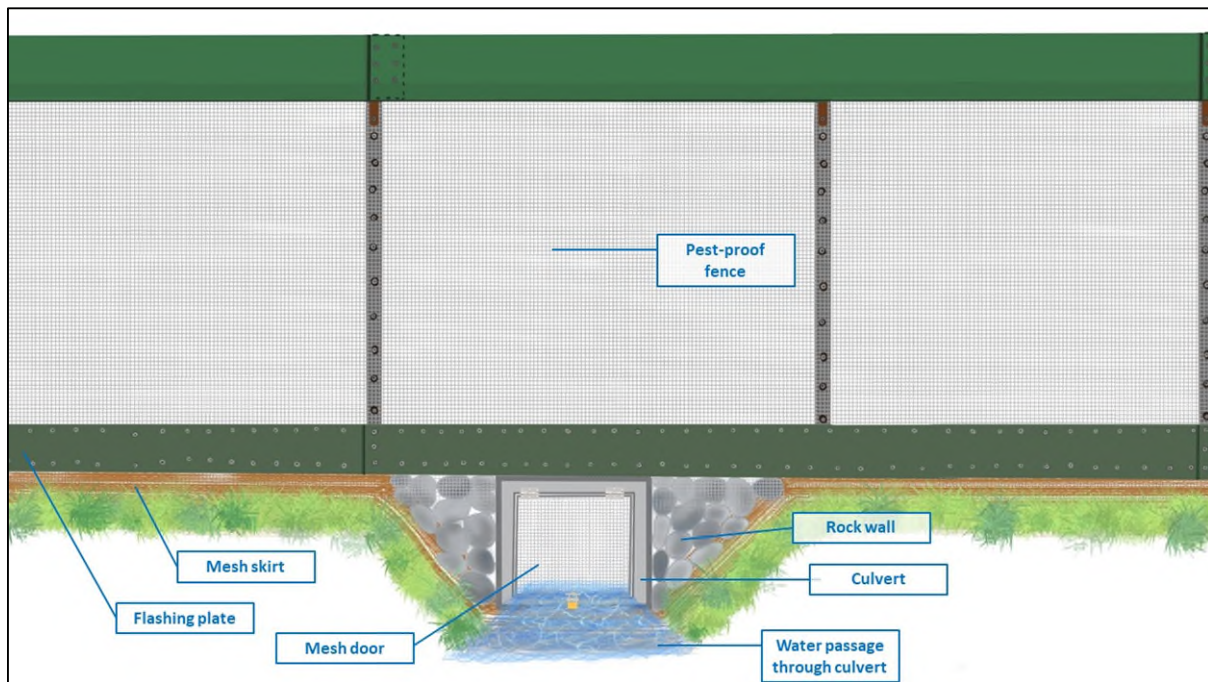


Image of predator-proof culvert under fence.



Ongoing monitoring and management of the flow is paramount to maintain the integrity and effectiveness of the fence. This will be achieved with regular checks of the stream flow, either manually or with remotely operated monitoring devices, in order to ensure that any sudden increases of flow do not undermine the integrity of the culverts or the fence. The strength of the flow and erosion damage can also be managed in the long-term with targeted landscaping of the edges and vegetation of the stream near the upstream culvert, and creating features such as rocky inlets that absorb and break down strong flows.

## 5. STRUCTURAL INTEGRITY AND MAINTENANCE

The fence will be able to withstand the environmental conditions of Jersey, which include heavy rainfall and storms. The average rainfall in Jersey is 875mm/year (Jersey Met Climate Statistics, 1981-2010), with the wetter season being from September to March. The prevalent winds are westerly, with an annual average speed of 15.2mph (data from Jersey Airport and Weatherspark statistics, 1980-2016). The Island can occasionally experience violent storms (64-72mph) which can be accompanied by gusts of hurricane-strength winds (73mph and over). Structural integrity of the fence has been a particular problem for fences exposed to hurricane-type winds, such as the case of the Corvo seabird project in the Azores (T. Pipa *pers. comm.*). It is believed that the fence specifications can withstand the weather conditions of Jersey (B. Bell *pers. comm.*). During the recent storm Ciaran in November 2023 the 'demo' section of fence, built with materials of less quality than the real fences, withstood the Force 12 gusts without any apparent damage.

Other factors that could compromise the integrity of the fence are the effect of animals, vegetation, inanimate objects, and accidental or deliberate human action. A system to detect damage will be implemented, involving visual checks, automated alarms, and potentially security cameras if vandalism or theft become a problem.

Potential damage from tree branches has been mitigated by choosing a route away from most trees and large shrubs.

Beside maintaining the structural integrity of the fence, the vegetation along a two-metre strip either side of it will be kept at a height of 100-150mm, in order to avoid damage to the fence and to prevent animals using the vegetation to jump over the fence. This strip of cleared ground might create good conditions for native coastal grasses and it will also allow easy access for fence checks, maintenance works, and drainage for surface run-off.

Evidence from other fences of this type suggests that their lifespan in a coastal or high salt-prone area is between 20 and 30 years. Small maintenance works are predicted to be necessary every five to ten years.

## **6. SUSTAINABILITY**

It is of high priority to this project a management regime that mitigates the environmental impact of the fence installation and maintenance.

Beyond the construction and decommissioning methods stated in the 'Project Implementation' document, the carbon footprint of the project will be limited by sourcing materials via local providers, and having the bulk of the materials that cannot be produced in Jersey imported in one single shipment.

The Ecosanctuary aims to encourage the public to spend time outdoors and to visit the area. All public communications regarding seasonal activities such as puffin watches and seabird trail walks will include recommendations to travel to the site by bus, bicycle or foot, in order to minimise the public's carbon footprint. The project has also engaged with the StOuen Parish and local infrastructure authorities to promote the bus service to the area, installation of bike racks, and charging points for electric vehicles.

## **7. ADDRESSING CONCERNS FROM PUBLIC CONSULTATION**

### **7.1 Public Access**

The avoidance or minimisation of the impact on the public use of the site was prioritized from the early stages of development, and these considerations influenced heavily in the choice of site and the route of the fence.

To avoid any impacts on the direct use of the area, the fence is sited away from footpaths and popular destinations such as the Plémont Point headland, which is visited regularly by residents, tourists, cyclists, dog-walkers, bathers and spear fishermen.

Instead, the fence will follow the coastal cliffs, 50m below the public footpath.

The only group of users that regularly use this site are shoreline anglers and low-water fishermen. The fence will be an obstacle to their traditional fishing marks below the cliffs. Consultations with this collective started in 2022 and have resulted in a design that includes pedestrian gates located at the points where the fence intersects their traditional routes to their fishing marks. A system will be established to ensure that all fishermen, and any other members of the public that wish so, can access the site via these pedestrian gates.

## **7.2 Visual Impact**

The fence will be visible from some sections of the public cliff path. Whilst it is not possible to alter the technical specifications of the fence, for risk of compromising its effectiveness, some aesthetical aspects of it can be modified to help it blend with the surroundings. The measures taken to mitigate the visual impact of the fence are:

1. Route: The fence will be placed between 50 meters below the public footpath
2. Colour of posts: The galvanized steel posts will naturally wear creating a dull effect that will not be easy to notice against the background from the coastal footpath.
3. Colour of hood: The hood is likely to be the most visible element of the fence, and it will be powder coated with a colour matching the surroundings without reducing its effectiveness.

## 8. REFERENCES

Bell, E. (2014). *FEASIBILITY STUDY FOR A 'MAINLAND ISLAND' ON SAINT LUCIA: A PEST- PROOF FENCE AT A MAINLAND SITE.*

Bell, P. (2014). *Predator Free Rakiura Halfmoon Bay Project—Analysis of options for proposed predator fence.*

Bode, M., Brennan, K. E. C., Morris, K., Burrows, N., & Hague, N. (2012). Choosing cost-effective locations for conservation fences in the local landscape. *Wildlife Research*, 39(3), 192.

Bode, Michael & Wintle, Brendan. (2009). How to Build an Efficient Conservation Fence. *Conservation biology : the journal of the Society for Conservation Biology*. 24. 182-8.

Cape Kidnappers Sanctuary [REDACTED]

Connolly, T. A., Day, T. D., & King, C. M. (2009). Estimating the potential for reinvasion by mammalian pests through pest-exclusion fencing. *Wildlife Research*, 36(5), 410.

Day & MacGibbon (2002) Escape behaviour and physical abilities of vertebrate pests towards electrified and non-electrified fences. Xcluder™ Pest Proof Fencing Company, unpublished report.

Ecology, R., White, G., & Hirons, G. (2019). *The predator exclusion fence manual*. Royal Society for the Protection of Birds, Sandy, UK.

Ka'ena Point Ecosystem Restoration [REDACTED]

Kilauea Point Natural Refuge [REDACTED]

MacGibbon, R.J. & Calvert (2002). Evaluation of the Effectiveness and Suitability of Xcluder™ Pest Proof Fencing Technology as a Conservation Management Tool in Hawaii. Xcluder™ Pest Proof Fencing Company unpublished report. 49 pp.

McInturff, A., Xu, W., Wilkinson, C. E., Dejid, N., & Brashares, J. S. (2020). Fence Ecology: Frameworks for Understanding the Ecological Effects of Fences. *BioScience*, baa103.

Pacific Rim Conservation [REDACTED]

Rotokare Scenic Trust [REDACTED]

Turnbull, A. (2021). *Zealandia Te Māra a Tāne: Predator Exclusion Fence Design and Maintenance. New Zealand.*

Xcluder Pest Proof Fencing [REDACTED]

Young, L. C., Behnke, J. H., Vanderwerf, E. A., Raine, A. F., Kohley, C. R., Dalton, M., Mitchell, M., Tonneson, H., DeMotta, M., Wallace, G., Nevins, H., Hall, C. S., & Uyehara, K. (2018). *The Nihoku Ecosystem Restoration Project: A case study in predator exclusion fencing, ecosystem restoration, and seabird translocation.*

Young, L. C., Vanderwerf, E. A., Mitchell, C., Yeun, E., & Miller, J. (2012). *The use of predator proof fencing as a management tool in the Hawaiian Islands: A case study of Ka`ena Point Natural Area Reserve.*

Zealandia Ecosanctuary

**APPENDIX. VISUALISATIONS OF THE FENCE ON THE SITE**









## **Annex 3 – Plémont Seabird Sanctuary: Management Plan**

# PLÉMONT SEABIRD SANCTUARY



## MANAGEMENT PLAN



September 2024

Cristina Sellarés de Pedro MSc

BIRDS ON THE EDGE PARTNERSHIP

## TABLE OF CONTENTS

1. Description of works
    - 1.1 Invasive plant removal
    - 1.2 Habitat restoration and enhancement
    - 1.3 Fence installation and decommissioning
      - 1.3.1 Aims
      - 1.3.2 Site Access and Site Clearance
      - 1.3.3 Waste management
      - 1.3.4 Fence Installation
      - 1.3.5 Gates Installation
      - 1.3.6 Culvert installation
      - 1.3.7 Decommissioning: dismantling, removal, disposal and reinstatement
    - 1.4 Non-native predator eradication
    - 1.5 Biosecurity protocol
      - 1.5.1 Prevention
      - 1.5.2 Detection
        - 1.5.2.1 Surveillance
        - 1.5.2.2 Reporting
      - 1.5.3 Incursion response
    - 1.6 Ecological monitoring
    - 1.7 Community programme
  2. Schedule of works year 1 to 5
  3. Project targets (key performance indicators)
  4. References
- Appendix. Statement from Cape Sanctuary, NZ, 15 March 2024

## 1. DESCRIPTION OF WORKS

### 1.1 Invasive plant removal

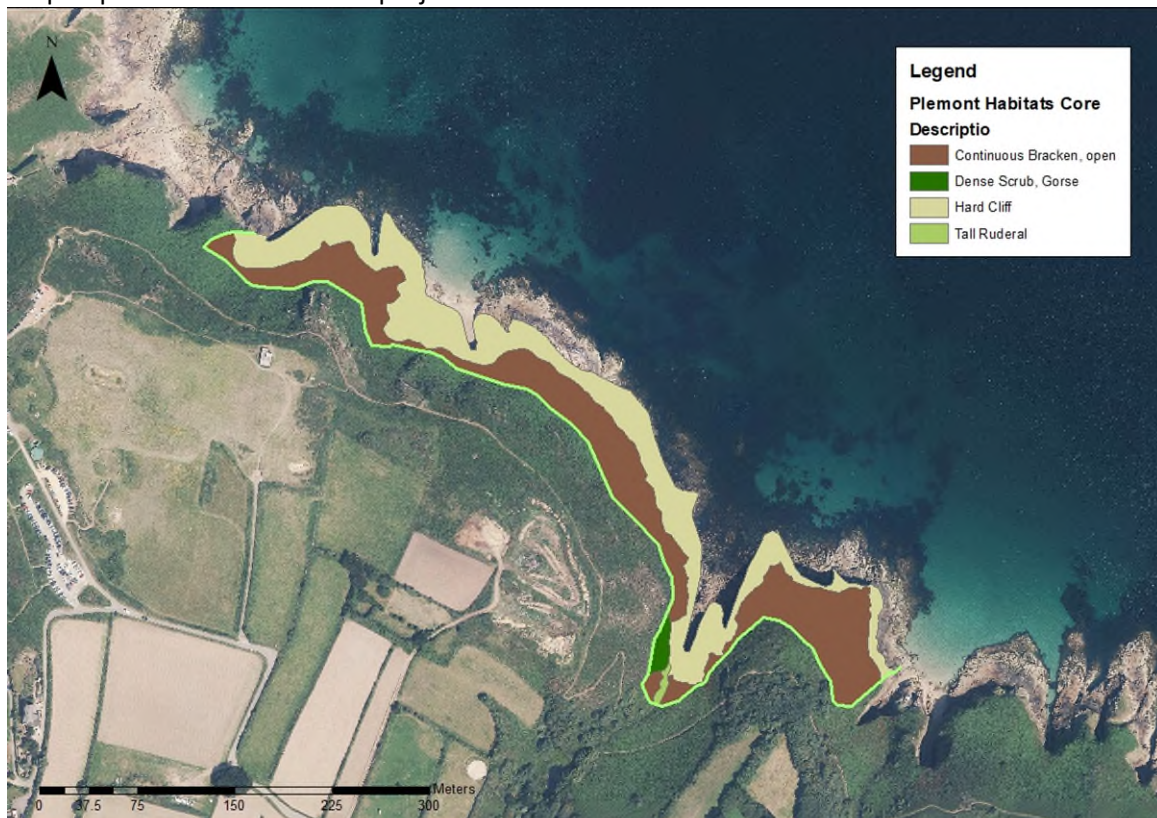
Bracken is the dominant plant at the project site, taking 96.3% of all the ground suitable for plants. Bracken is a species-poor type of habitat, and it has been described as 'invasive' in this area by the Jersey Integrated Landscape and Seascape Character Assessment (ILSCA), due to the fact that it supports a very limited number of native plants and animals.

Removing a large portion of the bracken will open up areas to be colonized by native plants and create a mosaic of coastal habitats such as grassland and heathland. Coastal grasslands on sloped above sea-cliffs are the preferred nesting environments by burrowing species such as Atlantic puffins and Manx shearwaters.

The following methods have been reviewed by local botanists and members of the Botany Section of La Société Jersiaise in an advisory capacity.

- a. Removal: This is the preferred method to remove bracken and it will be achieved by cutting bracken by hand, with electric tools, or with a specialised robotic machine.
- b. Curtailing new growth: This will be achieved with a grazing herd of Manx Loaghtan sheep, which at the right times of the year will trample the new shoots of bracken, whilst encouraging the re-growth and colonization of native grasses and other coastal plants.
- c. Chemical control: if there are some important areas where bracken cannot be cut, such as the 2m-strip either side of the predator fence, a regulated and licenced chemical will be used for short-term treatment.

Map of present habitats at the project site



Map of areas where bracken will be removed from (in brown)



## 1.2 Habitat restoration and enhancement

The management of habitats will aim to increase native plant diversity. Following advice from local experts, works will be dependent on what plant species colonise the areas where the bracken has been removed from.

There is potential at the site for a significant increase in native plant diversity after removal of bracken. The recent restoration of adjacent land, at the National Trust for Jersey headland, has resulted in diversity levels higher than expected: an average of 18 plant species on 2x2m quadrants.

Habitats will also be enhanced to increase the appeal of the site to prospecting seabirds and to facilitate quick colonisation by prospecting pairs. The following techniques will be used:

- a. Artificial burrows and nest boxes: These will be placed in suitable nesting sites to facilitate colonisation by new pairs of burrowing seabirds, such as Atlantic puffin, Manx shearwater and storm petrel. These species can dig their own burrows, but finding them ready-made quickens the colonization process. Artificial burrows are also easy to manage, allowing researchers to monitor breeding pairs and their chicks with little disturbance.
- b. Decoys of puffins and other seabirds: The use of decoys at secure breeding sites is a technique that creates the illusion of a large, safe colony, attracting the interest of prospecting birds looking for a mate and a place to nest. Decoys of puffins, razorbills and guillemots will be placed at suitable sites during the breeding season, and will be removed during the rest of the year.
- c. Broadcast of colony sounds: This method sends the message of a large, secure seabird colony, but it is particularly useful for nocturnal birds such as Manx shearwaters. The calls will be played at remote locations, after tests to ensure that they cannot be heard from the nearest properties.

## BIRD DECOYS

Bird decoys are used to attract wildlife to sites. Traditionally used for hunting, this technique has been used since the 70s in over 750 seabird conservation projects to increase populations of threatened seabirds or even start new populations.



Decoys are particularly useful for seabird species, as they are colony breeders who nest in open landscapes, that are easy to see from far away.



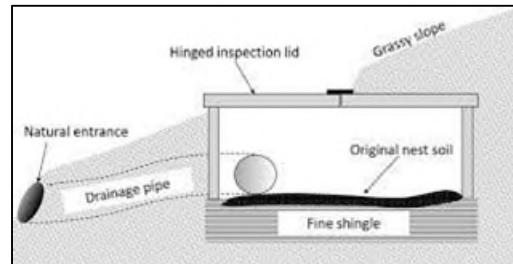
Puffin decoys have been used to increase numbers at many colonies as well as starting new colonies in Maine, US, Isle of Man, UK, and Hrisey Island, Iceland.



## ARTIFICIAL BURROWS

Like the popular bird-boxes used in gardens, artificial burrows provide safe nesting homes for wild birds who nest underground, such as puffins, petrels, shearwaters, and even penguins.

Artificial burrows have a high uptake and birds that use them have a higher breeding success rate and produce more chicks than those in natural burrows.



Artificial burrows are usually safer than natural burrows against the elements.

They also facilitate the monitoring of the eggs and chicks via an observation hatch or a pinhole camera.

Artificial burrows can also be used to translocate chicks from dangerous sites into fenced sanctuaries. When the grown up chicks are ready to breed, they will return to the site where they fledged from, starting a new colony within the boundaries of the safe sanctuary.

In Lundy Island artificial burrows for Manx shearwaters have allowed the population to recruit new breeding pairs increasing the population from 6,000 to over 25,000 since the rats were eradicated.

*Note: Puffin decoys and artificial burrows were installed at the project site in 2019 but have failed to attract new pairs; this is believed to be linked to the presence of rats and ferrets in the area.*



## **1.3 Fence installation and decommissioning**

### **1.3.1 Aims**

Due to the sensitivity of the coastal area where the predator exclusion fence is to be located, installation methods will aim to keep disruption and disturbance to a minimum. This will be balanced with the requirements to undertake the work in accordance with relevant health and safety at work. All works will follow the recommendations for mitigations detailed in the Preliminary Ecological Appraisal Report.

### **1.3.2 Site Access and Site Clearance**

Prior to commencement of installation works the route of the fence will be cleared of vegetation by strimming and mowing. This will be undertaken over a width of approximately 4m and will represent the 2m bands at either side of the fence that will be maintained clear of vegetation for the duration of the project. In addition, access will be cleared in the same manner from agreed access points from adjacent roads. These will be permitted to regrow following completion of the installation.

All vegetation clearance will be undertaken at times and in a manner as advised by ecologists to minimize disruption to protected species.

As much as possible work will be undertaken using lightweight tracked mini-diggers and suitable off-road vehicles to assist with creation of post holes, shallow excavation of the trench to embed the mesh skirt and for moving materials. This is to reduce risks from manual handling injury.

There are likely to be some locations where excavation will be required to level the ground to ensure safe access. This will be kept to a minimum. All excavated soil will be stockpiled locally and the ground reinstated to its original profile upon completion.

Where the ground is too steep or unsafe for mechanised plant the installation will be by hand. Temporary guide ropes and fall prevention lines will be installed by qualified individuals to reduce risks from falling. Installation will be undertaken by individuals trained in industrial rope access techniques. This will require anchor points for safety ropes which may consist of either vehicles, driven steel anchors or anchors bolted to the rock faces. All temporary anchors will be removed, except where they may be required for future access for maintenance and repair of the fence.

A site compound comprising welfare facilities, site office and storage containers will be set up in an agreed location. Unsurfaced areas that are frequently trafficked will be protected with plastic 'ground-boards' to reduce risk of rutting and damage to turf.

All small plant (generators, breakers etc) will be stored and refuelled on drip traps or similar. All plant will be regularly serviced to reduce likelihood of breakdowns and leaks. Spill kits will be available on all work sites in case of oil or fuel spill.

Due to the remote location of much of the work the emergency services will be briefed in advance and when working near or on the cliffs the coastguard will be kept informed of work locations daily in case of emergency. Trained rope access personnel will have training to assist recovery of injured persons from cliffs and steep slopes. All personnel on site will have, as a minimum, a basic level of first aid training.



### **1.3.3 Waste Management**

Prior to works commencing the route will be surveyed in detail in order to accurately determine the quantity of materials required. Waste is expected to be limited as all materials will be delivered to site in the required quantity and size, minimising cutting on site.

All waste will be collected and returned to the site compound at the end of each shift, where it will be collected in a covered skip. This is expected to consist primarily of packaging such as 'post-crete' bags, fixing boxes, protective wrappings and ties.

### **1.3.4 Fence Installation**

Fence posts will be installed in holes dug by either mechanised auger or by hand, with the posts being concreted into the holes. The size of the holes will be kept to a minimum and the least amount of concrete required will be used. All concrete will be kept below ground level and the hole will be covered with topsoil to reduce visual impact. Concrete used will be 'dry-mix' and will be hydrated by wetting after placing and compacting around the post.

Where the ground is rocky, but the rock is fractured and weak, the rock will be broken up and loosened with a breaker in order to excavate the hole. Where rock is strong and intact the post will be bolted to the rock.

The trench required for embedment of the base of the mesh skirt will be dug with mini-digger, or by hand. Where the ground is rocky the base of the mesh will be held down with concrete and fixed to the rock with bolts.

### **1.3.5 Gates Installation**

Gate posts will be installed in the same manner as fence posts however more concrete will be needed to ensure the gate is secure. A strip of ground beneath the gate will be filled with concrete and fitted with reinforcement bars to prevent damage if driven over. This will be level with the ground to create a flat surface for the gate to close over/up to, in order to prevent predators getting beneath the gate.

### **1.3.6 Culvert Installation**

Prior to culvert installation the vegetation will be cleared in accordance with requirements by an ecologist to avoid unnecessary disruption to protected species. Any water in the stream will be diverted, either by excavation of a temporary bypass or by pumping.

The ground beneath the culvert will be prepared by removing soft clay and silt and replacing with granular fill to create a stable base. To support the steel mesh box and swinging gate a reinforced concrete pad will be cast.

The culvert is expected to be 600mm diameter concrete pipe. This shall be buried with fill to bring the ground level up. The head wall (upstream of the culvert) will be built from cement filled bags. Between the fill and the headwall a fine mesh will be placed to prevent burrowing animals digging through. The steel mesh box with a swinging gate will be bolted down to the concrete slab.

The sides of the stream bank will be reinstated, and seeded with general purpose grass seed to reduce erosion. The diverted stream water will be reinstated to flow through the culvert.

### 1.3.7 Decommissioning: dismantling, removal, disposal and reinstatement

At the end of the project the fence and its components will be readily removed.

- Mesh fence, metal hood and fittings will be removed and recycled.
- Posts will be pulled or dug out, the concrete adhering to the posts will be broken off, collected and crushed for reuse as fill elsewhere.
- Aluminium posts will be reused or recycled. Timber posts will be reused if not damaged, or could be reclaimed for other purposes (for example at Acorn Enterprises or similar). Any timber unsuitable for reuse could also be stacked at designated locations to create wildlife habitats or sent to the energy recovery facility at La Collette.
- Post holes will be filled with clean soil.
- The culverts could be removed to return the stream to its original state, or could also be partially blocked to create ponds/wetland habitats. Concrete culvert parts will be reused elsewhere or crushed and recycled. Concrete bags will be crushed and recycled.

Images of the project site after removal of the 'demonstration fence' showing the post holes after removal of concrete and being filled with ground soil.



#### 1.4 Non-native predator eradication

The number of predators left inside the ecosanctuary in the period after the fence has been installed is predicated to be low, as ecological research suggests low densities of them on the lower part of the cliffs, and because they are likely to avoid the area during the installation works.

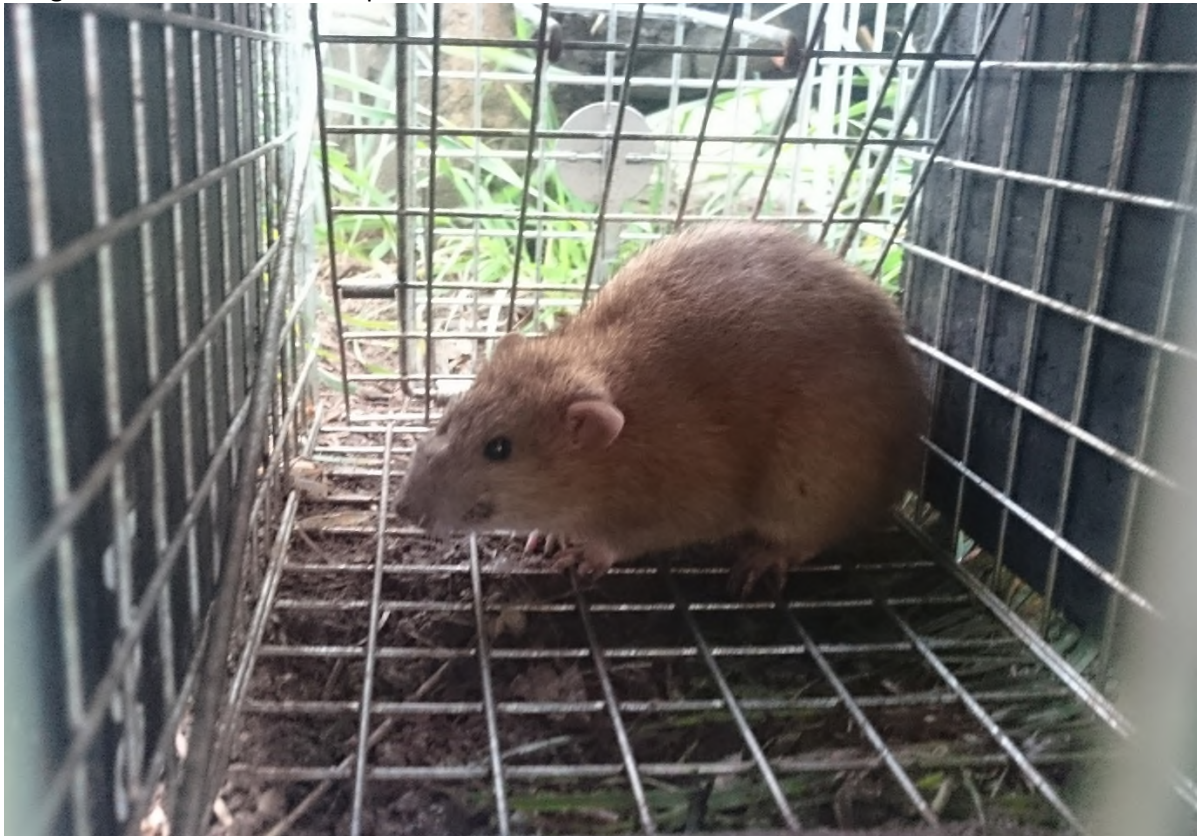
All trapping, handling and potentially lethal techniques will be used under licenced by the Government of Jersey Natural Environment authorities and will follow agreed methods and mitigations listed in the relevant risk assessments.

The eradication will be phased by prioritising species and techniques as follows:

- a. Monitoring: Continuation of the monitoring programme of non-native predators with trail cameras, tracking tunnels and wax blocks along the fence and on a 100m square-grid.
- b. Phase 1: Live traps on a 50m square-grid. All non-native mammals trapped will be relocated outside the fenced ecosanctuary. Species considered domestic such as ferrets and cats, and any injured wildlife, will be surrendered to a licenced veterinarian authority. This phase will conclude when no more cats, hedgehogs and ferrets are detected within the ecosanctuary.
- c. Phase 2: To be launched if rats are still present inside the ecosanctuary and will use rat-specific lethal traps in a 50m-square grid.
- d. Phase 3: In conjunction with Phase 2, after the rate of capture drops below a sustained level, poisoned bait to be placed on a 50m square-grid for a limited period during winter months.

Eradication to be undertaken for a minimum of two consecutive winters (November – February) and to be repeated if signs of target species persist.

Image of a rat in a humane trap.



## **1.5 Biosecurity protocol**

A predator-exclusion fence is not a perfect, infallible tool. Chances of re-incursions exist. The aim of a biosecurity plan is to provide clear context and guidance to the management of an ecosanctuary in order to maintain the site's predator-free status and to prevent, detect, and response to any potential incursion.

A biosecurity protocol will be prepared during the first year of the project and prior to eradication works, to cover all aspects details of the biosecurity elements:

### **1.5.1 Prevention**

The re-invasion of the sacntuary can be prevented with a deep understanding or all the potential pathways and a number of measures placed on these pathways to reduce their risks.

The biosecurity protocol will feature:

- a. A detailed list of all known incursion pathways, described and categorised with a risk score.
- b. A table with all the key actions, resources, and responsibilities to ensure that the excluded predators do not re-invade the ecosanctuary via any of the pathways.

All the stakeholders and relevant members of the immediate community will be familiarised with the biosecurity protocol and will be kept up to date of any changes within in.

### **1.5.2 Detection**

#### **1.5.2.1 Surveillance**

The predator monitoring that has been undertaken since 2017 will be replaced by a new research programme during and after the eradication phase. This programme will cover the entirety of the ecosanctuary and will feature a combination of passive surveillance and active detection.

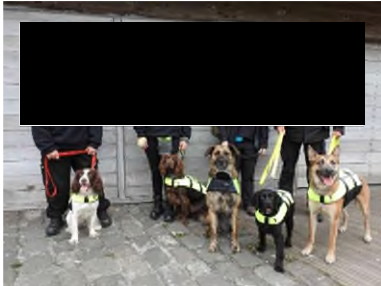
- a. Passive surveillance:
  - Trail cameras.
  - Flavoured wax blocks to detect bite marks.
  - Chew cards to detect bite marks
  - Inked tracking tunnels.
- b. Active detection:
  - Scent-trained detection dogs.
  - Thermal imaging surveys at dusk.

Passive surveillance will involve monitoring stations set up on a 100m square-grid across the ecosanctuary, along the perimeter, and at high-risk areas in higher densities, such as gates and both ends of the fence. Each monitoring station might feature a trail camera, flavoured wax blocks, or inked tracking tunnels. All bait used will be wildlife-safe. Each station will be checked fortnightly.

Active detection will involve ground-sweeping exercises across the ecosanctuary with dogs trained to pick up the scent of rats and ferrets, and dusk surveys from vantage points with a thermal imager, that will pick the thermal signature of nocturnal mammals and other animals.

## CONSERVATION DETECTION DOGS AND BIOSECURITY

The use of dog's scent detection skills has evolved from roles in search and rescue, custom checks, or bomb detection, to conservation purposes, where it has increased dramatically over the past decade. Dogs are being trained to detect elusive species, environmental cues, emerging diseases, or even threats to wildlife.



Dogs are helping to monitor endangered species by finding carcasses, scats, or scents of live animals and their burrows. In 2021, the RSPB used conservation dogs to find the first breeding colony of storm petrels in the Isle of May.

Dogs can also be used for eradication and biosecurity purposes. In the Orkney Islands dogs are being used as part of the stoat eradication project.

In 2018 two trained dogs walked 1,550 miles across South Georgia Island, seeking signs of rats and mice. Their skills helped confirm that the eradication project had been successful and that South Georgia is now pest and predator free.

The use of trained dogs can be the fastest way to find endangered species or detect potential threats in vast landscapes of difficult terrain.



Using dogs saves valuable time, effort and resources, reducing disturbance at sensitive sites.

Dogs trained to detect rats and ferrets can be used at the Plémont Seabird Sanctuary to monitor the progress of the predator eradication, and to identify potential incursions during routine biosecurity checks.

Detection dogs will be one component of a complex Biosecurity Protocol that will also involve structural checks, remote monitoring, and community awareness campaigns to reduce the threat at the pathways.



### **1.5.2.2. Reporting**

A formal reporting procedure will be put in place to collect any sightings or signs of incursions, whether from monitoring stations, surveys, or via stakeholders and members of the general public. Annual information campaigns, interpretation boards, and letters to neighbours of the site will raise awareness on the predator-free status and provide a clear and simple reporting system, as well as a guide to identifying the threat species. The reporting protocol will aim to:

- a. Raise community awareness and make all information easily accessible.
- b. Ensure correct species identification and collection of evidence when possible.
- c. Establish a simple and direct reporting pathway that reaches the biosecurity team.
- d. Provide the designated responders with a guide to extract the key information and course of action.
- e. If the sighting has been reported by a member of the public, a follow-up interview and a site inspection to be carried out whenever possible.
- f. Keep records of all reports and locations, even the ones that are confirmed to be false.

### **1.5.3 Incursion response**

A team of staff and designated responders will be established prior to the start of the biosecurity phase, and will be familiar with the biosecurity protocol and the incursion response plan. All equipment necessary to respond to an incursion will be stored in a secure and accessible location.

Any confirmed sign of an excluded species will be acted upon with speed. The response will be initiated by the designated responder who will follow an established response plan. The response plan will involve:

- a. Mobilising the response team.
- b. Informing all stakeholders and affected parties.
- c. Following a best communication procedure.
- d. Depending on the species, establishing a 50m grid of human traps or lethal traps and bait stations for at least 250m in every direction from the sighting or location.
- e. Deploying the detection dogs to re-locate the target at regular intervals.
- f. Undertaking dusk surveys to re-locate the target.
- g. Checking all cameras and bait stations daily.
- h. Recording all bait taken from each station on a daily basis.
- i. Reducing the amount of bait in the boxes, after two weeks.
- j. Increasing the grid of bait stations, if take of bait does not stop after two weeks.

The end of the response will be decided after two weeks without any bait being taken.

After an incursion has taken place, a review will examine all aspects of the response and possible causes of the incursion, in view to enable refinement of the response plan and biosecurity measures.

## 1.6 Ecological monitoring

The aims of the ecological monitoring programme are to establish a baseline of biodiversity and habitat data in order to monitor changes prompted by the creation of the ecosanctuary and in line with the project targets.

Table 1 lists the surveys that comprise the long-term ecological monitoring programme so far. New monitoring schemes will include a habitat Phase 1 assessment, annual habitat condition assessments, and baseline monitoring for small native mammals. The methodology of each survey can be found in the Plémont Seabird Sanctuary Project 2024 Biodiversity Report.

Table 1. Research schemes and surveys implemented since 2017

Method	Target species
Desk review	Atlantic Puffin; Northern Fulmar; Storm petrel; Razorbill; Common Guillemot; Manx shearwater
Habitat condition assessment	Plant species and habitat types
Seabird census and productivity	All seabirds
Playback survey	Burrowing seabirds
Bird Ringing	Nocturnal seabirds
Acoustic recording	All birds
Breeding bird survey	Terrestrial birds
Dusk survey	Nocturnal mammals Nocturnal birds
Motion-activated cameras	All species
Mark/recapture	Non-native predators
Radio-tracking	Non-native predators (feral ferret)
Flavoured wax blocks	Non-native predators
Tracking tunnels	Non-native predators
Refugia transect	Reptiles Native mammals

## 1.7 Community programme

The community programme started in 2018 with the aims of informing, engaging and empowering the community to become the leading voice in the sanctuary project.

The Ecosanctuary aims to offer resources and activities to all sectors of the community. Details about the community programme can be found on the document '*Case for the Plémont Seabird Sanctuary*' of this application.

## 2. SCHEDULE OF WORKS YEAR 1 TO 5

	2025				2026				2027				2028				2029				
<b>INVASIVE PLANT REMOVAL</b>																					
Bracken cutting			■	■			■	■			■	■			■	■			■	■	
Grazing							■	■			■	■			■	■			■	■	
Chemical (optional)							■														
<b>HABITAT RESTORATION &amp; ENHANCEMENT</b>																					
Artificial burrows and nest boxes				■	■			■	■			■	■			■	■				
Seabird decoys	■	■	■		■	■	■		■	■	■		■	■	■		■	■	■		
Colony calls		■				■				■				■				■			
<b>FENCE</b>																					
Installation			■	■																	
Maintenance					■	■	■	■	■	■	■	■	■	■	■	■	■	■	■	■	
<b>NON-NATIVE PREDATOR ERADICATION</b>																					
Monitoring	■	■	■	■																	
Phase 1 Live traps			■	■	■	■		■	■												
Phase 2 Lethal traps (rats only)							■	■			■	■			■	■					
Phase 3 Bait (optional)											■	■			■	■					
<b>BIOSECURITY PROTOCOL</b>					■	■	■	■	■	■	■	■	■	■	■	■	■	■	■	■	
<b>ECOLOGICAL MONITORING (ongoing)</b>	■	■	■	■	■	■	■	■	■	■	■	■	■	■	■	■	■	■	■	■	
<b>COMMUNITY PROGRAMME (ongoing)</b>	■	■	■	■	■	■	■	■	■	■	■	■	■	■	■	■	■	■	■	■	



### 3. PROJECT TARGETS (KEY PERFORMANCE INDICATORS)

	5 YEARS	10 YEARS	15 YEARS	20 YEARS
INVASIVE BRACKEN REMOVAL	20% of total bracken area	40% of total bracken area	60% of total bracken area	80% of total bracken area
HABITAT RESTORATION AND ENHANCEMENT	Bracken growth curtailed	Plant diversity maintained	Plant diversity increased	Plant diversity secured
FENCE AND BIOSECURITY PROTOCOL	Predator-exclusion status	Predator-exclusion status	Predator-exclusion status	Predator-exclusion status
BIODIVERSITY GAINS	Prevent extinction	Stabilize populations	Increase populations	Increase diversity
Seabirds <ul style="list-style-type: none"> <li>- Puffin</li> <li>- Razorbill</li> <li>- Guillemot</li> <li>- Fulmar</li> <li>- Manx shearwater</li> <li>- Storm petrel</li> <li>- Lesser B.B. Gull</li> <li>- E. Shag</li> </ul>	Prevent extinction Prevent extinction Increase sightings Prevent extinction Increase sightings Increase sightings Prevent extinction Prevent extinction	Stabilize population Stabilize population Establish breeding Stabilize population Establish breeding Establish breeding Stabilize population Stabilize population	Increase population Increase population Increase population Stabilize population Increase population Increase population Increase population Increase population	Increase population Increase population Increase population Stabilize population Increase population Increase population Stabilise population Stabilise population
Land birds <ul style="list-style-type: none"> <li>- Dartford warbler</li> <li>- Stonechat</li> <li>- Linnet</li> <li>- Meadow pipit</li> <li>- Rock pipit</li> </ul>	Increase sightings Increase sightings Increase sightings Increase sightings Prevent extinction	Establish breeding Establish breeding Establish breeding Establish breeding Stabilize population	Increase population Increase population Increase population Increase population Increase population	Stabilise population Stabilise population Stabilise population Stabilise population Stabilise population

	<b>5 YEARS</b>	<b>10 YEARS</b>	<b>15 YEARS</b>	<b>20 YEARS</b>
Reptiles	Prevent extinction	Stabilize populations	Stabilize populations	Stabilize populations
Native mammals	Prevent extinction	Stabilize populations	Stabilize populations	Stabilize populations
COMMUNITY PROGRAMME	4 stakeholder meets / year 4 public events / year 4 public talks / year 4 school visited / year Student programme started 500 feedback surveys Interpretation signs 2 art exhibitions Nature And Kids launched ('NAK')	4 stakeholder meets / year 4 public events / year 4 public talks / year 4 school visited / year 4 visiting students / year 500 feedback surveys Increase in awareness Increase in participation 2 art exhibitions NAK 3 events/year	4 stakeholder meets / year 4 public events / year 4 public talks / year 4 school visited / year 4 visiting students / year 500 feedback surveys Increase in awareness Increase in participation 2 art exhibitions NAK 3 events/year	4 stakeholder meets / year 4 public events / year 4 public talks / year 4 school visited / year 4 visiting students / year 500 feedback surveys Increase in awareness Increase in participation 2 art exhibitions NAK 3 events/year

**LANDSCAPE OF HABITATS AND BIODIVERSITY TARGETS**



#### 4. REFERENCES

- Anderson, G. Q. A., & Green, R. E. (2009). The value of ringing for bird conservation. *Ringling & Migration*, 24(3), 205–212.
- Arnold, R. (2001). A history of the birds of Puffin Island, in P Hope Jones & P Whalley (2004) *Birds of Anglesey*. Menter Mon Llangejni.
- Ascension Island Government Conservation and Fisheries Directorate (2020). Ascension Island Biosecurity Strategy 2020-2025.
- Bedolla-Guzmán, Y., Masello, J. F., Aguirre-Muñoz, A., & Quillfeldt, P. (2016). A wood-concrete nest box to study burrow-nesting petrels. *Marine Ornithology* 44: 249–252.
- Butcher, B. (2018). *UK Habitat Classification – Habitat Definitions V1.0*. UK Habitat Classification Working Group.
- Buxton, R. T., & Jones, I. L. (2012). Measuring nocturnal seabird activity and status using acoustic recording devices: Applications for island restoration: Acoustic Monitoring of Nocturnal Seabirds. *Journal of Field Ornithology*, 83(1), 47–60.
- Day, J., Robertson, P., & Symes, N. (2003). *The scrub management handbook: Guidance on the management of scrub on nature conservation sites*. English Nature.
- Ecology, R., White, G., & Hirons, G. (2019). *The predator exclusion fence manual*. Guidance on the use of predator exclusion fences to reduce mammalian predation on ground-nesting birds on RSPB reserves. Royal Society for the Protection of Birds, Sandy, Bedfordshire.
- Hawes, C. (2011). *A Handbook of Field Sampling Protocols for Biodiversity Indicator Monitoring*. James Hutton Institute.
- Herrera-Giraldo, J., Figuerola-Hernández, C. E., Wolf, C. A., Colón-Merced, R., Ventosa-Febles, E., Silander, S., & Holmes, N. D. (2021). The use of social attraction techniques to restore seabird colonies on Desecheo Island, Puerto Rico. *Ecological Solutions and Evidence*, 2(2), e12058.
- Keitt, B., Griffiths, R., Boudjelas, S., Broome, K., Cranwell, S., Millett, J., Pitt, W., & Samaniego-Herrera, A. (2015). Best practice guidelines for rat eradication on tropical islands. *Biological Conservation*, 185, 17–26.
- Lea', P., & Road, P. (2013). Preparation of New Habitat Condition Monitoring Objectives. Penny Anderson Associates Ltd. Department of the Environment, Jersey.
- Lor, S. (1991). Artificial and Modified Burrows Establishment for the Atlantic Puffin (*Fratercu la arctica*) on Petit Manan Island , Maine. Biology Department, Ripon College, Ripon, WI 54971.
- Lovering, T. (2008). Core management plan including conservation objectives for Skomer and Skokholm SPA (Special Protection Area). Countryside Council for Wales.
- Newton, I. (2014). Is bird ringing still necessary? *British Birds* 107, 572-574.
- Oppel, S., Hervias, S., Oliveira, N., Pipa, T., Silva, C., Geraldés, P., Goh, M., Immler, E., & McKown, M. (2014). Estimating population size of a nocturnal burrow-nesting seabird using acoustic monitoring and habitat mapping. *Nature Conservation*, 7, 1–13.

- Pacific Invasives Initiative. Guidelines for project managers. *Resource Kit for Rodent and Cat Eradication*. <https://www.pacificinvasivesinitiative.org/rk/>
- Perkins, A. J., Bingham, C. J., & Bolton, M. (2018). Testing the use of infra-red video cameras to census a nocturnal burrow-nesting seabird, the European Storm Petrel *Hydrobates pelagicus*. *Ibis*, *160*(2), 365–378.
- Perkins, A. J., Douse, A., Morgan, G., Cooper, A., & Bolton, M. (2017). Using dual-sex calls improves the playback census method for a nocturnal burrow-nesting seabird, the Manx Shearwater *Puffinus puffinus*. *Bird Study*, *64*(2), 146–158.
- Priddel, D., & Carlile, N. (1995). An Artificial Nest Box For Burrow-Nesting Seabirds. *Emu - Austral Ornithology*, *95*(4), 290–294.
- Russell, J. C., Beaven, B. M., MacKay, J. W. B., Towns, D. R., & Clout, M. N. (2008). Testing island biosecurity systems for invasive rats. *Wildlife Research*, *35*(3), 215.
- Russell, J. C., Towns, D. R., & Clout, M. N. (2008). *Review of rat invasion biology: Implications for island biosecurity*. Science & Technical Pub., Dept. of Conservation.
- SEARS Natural Scotland (2008) Bracken Control: A Guide to Best Practice. Scottish Government.
- Shonfield, J., & Bayne, E. M. (2017). Autonomous recording units in avian ecological research: Current use and future applications. *Avian Conservation and Ecology*, *12*(1), art14.
- Turnbull, A. (2021). Zealandia Te Māra a Tāne: Predator Exclusion Fence Design and Maintenance. *New Zealand*. Karori Sanctuary Trust, Wellington, New Zealand.
- Thomas, S., Varnham, K. & Havery, S. (2017): Current Recommended Procedures for UK (bait station) rodent eradication projects. (Version 4.0) Royal Society for the Protection of Birds, Sandy, Bedfordshire.
- Voříšek, P. (2008). *A best practice guide for wild bird monitoring schemes* (1st ed). Pan-European Common Bird Monitoring Scheme (PECMBS) : European Bird census Council (EBCC) : Birdlife International ; Statistics Netherlands ; Royal Society for Protection of Birds (RSPB) ; Czech Society for ornithology (CSO).
- Wilson, U. W. (1986). *Artificial rhinoceros auklet burrows: a useful tool for management and research*. *J. Field Ornithol.*, *57*(4):295-299

## APPENDIX. STATEMENT FROM CAPE SANCTUARY, HAWKE'S BAY, NZ 15/3/24

“There's no quick recipe for restoration. When every day is spent in a place of slow-moving progress it's often difficult to see just how far we've come. The 'sea-bird site' perched above the Ocean Beach Cliffside's is a perfect example of this.

In 2005, co-founder Andy decided that a two-hectare slice of eroded farmland was destined for greatness.

By 2007 a pest-proof fence enclosed thousands of planted native trees indistinguishable from the weed riddled hillside.

By 2009 Ngaio, harakeke and akeake had emerged, teasing us with promise of a forested future.

By 2014 flowers from a loosely knitted canopy convinced the return of tuī, korimako and riroriro.

By 2016 grey-faced petrels filled sea-bird boxes kept cool by the shelter of harakeke.

By 2021, a smorgasbord of shades of green had erupted from the hillside.

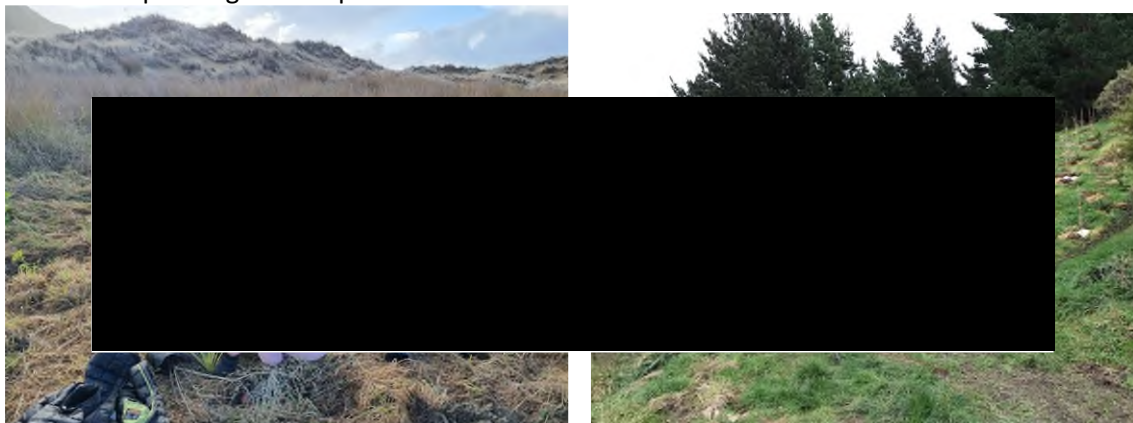
By 2022 seedlings, ferns and mosses had covered the once barren, cracked, clay-baked soil.

By 2023, our sea-bird was a self-sufficient ecosystem, where our forest supported a plethora of reptile, bird and invertebrate life. “



# CAPE SANCTUARY

Volunteers planting native species.



Volunteers installing nest boxes for burrowing seabirds.



Little blue penguin using a purpose-built nest box. Over 800 kiwis have been re-introduced at the sanctuary.



View of the ecosanctuary.



All images have been sourced from Cape Sanctuary facebook page ©.

